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#### **Critical Review**

## Plastic particle ingestion by wild freshwater fish: a critical review

France Collard, Johnny Gasperi, Geir Wing Gabrielsen, and Bruno Tassin

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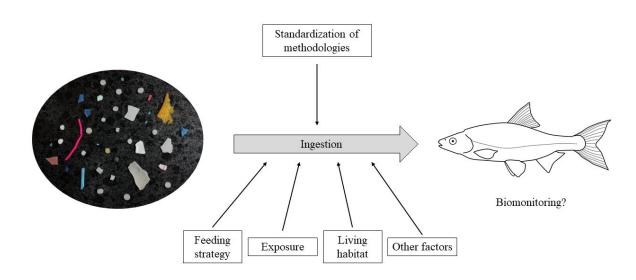
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1	Plastic particle ingestion by wild freshwater fish: a critical review
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5	France Collard <sup>*1,2</sup> , Johnny Gasperi <sup>1,3</sup> , Geir W. Gabrielsen <sup>2</sup> , Bruno Tassin <sup>1</sup>
6	
7	<sup>1</sup> Laboratoire Eau Environnement et Systèmes urbains (LEESU), Université Paris-Est Créteil,
8	61 avenue du Général de Gaulle, 94010 Cedex Créteil, France.
9	<sup>2</sup> Norwegian Polar Institute, Fram Centre, NO-9296 Tromsø, Norway.
10	<sup>3</sup> Water and Environment Laboratory (LEE), Geotechnical engineering, Environment, Natural
11	hazards and Earth Sciences Department (GERS), French Institute of Science and Technology
12	for Transport, Development and Networks (IFSTTAR), IRSTV, 44340 Bouguenais, France.
13	
14	
15	*Corresponding author:
16	francecollard16@gmail.com
17	phone: +4777750655
18	Current address:
19	Norwegian Polar Institute
20	Research Department
21	Hjalmar Johansens gate 14
22	9296 Tromsø, Norway
23	
24	

## 25 Graphical abstract



29 Abstract

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Plastic pollution, especially microplastics (MP) pollution, is a hot topic in both mainstream 31 media and scientific literature. Although rivers are potentially the major transport pathway of 32 this pollution to the sea, plastic contamination in freshwater bodies is comparatively 33 understudied. Microplastic pollution in freshwater fish is of growing interest, and while few 34 studies exist, discrepancies do occur in the sampling, extraction, and identification of MP and 35 in the expression of the results. Even though those differences hamper comparisons between 36 some studies, a comparative work has been performed to identify the factors influencing MP 37 38 ingestion by fish and consequently to target potential ecological traits that can be used to monitor species. Monitoring plastic ingested by fish will give relevant ecological information 39 on MP pollution. This review focuses on MP ingestion by wild freshwater and estuarine fish. 40 41 In addition to providing an overview of the existing data concerning contamination levels in wild freshwater fish, we aimed to (1) propose several overall recommendations on the 42 methodologies applicable to all biota, (2) compare MP contamination levels in fish and in their 43 environment, and (3) determine which parameters could help to define fish species for 44 monitoring. 45

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1. Introduction 48

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50 Due to their exponentially increasing production since the 1950s,<sup>1</sup> plastic materials are polluting 51 all types of environments: marine surface waters,<sup>2,3</sup> deep-sea sediments,<sup>4,5</sup> arctic sea ice,<sup>6,7</sup> 52 soils,<sup>8</sup> and even air.<sup>9,10</sup> As an expected and direct consequence, the number of species or taxa 53 exposed to plastic pollution is alarming.<sup>11</sup> Plastic pieces of all sizes are found in the

environment, and thus are ingested by many different organisms: marine mammals,<sup>12-14</sup> marine 54 and terrestrial birds,<sup>15-17</sup> crustaceans,<sup>18,19</sup> worms,<sup>20</sup> and fish,<sup>21-23</sup> Scientific research is currently 55 more focused on the marine environment. The reasons for this focus are probably due the high 56 economic value of marine resources and abundant funding allocated to marine research and 57 monitoring. In particular, microplastics (MP), which are plastic pieces with sizes less than 5 58 mm,<sup>24</sup> are well known to pollute seas and oceans. Numerous laboratory studies have shown that 59 this exposure leads to plastic ingestion by various organisms and is associated with some 60 negative impacts,<sup>25</sup> such as neurotoxicity,<sup>26</sup> a change in swimming behavior<sup>27</sup> and reduction of 61 predatory performances.<sup>28</sup> 62

Human populations are closely linked to water from both marine and freshwater sources. A 63 considerable fraction of the human population lives in the near-coastal zone.<sup>29</sup> Approximately 64 half of the world's population lives within 3 km of a freshwater body and only 10% of the 65 population lives farther than 10 km away.<sup>30</sup> However, given that proximity, between 1.15 and 66 2.41 million tons of plastics are released into the oceans via rivers each year.<sup>31</sup> Conventional 67 wastewater treatment plants may act as an MP source to rivers,<sup>32,33</sup> and inappropriate waste 68 management<sup>34</sup> combined with a high population density<sup>35</sup> may both be positively correlated 69 with riverine plastic loads. 70

Increasing number of studies have revealed that freshwaters are contaminated, sometimes 71 largely by plastic, regardless of its size.<sup>36-39</sup> The number of studies about MP pollution and its 72 impacts is growing (in March 2019, searching the Web of Science for "plastic marine pollution" 73 returned 3 results in 2000 and 351 results in 2018), but a small number of these studies deal 74 with a freshwater environment (0 out of the 3 results found in 2000 and 42 out of the 351 results 75 found in 2018). The same observation can be made for fish; searching the Web of Science for 76 the keywords "fish plastic pollution" gave 4 results in 2000 and 99 in 2018, in which 0 and 10 77 results concern freshwaters, respectively. Despite the large amounts of plastic debris input into 78

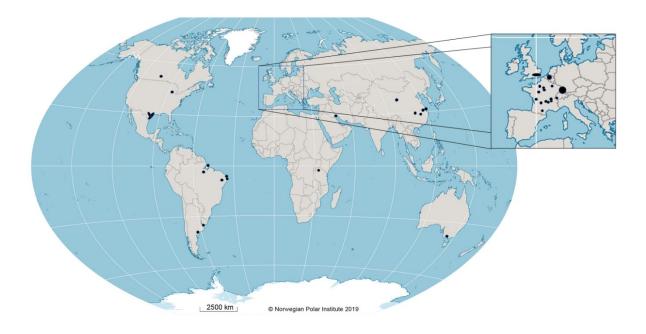
relies on visual sorting. In addition, the results are reported in one or two units while at least three are used in publications. Those discrepancies lead to difficulties in comparing data.

Fish are the main taxa studied with regard to MP ingestion in freshwater environments<sup>40</sup> and 84 constitute an economically, ecologically and highly diversified group.<sup>41,42</sup> Microplastic 85 ingestion by freshwater fish has been studied worldwide, but these studies have although 86 focused on limited regions (Fig. 1) in comparison with the studies of plastic ingested by wild 87 marine fish.<sup>43</sup> In total, the studies focusing on freshwater fish, while few, have described MP 88 ingestion by more than 200 species. This large number indicates that some studies have 89 collected few samples. Few published papers have studied the interactions between MP and 90 91 freshwater biota compared to the number of marine studies, although scientific interest in MP ingestion by freshwater fish is rising. We thus believe that a summary of recommendations for 92 the methods and expression of results and factors leading to MP ingestion by freshwater fish 93 would be useful for the scientific community, especially regarding monitoring. Three main 94 factors have been investigated: the MP levels in abiotic compartments, the living habitat and 95 96 the feeding strategy. Other factors could have been reviewed, but there are too few studies to allow for a reliable discussion. Exposure through abiotic compartments could be directly 97 correlated with the ingestion of MP. As further discussed in this review, MP ingestion does not 98 99 always depend on abiotic compartment exposure. Additionally, there is not a clear relationship between MP ingestion and the living habitat and feeding strategy. Some studies have found that 100 benthic or demersal fish ingested more MP,<sup>44,45</sup> while others reported the contrary.<sup>22</sup> 101 Monitoring and assessment are essential steps towards addressing specific questions about 102

103 marine litter, including MP. Monitoring and assessment are needed to assess the state or level

104 of pollution and provide objective information to design mitigation measures as well as assess 105 their effectiveness and promote adaptive management.<sup>46</sup> Recommendations for defining 106 indicator fish species are very few but crucial to improve the comparability between studies.

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109 Figure 1. Global overview of the studies performed on MP ingestion by wild freshwater fish.

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In this critical review, we aim to (1) propose several recommendations for the harmonization of methodologies, which is applicable not only to MP contamination in freshwater fish studies but also to all biota, (2) compare MP contamination levels in fish and their environment, and subsequently (3) determine which parameters could help to define fish species for monitoring. Our work is based on the occurrence of MP in the guts of fish. The translocation of plastic particles to other organs is beyond the scope of this review but should be considered in further studies.

119 2. Methodologies for the study of MP contamination in fish guts

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Globally, studies on MP ingestion by fish, both marine and freshwater, follow a general pattern 121 regarding their methodology, which include sampling, gut content (GC) extraction, MP 122 extraction (visual sorting or chemical treatment, which is not always performed), and 123 identification (including spectroscopic analyses which are not always performed). No specific 124 125 method has been defined as the method to be used in fish GC analyses. In addition, this general pattern is not specific to studies about fish, as it is also used for other organisms or even for 126 water or sediment matrices. In the two latter cases, density separation is often chosen to carry 127 128 out the MP extraction. Whatever the matrix, discrepancies in methodologies occur, including in the study of MP in the GCs of wild freshwater fish. There are variations in all the steps used 129 to extract and identify MP. In the next sections, we highlight the main issues of the key protocol 130 steps used to extract MP from GCs (Fig. 2) and propose some recommendations to increase the 131 comparability of studies for future research. We will discuss the following main steps: the 132 sampling, extraction, and identification of MP and the expression of the results. Within those 133 steps, we highlight some parameters that we think are of great importance including the 134 sampling size, method for digesting the GCs, definition of the size threshold for the extracted 135 136 particles, spectroscopic analyses, and the way results are provided. In addition, some recommendations for preventing and evaluating the contamination will be proposed. We 137 selected the most important criteria, as other reviews have already focused on methodologies 138 for the sampling, extraction and analyses of MP in biota or abiotic compartments.<sup>40,47,48</sup> 139

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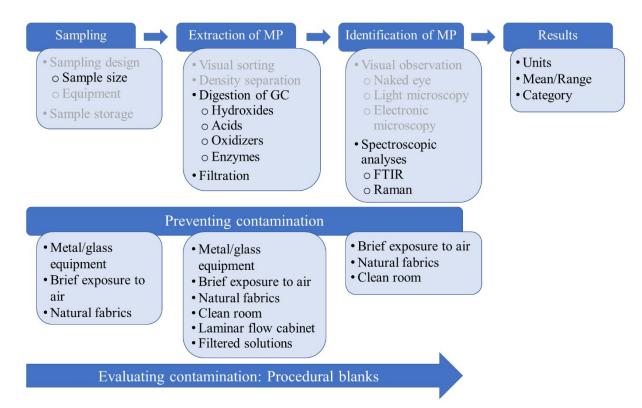




Figure 2. Flowchart of the main steps used to analyze microplastics in fish gut contents. The
steps shown in black are discussed in this review. GCs: gut contents, FTIR: Fourier transform
infrared.

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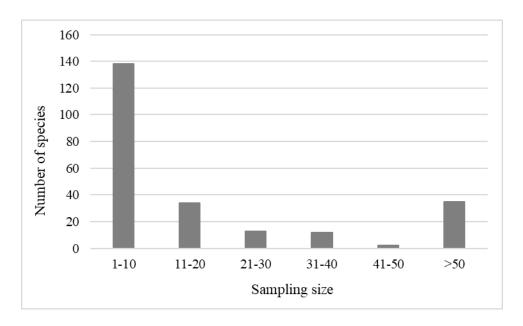
146 2.1. Sampling: Sample size

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Among all the species studied, more than twenty have been analyzed at least twice, and most 148 of these species have been studied thanks to studies conducted in South America.<sup>49-52</sup> In 149 addition, when considering a sampling of at least 10 individuals per species, only seven out of 150 twenty species can be spatially compared. A review focusing on wild marine fish found that the 151 detection of plastic ingestion was positively correlated with an increased sample size (up to 152 n=10).43 In many studies reviewed here, the number of individuals representing each species 153 was low (n < 10), which is obviously too small to define solid trends in MP contamination. In 154 several publications, data were given for one or two individuals, limiting the representativeness. 155

Given that a single dataset is available and the number of fish sampled is usually low for most 156 157 studied species, very local one-time events may influence the results and impair the global understanding of MP ingestion by fish. A large sample size is highly important since it will 158 provide enough data to perform reliable statistical analyses. Previously, a threshold of 50 159 individuals has been defined as sufficiently reliable regarding to achieve statistical power, 48,53,54 160 while 40 specimens were considered to be an adequate sample size for the monitoring of plastic 161 ingestion by northern fulmars.<sup>55</sup> Fig. 3 shows that most studies did not sample more than 50 162 specimens per species. This chart highlights the need for a better sampling design to statistically 163 support the results. 164

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167 Figure 3. Number of species for each category of sampling size. This chart is based on studies168 shown in Table 1.

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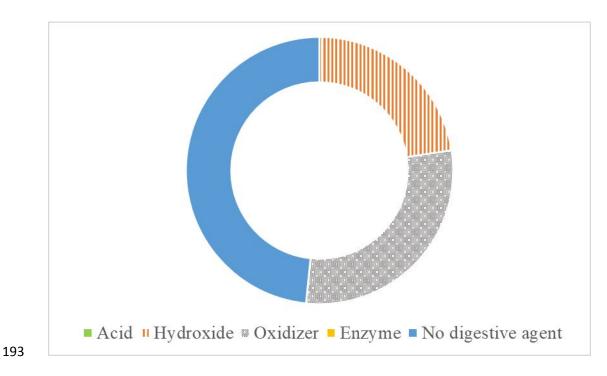
#### 170 2.2. Extraction of MP: Digestion of fish gut contents

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172 The digestion of organic matter is not mandatory but recommended,<sup>46</sup> especially when aiming

to isolate MP from biota. Several methods exist regarding GCs, which are indiscriminate for

marine and freshwater fish species. Fig. 4 shows that when present, the most common agents 174 are oxidizers (e.g., H<sub>2</sub>O<sub>2</sub> and NaClO)<sup>44,56,57</sup> and hydroxides (e.g., KOH).<sup>58,59</sup> No studies reported 175 acids as the main digesting agent, which is most likely because acids alter some polymers, such 176 as polyethylene terephthalate (PET), high density polyethylene (HDPE) and polyamide (PA). 177 Acids can also lead to suboptimal digestion.<sup>58,60-62</sup> Remarkably, to our knowledge, no studies 178 have used enzymes as digesting agents yet. While the efficiency of such agents has been 179 assessed,<sup>62-64</sup> the cost and duration of this treatment might discourage researchers. Recently, 180 von Friesen et al.<sup>65</sup> proposed a protocol more efficient than that using KOH with the convenient 181 advantages of being less time-consuming and commercially available at a low price. This 182 183 protocol has been thoroughly tested using bivalves but suits a broader range of organisms according to the authors. No studies on fish GCs have used Fenton's agent either, which is 184 probably due to the efficiency of the previously cited methods. This agent does not affect plastic 185 186 polymers and reduces the sample preparation time.<sup>66</sup> Both enzymes and Fenton's agent could be a solution to achieving the efficient, rapid and inexpensive digestion of the organic matter in 187 the GCs of freshwater fish but could probably also be used in other biological matrices such as 188 the liver. Some GCs can be very fatty and challenging to digest. No study has identified a means 189 of treating those samples properly. Finding an easy way to process fatty GCs would be useful 190 191 for further research on fish gut contamination.



194 Figure 4. Proportion of studies according to the main digestive agent used.

195

Only a couple of studies have reported data on all types of anthropogenic particles, *i.e.*, small 196 pieces created or handled by humans, such as plastics, dyed particles or textile fibers.<sup>57,67</sup> 197 198 Indeed, studies have only focused on plastic materials and have developed protocols for that purpose. To our knowledge, few studies have tested the protocols used to extract MP from biota 199 on natural or modified natural materials such as cellulose or cellulose acetate.<sup>58,68,69</sup> Dehaut et 200 al.58 tested three extraction protocols on cellulose acetate materials, in which two contained 201 hydroxides (NaOH & KOH) and one contained a combination of a hydroxide and an oxidizer 202 (NaOH & K<sub>2</sub>S<sub>2</sub>O<sub>8</sub>). All the protocols led to a mass reduction and a size and shape modification 203 of the cellulose acetate while two of them allowed for identification by Raman spectroscopy. 204 On the other hand, cellulose fibers, along with plastic polymers, were tested with an oxidizing 205 agent and a combination of an oxidizer and an acid (NaClO & HNO<sub>3</sub>).<sup>68</sup> Cellulose fibers were 206 left unaffected by both treatments and Raman spectroscopy could successfully be performed. 207 A potassium hydroxide-based protocol has also been tested on several natural materials.<sup>69</sup> 208 Cellulose acetate has been found to lose a significant part of its mass after treatment, which is 209

similar to the study by Dehaut et al.<sup>58</sup> Some natural materials, such as sheep wool, were dissolved, while others showed a resistance to KOH. The extraction method should then be carefully chosen when textile fibers are targeted.

Data on anthropogenic particles made from natural materials such as wool and cotton should be included in further studies. Laboratory studies have mainly focused on plastic materials. Textile fibers, independent of their composition, are found in high quantities in the environment<sup>3,70</sup> but their impacts are unknown. These particles contain additives and dyes<sup>16,70,71</sup> and may constitute a threat to the environment<sup>72,73</sup>, aquatic organisms and human health.<sup>71,74</sup> All anthropogenic particles, regardless of their size, should therefore be studied in both in situ and in vitro studies.

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#### 221 2.3. Extraction of MP: Filtration and targeted MP sizes

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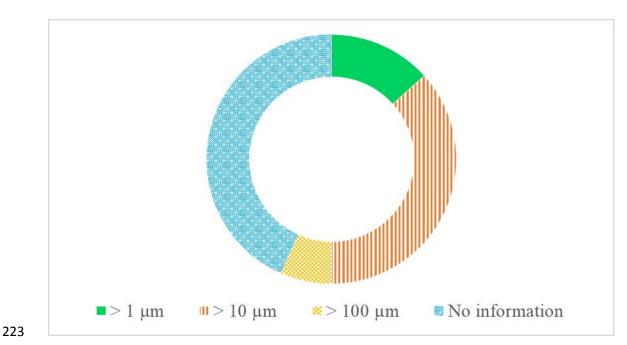


Figure 5. Proportion of studies according to their lowest limit of MP extracted. Based on Table

225 1.

The minimum size of the targeted particles is not always mentioned (Fig. 5). Generally, the 227 228 particle isolation process includes a filtration/sieving step, in which a precise mesh size practically defines the threshold for particles to be further analyzed. In several studies, that 229 mesh size is close to the micrometer scale (Fig. 5). Although this mesh size perfectly fits the 230 microplastic size definition, from 5 mm down to a few microns,<sup>75</sup> the limiting factor is then the 231 spectroscopic analyses. Depending on the technique and equipment used, the size threshold 232 allowing for the analysis of a particle varies. For instance, when dealing with particles between 233 1 and 20 µm in length, Raman spectroscopy is recommended.<sup>76</sup> However, if the aim of the study 234 is to investigate direct ingestion by fish, it might not be worth studying such small MP. Adult 235 236 fish are unlikely to ingest micron-sized particles because these particles will probably be ejected into the surrounding water from the branchial system, which is not adapted to retain such small 237 particles.<sup>77</sup> The filtration step could then be made easier by using larger mesh sizes depending 238 239 on the studied species, mainly if no digestion step is performed. However, if the transfer of MP from prey is considered in the study, looking at smaller particles might be of interest. Regarding 240 filtering fish, their filtration system defines the size limit of particles that are retained and further 241 swallowed. For those species, correlating the filtration-limited size to the size of MP ingested 242 might be a way to discriminate MP coming from prey (indirect ingestion) and MP ingested 243 244 through feeding (direct ingestion). This assumption has to be assessed by future studies. A size of 5 µm could be the lowest limit of investigated MP sizes, as it almost perfectly covers the MP 245 size definition, which comprises direct and indirect MP ingestion and allows for reliable 246 spectroscopic analyses. We encourage authors to provide at least the size classes of the ingested 247 particles, as well as the size distribution. This information will allow for the comparison of 248 those size classes according to the minimal size targeted and will provide baseline data for 249 studies further dealing with the uptake of ingested MP by other organs and tissues. 250

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252 2.4. Ic

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2.4. Identification of MP: Spectroscopic analyses

Half of the studies on plastics in freshwater fish used spectroscopic analyses to confirm the 254 plastic composition of the particles found, and most of these studies are the oldest (Table 1). 255 When considering other matrices such as sediment or water, such techniques are often set 256 aside.<sup>76</sup> Thus, these techniques only rely on visual observation, which may bias the results.<sup>78-80</sup> 257 258 Visual observations with a stereomicroscope tend to underestimate the amount of microplastic fragments and, conversely, tend to overestimate the amount of MP fibers compared to the 259 results obtained by Fourier transform infrared (FTIR) analyses.<sup>79</sup> Markic et al.<sup>43</sup> found that 260 261 plastic was detected more often when a chemical digestion-coupled spectroscopic analyses was performed compared to visual methods alone. The smaller the particles are, the more difficult 262 it is to distinguish them from biological tissues and to assess their synthetic composition with 263 264 the naked eye. The 'hot point-test' is also sometimes used to determine the plastic composition of particles.<sup>18,81</sup> This test consists of touching a particle with a hot needle and observing whether 265 the needle leaves a mark on the particle.<sup>48,82</sup> Although convenient and cheap, this technique 266 should not be solely used for identification but used as a complement to visual sorting prior to 267 spectroscopic analyses.<sup>48</sup> In a report, the GESAMP group of experts<sup>40</sup> stated that spectroscopy 268 269 is recommended for particles smaller than 1 mm. Consequently, using either Raman or FTIR spectroscopy when working with microplastic pollution is highly recommended.<sup>46,80</sup> Both 270 techniques usually allow for precisely identify the particle composition to be precisely 271 identified and have the strong advantage of being nondestructive in contrast with other 272 techniques such as pyrolysis-gas chromatography coupled with mass spectrometry (py-GC/MS) 273 or thermogravimetric analysis coupled with solid-phase extraction (TGA-SPE).<sup>83-85</sup> These 274 techniques are complementary, as they give different information but using two or three of these 275 techniques is obviously expensive and time-consuming. Therefore, depending on the 276

information needed or the type of particle, researchers have to choose the best technique. Raman 277 278 spectroscopy constitutes a better choice if results about additives are needed. Indeed Raman spectroscopy can provide additional information on dyes, for example, and will identify the 279 composition of small particles (< 20  $\mu$ m) that FTIR will miss.<sup>79,84,85</sup> On the other hand, FTIR 280 spectroscopy is less time-consuming and may allow for the identification of black particles that 281 lead to fluorescence issues in Raman spectroscopy.<sup>84</sup> All these techniques are expensive and 282 time consuming; thus, a random subsample is often collected prior to the analyses, and the 283 results are extrapolated to the total number particles found.<sup>10,44,86-88</sup> This random subsample is 284 usually poorly described and might result from subjective choices, especially regarding the 285 286 sizes and colors. As selecting the particles may lead to bias, we instead recommend selecting a zone covering both the center and the edge of the filter (a half or a quarter) used during MP 287 extraction. 288

Spectroscopic analyses provide useful and reliable data on particles. However, part of this data is often set aside because studies focus only on plastic polymers (Table 1). As explained in a previous section, we encourage researchers to provide all information, even on natural and semisynthetic materials.

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294	2.5.	Expressi	on of the	results
			011 01 0110	1.0000000

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We recorded three different units of MP contamination in GCs, *i.e.*, the percentage of contaminated individuals, the number of MP per individual and the number of MP in one gram of the GCs. Additionally, units of contamination differ between studies making comparisons impossible. All three units are easy to calculate and, although the most common (Table 2), the percentage of occurrence is not the most representative unit. This unit gives an overview of the number of fish having ingested at least one piece of plastic in a given time but does not give any valuable quantitative information. The number of MP per gram of GCs could be considered to be the most accurate unit but greatly depends on the fullness of the stomach or gut. As it is not clear whether MP have a greater retention time than food in the digestive tract,<sup>89,90</sup> expressing the quantity of MP per mass of gut contents is not the most consistent option. It is instead preferable to express the results as the number of MP per individual and for comparison purposes, we recommend authors and researchers to report the results with the three units described in this review.

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310 2.6. Procedural blanks

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Procedural blanks (PBs), also termed controls or negative controls,<sup>47,91</sup> are used in all procedure 312 steps with the reagents only, and the sample in itself is excluded. Few studies that we have 313 314 reviewed made PBs although most of the recent studies did. Even if contamination is prevented as best as possible, including PBs is recommended to evaluate possible contamination of the 315 whole treatment protocol<sup>46,48</sup> and, if needed, to adapt the results according to those blank 316 data.<sup>18,82,91</sup> Plastic contamination levels depend on many parameters, including the type of 317 equipment used for sampling, the extraction and the analyses.<sup>92</sup> Given that microplastic fibers 318 are also found in the air.<sup>9,10</sup> tap water<sup>93</sup> and table salt.<sup>94,95</sup> in which the latter is used in 319 laboratories to perform density separation, PBs are necessary to evaluate the contamination 320 levels in environmental samples. Moreover, we recommend including some PBs before starting 321 the sample treatment. This will give an overview about the contamination level in the working 322 environment, which can be then compared to the expected levels in the samples. This step could 323 be critical if minimally contaminated samples are to be processed. In this case, PB levels would 324 easily exceed what may be found in the samples, leading to an impossible interpretation of the 325 results. Depending on the studies and equipment used, PB contamination can be important: 50 326

- 327 times higher than plastics within the samples,<sup>96</sup> or low: less than 5% of all plastics found in the
- 328 samples.<sup>44</sup> However, if PB levels are assessed first, precautions and decisions can be made to
- 329 prevent the samples from being highly contaminated during the treatment.

No. sp.	n/sp.	Country	Water body	part		Extracted particle size	Visual sorting		Digestii			c ana	roscopi alyses	Procedural blanks	Reference
	60, 60,			AP	PL	(µm)		Hyd.	Acid	Enz.	Ox.	Rn	FTIR		
3	62	Brazil	Е	-	Х	?	X	-	-	-	-	-	-	-	Possatto et al. <sup>97</sup>
2	239 & 330	Brazil	Е	-	Х	?	X	-	-	-	-	-	-	-	Dantas et al.49
3	240, 141, 44	Brazil	Е	-	X	?	X	-	-	-	-	-	-	-	Ramos et al. <sup>50</sup>
1	186	France	R	-	X	> 1.2	Х	-	-	-	-	-	-	-	Sanchez et al.98
2	20	Tanzania	L	-	X	> 500	X	X	-	-	-	-	X	-	Biginagwa et al.99
4	10	Switzerland	L	-	X	?	X	-	-	-	-	-	-	-	Faure et al. <sup>100</sup>
44	1-67	U.S.A.	R & E	-	Х	?	Х	-	-	-	-	-	X	-	Phillips & Bonner <sup>101</sup>
1	530	Brazil	Е	-	X	?	Х	-	-	-	-	-	-	-	Ferreira et al. <sup>102</sup>
2	318 & 118	U.S.A.	R	-	X	> 53	X	-	-	-	-	-	-	-	Peters & Bratton <sup>103</sup>
5	10-75	Canada	R	-	X	> 5	X*	-	-	-	X	-	-	Х	Campbell et al. <sup>96</sup>
6	20-40	China	L	-	Х	> 5	Х	-	-	-	X	-	X	Х	Jabeen et al. <sup>44</sup>
2	10 & 66	England	R & E	-	X	?	Х	-	-	-	-	-	X	-	McGoran et al. <sup>104</sup>
11	1-21	Argentina	Е	-	X	?	X*	-	-	-	X	-	-	-	Pazos et al. <sup>56</sup>
1	48	Brazil	R	-	Х	> 63	Х	-	-	-	-	-	-	-	Silva-Cavalcanti et al. <sup>105</sup>
27	2-215	Brazil	Е	-	Х	?	Х	-	-	-	-	-	-	-	Vendel et al. <sup>51</sup>
13	1-6	China	R	-	X	> 1.2	X*	Х	-	-	-	X	-	-	Zhang et al. <sup>59</sup>
1	60	France	R	Х	Х	> 5	-	-	-	-	X	X	-	Х	Collard et al. <sup>57</sup>
1	64	England	R	Х	X	> 1.2	Х	-	-	-	-	X	-	-	Horton et al. <sup>67</sup>
46	1-16	Brazil	Е	-	X	?	X	-	-	-	-	-	X	-	Pegado et al. <sup>52</sup>
4	4-17	Iran	Е	-	X	?	X	X	X	-	X	-	-	X+	Abbasi et al. <sup>106</sup>
11	1-17	U.S.A.	R	-	X	> 0.45	X*	-	-	-	X	-	X	Х	McNeish et al. <sup>91</sup>
1	10	China	L	-	Х	> 200	Х	X	-	-	-	X	-	Х	Xiong et al. <sup>107</sup>
16	1-63	Brazil	R	-	Х	?	Х	-	-	-	-	-	X	-	Andrade et al. <sup>108</sup>
1	78	Belgium	R	-	Х	> 8	Х	-	-	-	X	X	X	Х	Slootmaekers et al. <sup>109</sup>
1	11	China	L	-	X	> 8	Х	X	-	-	X	X	-	Х	Yuan et al. <sup>110</sup>
14	8-9	China	Е	-	X	> 20	Х	-	-	-	X	-	X	Х	Su et al. <sup>111</sup>

2	265 & 184	Brazil	Е	-	X	?	X	-	-	-	-	-	-	-	Ferreira et al. <sup>112</sup>
1	20	Argentina	Е	-	X	> 8	Х	-	-	-	Х	-	-	-	Arias et al. <sup>113</sup>
22	-	Germany	R & L	-	X	> 20	Х	X	Х	-	-	-	-	X	Roch et al. <sup>81</sup>
1	180	Australia	W	-	X	> 20	Х	X	-	-	-	-	X	X	Su et al. <sup>114</sup>

No.=number, sp.=species, R=river, L=lake, E=estuary, W=wetland, FR=fragments, FI=fibers, AP=anthropogenic particles excluding plastic particles,

333 PL=plastics, Hyd.=hydroxides, Enz.=enzymes, Ox.=oxidizers, Rn=Raman spectroscopy, FTIR=Fourier transform infrared spectroscopy.

- \*Visual examination after the digestion process. In this table, when the minimum size of the MP was not mentioned in the study, the porosity of the
- filters was considered. +No results for those blanks were provided.

Table 2. Summary of plastic characteristics from gut contents of freshwater fish.

	Location	Plastic co	ncentration or o	occurrence	Mean plastic	S	
Species (n)		Item/individual	Item/g GC	Contaminated individuals (%)	size or range (mm)	Spectroscopic identification?	Reference
Gobio gobio (78)	Several rivers, Belgium	-	-	9	0.67	Yes	109
Gobio gobio (186)	Several rivers, France	-	-	12	-	No	98
Squalius cephalus	Seine River, France	0.16*	0.16*	15	2.67*	Yes	57
22 species	Lakes & rivers, Germany	0.2	-	18.8	0.889	No	81
Alburnus alburnus Perca fluviatilis <b>Rutilus rutilus (10)</b> Leuciscus leuciscus	Lake Geneva, Switzerland	3.1 0 0 0.3	-	7.5	-	No	100
Platichthys flesus Osmerus eperlanus	Thames Estuary, England	-	-	85 20	-	Yes	104
Rutilus rutilus (64)	River Thames, England	0.69	-	32.8	-	Yes	67
Platycephalus indicus Saurida tumbil Sillago sihama Cynoglossus abbreviatus	Musa Estuary, Iran	2.3 2.8 1.5 2.9	-	-	-	No, SEM/EDS	106
Lates niloticus Oreochromis niloticus	Lake Victoria, Tanzania	-	-	20 20	-	Yes	99
Lepomis macrochirus (318) Lepomis megalotis (118)	Brazos River, Texas, U.S.A.	-	-	45	-	No	103
44 species	Several rivers, Texas, U.S.A.	-	-	8.2	-	Yes	101
17 species	Several rivers, U.S.A.	~13	-	85	<1.5	Yes	91
Esox lucius Catostomus commersoni Notropis atheirnoides <b>Pimephales promelas (34)</b>	Wascana Creek, Canada	-	-	73.5	-	No	96

Eucalia inconstans							
11 species	Rio de La Plata, Argentina	19.2	-	100	0.06-4.7	No	56
Micropogonias furnieri	Bahía Blanca Estuary, Argentina	12.1	-	-	2.4 (median)	No	113
16 species	Xingu River, Brazil	-	-	26.7	1-15	Yes	108
46 species	Amazon Estuary, Brazil	1.2	-	13.7	1.82	Yes	52
Hoplosternum littorale	Pajeú River, Brazil	3.6	-	83	<1-12	No	105
69 species	Paraiba Estuary Mamanguape Estuary, Brazil	0.07 0.12	-	9	-	No	52
Cathorops spixii Cathorops agassizii Sciades herzbergii	Goiana Estuary, Brazil	-	-	18 33 17	-	No	97
<b>Stellifer brasiliensis (330)</b> Stellifer stellifer	Goiana Estuary, Brazil	-	-	6.9 9.2	-	No	49
Eugerres brasilianus (240) Eucinostomus melanopterus Diapterus rhombeus (44)	Goiana Estuary, Brazil	-	-	16.2 9.2 11.4	1-5	No	50
Cynoscion acoupa	Goiana Estuary, Brazil	-	-	64.2	<5	No	102
Centropomus undecimalis Centropomus mexicanus	Goiana Estuary, Brazil	1.5 1.4	-	-	~1.25	No	112
6 species	Taihu Lake, China	2.4	3.4	95.7	0.4-24.8	Yes	44
14 species	Yangtze Estuary Hangzhou Bay, China	0.3-4.5 0.5-5.3	0.3-6.2 0.1-8.8	-	22-100 25-100	Yes	111
13 species	Xiangxi River, China	-	-	25.7	0.3-1.8	Yes	59
Gymnocypris przewalskii	Qinghai Lake, China	5.4	-	-	-	Yes	107

Carassius auratus (11)	Poyang Lake, China	-	-	91	0.1-1	Yes	110
Gambusia holbrooki	Melbourne Area, Australia	0.60	-	19.4	0.09-4.86	Yes	114

338 GC=gut contents. The bolded items are the species studied in at least two publications with the number of individuals sampled. \*Personal unpublished

data and the data published Collard et al.<sup>57</sup>, which concerned all anthropogenic particles. When the global mean is not provided, we gave the total range.

- 340 3. Factors influencing MP ingestion by freshwater fish
- 341

Here, we aimed to discuss factors that are likely to influence MP ingestion by freshwater fish foraging for prey. Those factors include specific traits, such as the feeding strategy, and abiotic parameters, such as the environmental MP levels. However, the results in the discussed in situ studies are also influenced by the retention time of MP inside the gut, which is not well understood. The results might be different depending on the size of plastic pieces or the species. Also, the gut content reflects only the instantaneous diet, that is why repeated studies should be performed.

- 348
- 349 3.1. Levels in abiotic compartments
- 350

According to studies, several factors are at play when plastic is ingested by fish. The first factor to consider is probably the exposure. Very few studies have investigated both the abiotic contamination and MP ingestion by freshwater fish. Therefore, this section will give an overview of the contamination levels in those matrices as a first step in determining what induces or prevents MP ingestion.

Many different units are used for both water and sediment concentrations (Table S1), which is also 356 the case for marine environments.<sup>115,116</sup> Regarding water contamination, the results are often given in 357 items per cubic meter of water, and the format "items/volume" is the most commonly used. Regarding 358 the sediment, the results are usually expressed in items per kilogram of sediment (dry weight). 359 Remarkably, all except one study chose to express their results per dry weight. We thus encourage 360 researchers to further express sediment concentration per dry weight too. In addition, some authors 361 362 calculated the mean concentration while others chose to only show a range. Some authors also excluded values equal to zero from the mean calculation, and others did not give any concentration 363 mean or range. While a global mean value may be very restrictive and not representative of all the 364 parameters such as the different rivers sampled or difference in the sampling locations (close or far 365

366 from an outlet, close or far from urban activities, etc.), we recommend showing as many values as 367 possible. This will allow further studies to choose the most convenient and adapted value to be 368 compared to.

369 When considering studies that performed analyses in both water and sediment samples obtained from the same place, several trends and questions arise. Regarding the abundances of MP, some studies 370 found different patterns for the two compartments. 59,110,117,118 The authors of these studies discussed 371 those results by mentioning potential factors that could influence MP distribution. MP levels in lake 372 sediments are related to their distance from MP sources.<sup>119,120</sup> Yuan et al.<sup>110</sup> confirmed this fact. The 373 northern and middle regions of Poyang Lake are subject to anthropogenic activities and are the most 374 contaminated parts with regards to the sediment and surface waters, respectively. Sediment and water 375 obtained from the south of Poyang Lake, where less human activities occur, are less contaminated 376 than those of the northern and middle regions. Furthermore, the presence of a wastewater treatment 377 plant in the vicinity of a sampling station might induce higher concentrations of polyethylene 378 terephthalate (PET) fibers.<sup>117,121</sup> 379

The topography of the environment may also be a factor of the concentration or dilution of MP in sediment.<sup>122</sup> In addition, the hydrodynamics, which are influenced by the shoreline morphology, may favor or prevent the deposition of plastics in the sediment; a low velocity environment, such as a harbor will enable plastics to settle.<sup>59,97,120,123</sup>

Surprisingly, MP contamination in fish shows different patterns than MP contamination in water or 384 sediment for at least one characteristic (Table S2). Sometimes, a correlation between the water or 385 sediment contamination and fish contamination exists for only one or two parameter(s) (e.g., the 386 shape or the size range) or at one sampling location.<sup>91</sup> Therefore, this shows evidences that exposure 387 is not the only factor involved,<sup>67</sup> and perhaps, that exposure is not a factor at all.<sup>124</sup> Of course, MP 388 ingestion cannot occur if there is not any plastic in the environment, but Kim et al.<sup>124</sup> found that the 389 food concentration is surprisingly more important than the MP concentration in influencing MP 390 ingestion. Kim et al.<sup>124</sup> exposed zebrafish to various mixtures of MP and food and concluded that MP 391

ingestion increased with increasing concentrations of food. However, in that study, the authors 392 exposed fish to virgin plastic. It is now known that animals react differently to virgin or biofouled 393 plastic. The copepods Calanus finmarchicus and female copepods Acartia longiremis ingested 394 395 significantly more biofouled PS beads than virgin PS beads.<sup>125</sup> Similarly, procellariiform seabirds can be fooled by the dimethyl sulfide signature present on marine plastic particles.<sup>126</sup> In 2017, it has also 396 been shown that Engraulis mordax anchovies responded to medium and high concentrations of 397 biofouled plastic odors with foraging behaviors.<sup>127</sup> Those behaviors did not appear when these 398 anchovies were exposed to clean plastic particles. 399

Indirectly, surface runoff and seasons also indirectly act on the exposure level.<sup>112,128</sup> Surface runoff
increases the exposure to plastic debris.<sup>112</sup> Consequently, the exposure plays a role, but the ingestion
also depends on other parameters, such as the age of the plastic particle or the time of the year.

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#### 404 3.2. Living habitat

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Sediment is a sink for microplastics,<sup>4,129</sup> and therefore, demersal and benthic fish could be exposed 406 to more MP than pelagic species. Some studies have focused on both pelagic and benthic or demersal 407 fish with varying results. In China, demersal species ingested more plastics than pelagic species.<sup>44</sup> 408 The same assumption was made by McGoran et al.<sup>104</sup>; a demersal species, *Platichthys flesus* flounder 409 had ingested far more MP than a pelagic species, Osmerus eperlanus smelt. McGoran et al.<sup>104</sup> 410 assumed that P. flesus had ingested more MP, because it ingested sediment, likely leading to the 411 unintended ingestion of MP. In the marine Scottish waters, demersal fish species have ingested more 412 plastics than pelagic species.<sup>45</sup> They also found a significant difference between coastal and offshore 413 414 species, in which the former are more contaminated than the latter. Nevertheless, this latter category was only represented by demersal species, which may lead to the wrong conclusions about the 415 involvement of the fish species coastal habitat on MP ingestion. Not all studies focusing on MP 416 ingestion by pelagic and demersal fish found a difference between these two categories.<sup>130</sup> Moreover, 417

Rummel et al.<sup>22</sup> showed that pelagic feeders had more MP in their stomach than demersal fish. Again,
as methods vary from one study to another and as MP ingestion depends on several factors, the results
may vary when looking at only one factor.

421

422 3.3. Feeding strategy

423

A more commonly discussed parameter is the feeding strategy of fish and their trophic position. It is 424 thought that (top-)predator fish from both fresh and marine environments might be more at risk than 425 organisms at lower trophic levels because of high energy requirements and trophic 426 transfer.<sup>91,96,102,112,131</sup> Other predators such as marine mammals have ingested MP that are too small 427 to be preyed upon by such larger animals.<sup>13,132,133</sup> This suggests that secondary ingestion has occurred 428 and thus, MP have accumulated through the food chain. The developmental stage of fish might also 429 influence MP ingestion.<sup>67</sup> Indeed, many fish change their diet throughout their development.<sup>134-137</sup> 430 Ferreira et al.<sup>102</sup> studied MP contamination in several ontogenetic phases of *Cynoscion acoupa*. The 431 adults, which are mainly piscivorous, were more contaminated than juveniles and subadults, which 432 exhibit an omnivorous diet. This result leads to the same conclusion previously reported: predator or 433 piscivorous fish are more at risk than fish displaying another type of feeding strategy. Later, Ferreira 434 et al.<sup>112</sup> made the same conclusion for two other species (the snook Centropomus undecimalis and 435 Centropomus mexicanus) obtained from the same location, the Goiana Estuary. Three ontogenetic 436 phases were sampled and compared. Juvenile snook registered the lowest MP ingestion level while 437 the piscivorous adults, registered the highest. Moreover, peaks of MP ingestion by adults coincide 438 with peak of fish ingestion. Other studies did not find such differences,<sup>108</sup> indicating the possibility 439 440 for other parameters to be involved. For example, in the marine environment, debris categories were related to the feeding behavior of the sampled fish.<sup>138</sup> Opportunistic feeders ingested all debris 441 categories (i.e., metals, wood, and plastics), while pelagic feeders had only swallowed plastic bags, 442 443 benthic fish, and hard plastic pieces.

Based on the fish feeding habits, the visual characteristics of plastic pieces may be involved in MP 444 ingestion by particulate-feeders. In contrast to ram-feeding, which consists of passively filtering water 445 and is nonselective, particulate-feeding is a selective mode, where prey are visually detected before 446 their capture.<sup>139</sup> Then, ram-feeders are less likely to be influenced by the visual characteristics of 447 plastics.<sup>77</sup> Yuan et al.<sup>110</sup> noticed that white MP were more frequently found in fish stomachs than in 448 both water and sediment. Yuan et al.<sup>110</sup> supposed that Carassius auratus could mistake MP for 449 planktonic food. Ory et al.<sup>140</sup> studied Seriolella violacea juveniles in controlled conditions and 450 exposed them to food-shaped and food-sized polyamide particles with different colors. As expected, 451 the juveniles preferentially ingested black particles, which is the same color as food pellets. The other 452 colored particles were supposed to be cocaptured with the food, as they were ingested together with 453 the food pellets only. Nevertheless, attention should be paid to the digesting agent used in the 454 extraction process as some agents, such as NaClO, change or discolor the particles.<sup>68</sup> In a more global 455 framework, the likelihood that beached macroplastics were previously bitten by fish and other 456 organisms depends on the plastic colors.<sup>141</sup> Blue and yellow macroplastics were preferentially more 457 than were other colors. Similarly, plastic threads were the main plastic type ingested by Cynoscion 458 acoupa in the Goiana Estuary, while plastic threads represented only 1.4% of all the plastic debris in 459 the water column.<sup>102,142</sup> Several colors of threads were ingested, suggesting that the plastic shape was 460 more important than the color for this species. The same observation was made by Carson,<sup>141</sup> who 461 found that bottle shaped macroplastics displayed more bite marks than other shapes. Visual and 462 chemical cues thus interfere together with the feeding behavior of fish, making the understanding of 463 MP ingestion in fish even more challenging. Further studies that will consider several parameters are 464 needed to understand the different pathways involved. Increasing the knowledge of those factors will 465 466 help to suggest and define species for the monitoring of plastic pollution in both fresh and marine waters. 467

468

469 4. Freshwater fish for the biomonitoring of MP contamination?

470

The occurrence of a contaminant in an ecosystem does not mean that the contaminant will cause harm 471 to the populations inhabiting the ecosystem. Biota is needed to link the pollutant levels in abiotic 472 473 compartments and in organisms' tissue to potential adverse effects.<sup>143</sup> Using fish species for the assessment of the quality of aquatic ecosystems is highly relevant. Fish are found everywhere in 474 aquatic environments and have an intermediate trophic position, linking the lower trophic positions 475 to the higher trophic positions (Beyer 1996 in van der Oost et al.<sup>143</sup>). Nevertheless, attention must be 476 paid when choosing a species for monitoring. The GESAMP<sup>46</sup> report suggested a list of criteria that 477 helps to define good species for monitoring. This species, or group of species, must be representative 478 of different life histories, phylogenies, sizes, ages and developmental stages. General strategies are 479 proposed, which include opportunistic sampling, market sampling of commercial species and 480 481 targeting biota. In this report, it is stated that fish, as targeted species or those purchased from the markets, could constitute a good monitoring group for the ingestion of MP.<sup>46</sup> 482

Given the numerous factors influencing the ingestion of MP by fish, choosing only one indicator species or a group of species is challenging and perhaps impossible. The ecology and feeding habits of the chosen species must be known. It is important to know where the fish, as well as all its developmental stages, is located within the food web, for instance. Here, we present a summary of the parameters to be considered when choosing an indicator fish species, for both fresh and marine environments:

*Trophic position.* As discussed in the previous section, predators, especially top-predators, are
 more likely to ingest MP through food transfer. However, monitoring MP pollution with top predators will not accurately indicate the contamination level of the environment (water and
 sediment) but will give global information on the trophic web and the ecosystem.

*Feeding strategy.* Ram-feeding fish passively filter water without any visual search for food.
 They may be good indicators of the MP levels in the surrounding water. In this case, in
 addition to gut contents, we advise researchers to also examine the gill rakers, which are bony

or cartilaginous structures attached to the branchial arches that retain filtered particles inside 496 the mouth. Those analyses will complete the GC results. 497

- Migrations. Many fish migrate, either vertically (diel migrations) or horizontally (seasonal 498 \_ migrations), and sometimes fish migrate both ways. Depending on the area chosen to be 499 studied, all types of migration can be valuable. Diel migrations could inform researchers about 500 the MP levels in the pelagic zone, as the surface, seasonal and spatial migrations could provide 501 data for the same species but in different areas. Utilizing the same species is important, as it 502 avoids some bias caused by specific related factors, such as the fish length, the mouth structure 503 and the fish vision. Whatever the chosen species and area, the sampling protocol must be 504 designed considering those migration types. Fish must be sampled at the same time of the day 505 or year if only one location or one water column compartment is targeted. 506
- Commercial value. Collecting commercially valued fish is advantageous as this is a cost-507 \_ effective method for the assessment of human exposure through the chosen species.<sup>46</sup> Indeed, 508 commercial fish can be purchased at markets, which eliminates evident sampling costs. 509 However, it may not be possible to obtain precise information regarding the sampling location. 510 Living habitat. Benthic and demersal species will provide information about a sea \_ 511 compartment that is different than the sea compartment of pelagic species. Demersal fish will 512 give a less accurate overview of the benthic compartment than benthic fish, as demersal fish 513 are not fully associated with the sediment. 514

Distribution. For comparison and repeatability purposes, the chosen species should have the 515 \_ widest distribution possible or should at least belong to a higher taxon (genus or family) that 516 is distributed worldwide distributed in order to avoid biases linked to the different species 517 518 used in various studies.

519

520 In summary, the perfect fish species used to monitor MP pollution in fresh and marine water is either a top-predator or a ram-feeding species with a large distribution area and a commercial value. If the 521

522 species is restricted to a small area, similar species must be found in other areas of the world. The

migration of the selected fish must be well known and considered when sampling. Researchers must
then carefully choose the correct species according to the information they need.

525

526 5. Perspectives and recommendations

527

To provide perspectives for further research, we want to highlight three steps of protocols that are of great importance: the number of samples analyzed, the size limit of the extracted particles and their identification.

Sample sizes vary greatly between studies, as sizes vary from 1 to 330 individuals for a single species.
A sample size of 50 individuals has been determined to be sufficient for plastic pollution research.<sup>53</sup>
A couple of samples is not useful for research; in addition to being statistically irrelevant, time and/or
money can be saved by not sampling such small sample sizes and perhaps used for better purposes.
Achieving 50 individuals per studied criterion can be challenging, and therefore, we recommend
sampling as close to 50 individuals as possible by sampling the same species from a specific location,
for example.

Studies have rarely identified the target size range of MP extracted from fish gut contents (Table 2). 538 The only information available is the mesh size of the filters used during the extraction step when 539 present. Usually, that mesh size is approximately a micrometer, which allows researchers to extract 540 very small particles, but such particles are hardly observed visually and analyzed spectroscopically. 541 In contrast, focusing on large MP might lead to an underestimation of the number of ingested 542 particles. Both large and small fish are likely to ingest small MP ( $< 100 \mu m$ ) either by transfer from 543 544 prey to predator or through their feeding behavior. It is thus critical to specify which threshold is applied in each study to promote comparability. This threshold must be defined according to the 545 identification method used; FTIR and Raman spectroscopies allow researchers to analyze MP that 546 are smaller than the particles visible by visual observation through a stereomicroscope. 547

548 MP have often been identified by visual inspection, although the identification techniques used have 549 seemingly improved in recent studies. Only two studies on MP contamination in freshwater fish 550 included all dyed particles, even if the particles were not made of plastic (Table 1). We believe that 551 it is important to report all types of anthropogenic contamination in fish tissues, especially textile 552 fibers, which are usually dyed and released in large amounts into the environment.<sup>144,145</sup>

With regards to contamination, we suggest that (1) sample collection is performed with as few plastic 553 materials as possible, (2) all solutions, including tap water, must be filtered with filters having the 554 same porosity as the one used to filter the samples and stored in glass jars, (3) plastic materials should 555 be avoided during sample treatment, (4) if the laboratory work cannot be performed in a clean room, 556 the laboratory should then be ventilated as much as possible, (5) all equipment must be covered with 557 clean non-plastic materials when not in use and (6) the laboratory operator should wear nonsynthetic 558 clothes, a cotton lab coat and gloves. Avoiding plastic during the sampling and treatment steps is not 559 easy, especially regarding the textiles worn, solution storage and use of nonsynthetic gloves, but for 560 the latter at least, an FTIR or Raman spectrum can be recorded and matched with the potential 561 contaminating particles found in the samples. If contamination does occur, those particles can then 562 be excluded from the results. 563

As previously recommended by published reviews,<sup>40,47,48</sup> standardized protocols are required. Thus 564 we briefly recommend an overall optimal methodology that can be used when evaluating MP 565 pollution by working with fish gut samples: (1) the number of sampled individuals should be as close 566 to 50 as possible, (2) a MP target size down to 1 µm should be considered to fit the MP definition, 567 (3) enzymes should be used as digesting agents, (4) identification should be performed by 568 spectroscopy, (5) qualitative observations should be made by light microscopy, (6) as many data as 569 570 possible (averages and ranges, global contamination or per category, etc.) should be provided, (7) the results should be expressed using several units and per category studied (feeding strategy, 571 572 developmental stage, etc.), and (8) procedural blank samples should be prepared and analyzed. Spectroscopic analyses can be performed on a subsample of the total amount of extracted particles. 573

However, when focusing on fish contamination, it is uncommon to find tens of particles in a single individual. Therefore, we recommend analyzing all particles extracted from the fish GCs. Visual observations are not recommended for sorting or identifying microplastics but are required to take pictures and to measure the MP found. We encourage researchers to keep that step only for those purposes. The global evaluation of the protocol quality can be made thanks to the quality assessment system proposed by Hermsen et al.<sup>48</sup> This quality assessment has been adapted for MP ingestion from the CRED scoring system<sup>146</sup> and provides a good overview of the protocol strength.

581

This review examines MP in the digestive tracts of freshwater fish, but MP appear to translocate to 582 other organs in several aquatic taxonomic groups, including freshwater fish.<sup>57,61,111,147-149</sup> The impacts 583 of MP translocation are not well known, and MP pathways still need to be understood. The liver has 584 primarily been studied, but other tissues are worth investigating. Based on the human health 585 perspective, muscle is obviously the first tissue that comes to mind, as muscle is almost the only part 586 of the fish we consume. Few studies have examined MP in fish muscle, 57,106,111,147 and when MP are 587 found in the muscle, their pathways are unknown. In this regard, in vitro studies using biofouled 588 plastics are needed to provide more information about the translocation process(es). 589

Selecting one or more species suitable for monitoring is currently challenging. First, studies must be 590 performed on MP ingestion by fish and MP contamination in their nearby environment. Studies 591 should be performed at the same time and, of course, by utilizing proper methods with chemical 592 treatment and identification. This will allow scientists to determine which species reflects the best 593 environmental contamination and thus to decide which species can be defined as an indicator. 594 Furthermore, assessing the species monitoring process is only the first step towards obtaining a more 595 596 global understanding of MP pollution in biota. To date, most studies on MP ingestion by fish are quite descriptive and contain suggested explanations that are usually not demonstrated, and potential 597 impacts are listed. The quantity and quality of research require improvements to be made to assess 598 the ecological risk posed by MP.<sup>150</sup> A description of the contamination levels is obviously of great 599

600	importance to the scientific community, but we think that only studying MP contamination is not
601	sufficient. Microplastics alone have been found to negatively impact fish, <sup>25,151,152</sup> but MPs are part of
602	a cocktail of pollutants containing additives, e.g., phthalates and bisphenols, dyes, organic pollutants,
603	among others. These pollutants are also known to affect fish in several ways. <sup>153-155</sup> Consequently, MP
604	should be monitored concomitantly with other pollutants to more accurately reflect the ecological
605	state of freshwaters with regards to pollution.
606	
607	In conclusion, this review aims to highlight the need for the standardization of methodologies and to
608	give an overview of the factors involved in the ingestion of MP by freshwater fish. The first step of
609	standardization can be to follow several recommendations, such as suggesting that a sufficient
610	number of individuals be sampled and recommending the systematic use of spectroscopy to identify
611	polymer particles. With regards to biomonitoring, studies should be performed using the same overall
612	methodology and the same or very similar fish species. These species must be chosen for the purpose
613	of monitoring and based on their biological characteristics: top-predator and ram-filtering species
614	provide useful information with regards to plastic contamination.
615	
616	6. Supporting information
617	
618	The Supporting Information is available free of charge on the ACS Publications website at DOI:
619	XXX.
620	An overview of the studies reporting plastic contamination in freshwaters and a summary of data
621	regarding MP contamination in abiotic compartments and fish at the same location.
622	
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- 631 8. References
- 632
- 633 (1) Ryan, P. G. A Brief History of Marine Litter Research. In: Marine Anthropogenic Litter; Bergmann,
- 634 M., Gutow, L., Klages, M., Eds.; Springer International Publishing: Cham, 2015.
- 635 https://doi.org/10.1007/978-3-319-16510-3.
- 636 (2) Law, K. L.; Morét-Ferguson, S. E.; Goodwin, D. S.; Zettler, E. R.; Deforce, E.; Kukulka, T.;
- 637 Proskurowski, G. Distribution of Surface Plastic Debris in the Eastern Pacific Ocean from an 11-Year
- 638 Data Set. *Environ. Sci. Technol.* **2014**, *48* (9), 4732–4738. https://doi.org/10.1021/es4053076.
- 639 (3) Gago, J.; Carretero, O.; Filgueiras, A. V.; Viñas, L. Synthetic Microfibers in the Marine
- 640 Environment: A Review on Their Occurrence in Seawater and Sediments. *Mar. Pollut. Bull.* 2018,
- 641 *127*, 365–376. https://doi.org/10.1016/j.marpolbul.2017.11.070.
- 642 (4) Woodall, L. C.; Sanchez-Vidal, A.; Canals, M.; Paterson, G. L. J.; Coppock, R.; Sleight, V.; Calafat,
- A.; Rogers, A. D.; Narayanaswamy, B. E.; Thompson, R. C. The Deep Sea Is a Major Sink for
  Microplastic Debris. *R. Soc. Open Sci.* 2014, *1* (4), 140317–140317.
- 645 https://doi.org/10.1098/rsos.140317.
- 646 (5) Tekman, M. B.; Krumpen, T.; Bergmann, M. Marine Litter on Deep Arctic Seafloor Continues to
  647 Increase and Spreads to the North at the HAUSGARTEN Observatory. *Deep Sea Res. Part I*
- 648 *Oceanogr. Res. Pap.* 2017, *120*, 88–99. https://doi.org/https://doi.org/10.1016/j.dsr.2016.12.011.
- 649 (6) Obbard, R. W.; Sadri, S.; Wong, Y. Q.; Khitun, A. A.; Baker, I.; Thompson, R. C. Global Warming
  650 Releases Microplastic Legacy Frozen in Arctic Sea Ice. *Earth's Futur.* 2014, *2* (6), 315–320.
  651 https://doi.org/10.1002/2014EF000240.
- 652 (7) Peeken, I.; Primpke, S.; Beyer, B.; Gütermann, J.; Katlein, C.; Krumpen, T.; Bergmann, M.;
- Hehemann, L.; Gerdts, G. Arctic Sea Ice Is an Important Temporal Sink and Means of Transport for
  Microplastic. *Nat. Commun.* 2018, 9 (1), 1505. https://doi.org/10.1038/s41467-018-03825-5.
- 655 (8) Liu, M.; Lu, S.; Song, Y.; Lei, L.; Hu, J.; Lv, W.; Zhou, W.; Cao, C.; Shi, H.; Yang, X.; et al.
- 656 Microplastic and Mesoplastic Pollution in Farmland Soils in Suburbs of Shanghai, China. *Environ*.
- 657 *Pollut.* **2018**, *242*, 855–862. https://doi.org/10.1016/j.envpol.2018.07.051.
- 658 (9) Dris, R.; Gasperi, J.; Saad, M.; Mirande, C.; Tassin, B. Synthetic Fibers in Atmospheric Fallout: A
- 659 Source of Microplastics in the Environment? *Mar. Pollut. Bull.* **2016**, *104* (1–2), 290–293.

660 https://doi.org/10.1016/j.marpolbul.2016.01.006.

- 661 (10) Cai, L.; Wang, J.; Peng, J.; Tan, Z.; Zhan, Z.; Tan, X.; Chen, Q. Characteristic of Microplastics in the
  662 Atmospheric Fallout from Dongguan City, China: Preliminary Research and First Evidence. *Environ*.
- 663 Sci. Pollut. Res. 2017, 24 (32), 24928–24935. https://doi.org/10.1007/s11356-017-0116-x.
- 664 (11) *Marine Anthropogenic Litter*; Bergmann, M., Gutow, L., Klages, M., Eds.; Springer International
  665 Publishing: Cham, 2015. https://doi.org/10.1007/978-3-319-16510-3.
- (12) Lusher, A. L.; Hernandez-Milian, G.; O'Brien, J.; Berrow, S.; O'Connor, I.; Officer, R. Microplastic
  and Macroplastic Ingestion by a Deep Diving, Oceanic Cetacean: The True's Beaked Whale
- 668 Mesoplodon Mirus. *Environ. Pollut.* **2015**, *199*, 185–191.
- (13) Nelms, S. E.; Galloway, T. S.; Godley, B. J.; Jarvis, D. S.; Lindeque, P. K. Investigating Microplastic
- Trophic Transfer in Marine Top Predators. *Environ. Pollut.* **2018**, *238*, 999–1007.
- 671 https://doi.org/10.1016/J.ENVPOL.2018.02.016.
- (14) Zhu, J.; Yu, X.; Zhang, Q.; Li, Y.; Tan, S.; Li, D.; Yang, Z.; Wang, J. Cetaceans and Microplastics:
  First Report of Microplastic Ingestion by a Coastal Delphinid, Sousa Chinensis. *Sci. Total Environ.*2019, 659, 649–654. https://doi.org/10.1016/j.scitotenv.2018.12.389.
- 675 (15) Trevail, A. M.; Gabrielsen, G. W.; Kühn, S.; Van Franeker, J. A. Elevated Levels of Ingested Plastic
- 676 in a High Arctic Seabird, the Northern Fulmar (Fulmarus Glacialis). *Polar Biol.* **2015**, *38* (7), 975–
- 677 981. https://doi.org/10.1007/s00300-015-1657-4.
- 678 (16) Zhao, S.; Zhu, L.; Li, D. Microscopic Anthropogenic Litter in Terrestrial Birds from Shanghai,
- 679 China: Not Only Plastics but Also Natural Fibers. *Sci. Total Environ.* **2016**, *550*, 1110–1115.
- 680 https://doi.org/10.1016/j.scitotenv.2016.01.112.
- (17) Reynolds, C.; Ryan, P. G. Micro-Plastic Ingestion by Waterbirds from Contaminated Wetlands in
  South Africa. *Mar. Pollut. Bull.* 2018, *126*, 330–333.
- 683 https://doi.org/10.1016/J.MARPOLBUL.2017.11.021.
- 684 (18) Devriese, L. I.; van der Meulen, M. D.; Maes, T.; Bekaert, K.; Paul-Pont, I.; Frère, L.; Robbens, J.;
- 685 Vethaak, A. D. Microplastic Contamination in Brown Shrimp (Crangon Crangon, Linnaeus 1758)
- from Coastal Waters of the Southern North Sea and Channel Area. Mar. Pollut. Bull. 2015, 98 (1),
- 687 179–187. https://doi.org/10.1016/j.marpolbul.2015.06.051.
- 688 (19) Iannilli, V.; Di Gennaro, A.; Lecce, F.; Sighicelli, M.; Falconieri, M.; Pietrelli, L.; Poeta, G.; Battisti,

- 689 C. Microplastics in Talitrus Saltator (Crustacea, Amphipoda): New Evidence of Ingestion from
- 690 Natural Contexts. *Environ. Sci. Pollut. Res.* **2018**, *25* (28), 28725–28729.
- 691 https://doi.org/10.1007/s11356-018-2932-z.
- (20) Hurley, R. R.; Woodward, J. C.; Rothwell, J. J. Ingestion of Microplastics by Freshwater Tubifex
  Worms. *Environ. Sci. Technol.* 2017, *51* (21), 12844–12851. https://doi.org/10.1021/acs.est.7b03567.
- 694 (21) Bråte, I. L. N.; Eidsvoll, D. P.; Steindal, C. C.; Thomas, K. V. Plastic Ingestion by Atlantic Cod (
- Gadus Morhua ) from the Norwegian Coast. *Mar. Pollut. Bull.* **2016**, *112* (1–2), 105–110.

696 https://doi.org/10.1016/j.marpolbul.2016.08.034.

- 697 (22) Rummel, C. D.; Löder, M. G. J.; Fricke, N. F.; Lang, T.; Griebeler, E.-M.; Janke, M.; Gerdts, G.
- 698 Plastic Ingestion by Pelagic and Demersal Fish from the North Sea and Baltic Sea. *Mar. Pollut. Bull.*

**2016**, *102* (1), 134–141. https://doi.org/10.1016/j.marpolbul.2015.11.043.

- (23) Collicutt, B.; Juanes, F.; Dudas, S. E. Microplastics in Juvenile Chinook Salmon and Their Nearshore
  Environments on the East Coast of Vancouver Island. *Environ. Pollut.* 2019, *244*, 135–142.
  https://doi.org/10.1016/j.envpol.2018.09.137.
- 703 (24) Arthur, C.; Baker, J.; Bamford, H. Proceedings of the International Research Workshop on the
- 704 Occurrence, Effects and Fate of Microplastic Marine Debris. In *Proceedings of the International*
- 705 *Research Workshop on the Occurrence, Effects and Fate of Microplastic Marine Debris.*; Technical
- 706 Memorandum NOS-OR&R-30; Proceedings of the International Research Workshop on the
- 707 Occurrence, Effects, and Fate of Microplastic Marine Debris Courtney Arthur, Joel Baker, and
- Holly Bamford (eds.) National Oceanic and Atmospheric Administration: University of Washington
  Tacoma campus in Tacoma, Washington, USA, 2008.
- 710 (25) Anbumani, S.; Kakkar, P. Ecotoxicological Effects of Microplastics on Biota: A Review. *Environ*.
- 711 Sci. Pollut. Res. 2018, 25 (15), 14373–14396. https://doi.org/10.1007/s11356-018-1999-x.
- 712 (26) Barboza, L. G. A.; Dick Vethaak, A.; Lavorante, B. R. B. O.; Lundebye, A.-K.; Guilhermino, L.
- 713 Marine Microplastic Debris: An Emerging Issue for Food Security, Food Safety and Human Health.
- 714 *Mar. Pollut. Bull.* **2018**, *133*, 336–348. https://doi.org/10.1016/j.marpolbul.2018.05.047.
- 715 (27) Choi, J. S.; Jung, Y.-J.; Hong, N.-H.; Hong, S. H.; Park, J.-W. Toxicological Effects of Irregularly
- 716 Shaped and Spherical Microplastics in a Marine Teleost, the Sheepshead Minnow (Cyprinodon
- 717 Variegatus). *Mar. Pollut. Bull.* **2018**, *129* (1), 231–240.

718 https://doi.org/10.1016/j.marpolbul.2018.02.039.

- 719 (28) de Sá, L. C.; Luís, L. G.; Guilhermino, L. Effects of Microplastics on Juveniles of the Common Goby
- 720 (Pomatoschistus Microps): Confusion with Prey, Reduction of the Predatory Performance and
- 721 Efficiency, and Possible Influence of Developmental Conditions. *Environ. Pollut.* 2015, *196*, 359–

722 362. https://doi.org/10.1016/j.envpol.2014.10.026.

- 723 (29) Small, C.; Nicholls, R. J. A Global Analysis of Human Settlement in Coastal Zones. *J. Coast. Res.*724 2003, 19 (3), 584–599.
- 725 (30) Kummu, M.; de Moel, H.; Ward, P. J.; Varis, O. How Close Do We Live to Water? A Global
- Analysis of Population Distance to Freshwater Bodies. *PLoS One* **2011**, *6* (6), e20578.

727 https://doi.org/10.1371/journal.pone.0020578.

- (31) Lebreton, L. C. M.; van der Zwet, J.; Damsteeg, J.-W.; Slat, B.; Andrady, A.; Reisser, J. River Plastic
  Emissions to the World's Oceans. *Nat. Commun.* 2017, *8*, 15611.
- 730 https://doi.org/10.1038/ncomms15611.
- (32) Carr, S. A.; Liu, J.; Tesoro, A. G. Transport and Fate of Microplastic Particles in Wastewater
   Treatment Plants. *Water Res.* 2016, *91*, 174–182. https://doi.org/10.1016/j.watres.2016.01.002.
- 733 (33) Mason, S. A.; Garneau, D.; Sutton, R.; Chu, Y.; Ehmann, K.; Barnes, J.; Fink, P.; Papazissimos, D.;
- Rogers, D. L. Microplastic Pollution Is Widely Detected in US Municipal Wastewater Treatment

735 Plant Effluent. *Environ. Pollut.* **2016**, *218*, 1045–1054. https://doi.org/10.1016/j.envpol.2016.08.056.

- 736 (34) Free, C. M.; Jensen, O. P.; Mason, S. A.; Eriksen, M.; Williamson, N. J.; Boldgiv, B. High-Levels of
- Microplastic Pollution in a Large, Remote, Mountain Lake. *Mar. Pollut. Bull.* 2014, *85* (1), 156–163.
  https://doi.org/10.1016/j.marpolbul.2014.06.001.
- (35) Baldwin, A. K.; Corsi, S. R.; Mason, S. A. Plastic Debris in 29 Great Lakes Tributaries: Relations to
  Watershed Attributes and Hydrology. *Environ. Sci. Technol.* 2016, *50* (19), 10377–10385.
- 741 https://doi.org/10.1021/acs.est.6b02917.
- (36) Eriksen, M.; Maximenko, N.; Thiel, M.; Cummins, A.; Lattin, G.; Wilson, S.; Hafner, J.; Zellers, A.;
  Rifman, S. Plastic Pollution in the South Pacific Subtropical Gyre. *Mar. Pollut. Bull.* 2013, *68* (1–2),
- **744** 71–76.
- (37) Gasperi, J.; Dris, R.; Bonin, T.; Rocher, V.; Tassin, B. Assessment of Floating Plastic Debris in
  Surface Water along the Seine River. *Environ. Pollut.* 2014, *195*, 163–166.

747 https://doi.org/10.1016/j.envpol.2014.09.001.

- 748 (38) Lechner, A.; Keckeis, H.; Lumesberger-Loisl, F.; Zens, B.; Krusch, R.; Tritthart, M.; Glas, M.;
- 749 Schludermann, E. The Danube so Colourful: A Potpourri of Plastic Litter Outnumbers Fish Larvae in
- Europe's Second Largest River. *Environ. Pollut.* **2014**, *188*, 177–181.
- 751 https://doi.org/http://dx.doi.org/10.1016/j.envpol.2014.02.006.
- 752 (39) Dris, R.; Imhof, H.; Sanchez, W.; Gasperi, J.; Galgani, F.; Tassin, B.; Laforsch, C. Beyond the Ocean:
- 753 Contamination of Freshwater Ecosystems with (Micro-)Plastic Particles. *Environ. Chem.* 2015, *12*
- 754 (5), 539–550. https://doi.org/10.1071/EN14172.
- 755 (40) O'Connor, J. D.; Mahon, A. M.; Ramsperger, A. F. R. M.; Trotter, B.; Redondo-Hasselerharm, P. E.;
- 756 Koelmans, A. A.; Lally, H. T.; Murphy, S. Microplastics in Freshwater Biota: A Critical Review of
- Isolation, Characterization, and Assessment Methods. *Glob. Challenges* **2019**, 1800118.
- 758 https://doi.org/10.1002/gch2.201800118.
- 759 (41) Powers, D. Fish as Model Systems. *Science (80-. ).* **1989**, *246* (4928), 352–358.
- 760 https://doi.org/10.1126/science.2678474.
- 761 (42) Nelson, J. S. Fishes of the World; John Wiley & Sons, Inc.: Hoboken, 2006.
- 762 (43) Markic, A.; Gaertner, J.-C.; Gaertner-Mazouni, N.; Koelmans, A. A. Plastic Ingestion by Marine Fish
- in the Wild. Crit. Rev. Environ. Sci. Technol. 2019, 1–41.
- 764 https://doi.org/10.1080/10643389.2019.1631990.
- 765 (44) Jabeen, K.; Su, L.; Li, J.; Yang, D.; Tong, C.; Mu, J.; Shi, H. Microplastics and Mesoplastics in Fish
- from Coastal and Fresh Waters of China. *Environ. Pollut.* **2017**, *221*, 141–149.
- 767 https://doi.org/http://dx.doi.org/10.1016/j.envpol.2016.11.055.
- 768 (45) Murphy, F.; Russell, M.; Ewins, C.; Quinn, B. The Uptake of Macroplastic & Microplastic by
- 769 Demersal & Pelagic Fish in the Northeast Atlantic around Scotland. Mar. Pollut. Bull. 2017, 122 (1–
- 2), 353–359. https://doi.org/10.1016/j.marpolbul.2017.06.073.
- (46) GESAMP. Guidelines or the Monitoring and Assessment of Plastic Litter and Microplastics in the
   Ocean; Kershaw, P. J., Turra, A., Galgani, F., Eds.; 2019.
- (47) Lusher, A. L.; Welden, N. A.; Sobral, P.; Cole, M. Sampling, Isolating and Identifying Microplastics
  Ingested by Fish and Invertebrates. *Anal. Methods* 2017, *9* (9), 1346–1360.
- 775 https://doi.org/10.1039/C6AY02415G.

- (48) Hermsen, E.; Mintenig, S. M.; Besseling, E.; Koelmans, A. A. Quality Criteria for the Analysis of
- Microplastic in Biota Samples: A Critical Review. *Environ. Sci. Technol.* 2018, *52* (18), 10230–
  10240. https://doi.org/10.1021/acs.est.8b01611.
- (49) Dantas, D. V; Barletta, M.; da Costa, M. F. The Seasonal and Spatial Patterns of Ingestion of
  Polyfilament Nylon Fragments by Estuarine Drums (Sciaenidae). *Environ. Sci. Pollut. Res.* 2012, *19*(2), 600–606.
- (50) Ramos, J.; Barletta, M.; Costa, M. Ingestion of Nylon Threads by Gerreidae While Using a Tropical
  Estuary as Foraging Grounds. *Aquat. Biol.* 2012, *17* (1), 29–34. https://doi.org/10.3354/ab00461.
- Vendel, A. L.; Bessa, F.; Alves, V. E. N.; Amorim, A. L. A.; Patrício, J.; Palma, A. R. T. Widespread
   Microplastic Ingestion by Fish Assemblages in Tropical Estuaries Subjected to Anthropogenic
- 786 Pressures. *Mar. Pollut. Bull.* **2017**, *117* (1–2), 448–455.
- 787 https://doi.org/10.1016/j.marpolbul.2017.01.081.
- 788 (52) Pegado, T. de S. e S.; Schmid, K.; Winemiller, K. O.; Chelazzi, D.; Cincinelli, A.; Dei, L.; Giarrizzo,
  789 T. First Evidence of Microplastic Ingestion by Fishes from the Amazon River Estuary. *Mar. Pollut.*790 *Bull.* 2018, *133*, 814–821. https://doi.org/10.1016/J.MARPOLBUL.2018.06.035.
- (53) Galgani, F.; Hanke, G.; Werner, S.; Oosterbaan, L.; Nilsson, P.; Fleet, D.; Kinsey, S.; Thompson, R.
  C.; van Franeker, J.; Vlachogianna, T.; et al. *Guidance on Monitoring of Marine Litter in European Seas*; 2013.
- (54) ICES. OSPAR Request on Development of a Common Monitoring Protocol for Plastic Particles in
   Fish Stomachs and Selected Shellfish on the Basis of Existing Fish Disease Surveys; 2015.
- (55) Van Franeker, J.; Meijboom, A. *Litter NSV, Marine Litter Monitoring by Northern Fulmars; a Pilot Study*; Wageningen, 2002.
- 798 (56) Pazos, R. S.; Maiztegui, T.; Colautti, D. C.; Paracampo, A. H.; Gómez, N. Microplastics in Gut
- 799 Contents of Coastal Freshwater Fish from Río de La Plata Estuary. *Mar. Pollut. Bull.* **2017**, *122* (1–
- 2), 85–90. https://doi.org/10.1016/j.marpolbul.2017.06.007.
- 801 (57) Collard, F.; Gasperi, J.; Gilbert, B.; Eppe, G.; Azimi, S.; Rocher, V.; Tassin, B. Anthropogenic
- 802 Particles in the Stomach Contents and Liver of the Freshwater Fish Squalius Cephalus. *Sci. Total*
- 803 *Environ.* **2018**, *643*. https://doi.org/10.1016/j.scitotenv.2018.06.313.
- 804 (58) Dehaut, A.; Cassone, A.-L.; Frère, L.; Hermabessiere, L.; Himber, C.; Rinnert, E.; Rivière, G.;

- 805 Lambert, C.; Soudant, P.; Huvet, A.; et al. Microplastics in Seafood: Benchmark Protocol for Their
- 806 Extraction and Characterization. *Environ. Pollut.* **2016**, *215*, 223–233.
- 807 https://doi.org/10.1016/j.envpol.2016.05.018.
- 808 (59) Zhang, K.; Xiong, X.; Hu, H.; Wu, C.; Bi, Y.; Wu, Y.; Zhou, B.; Lam, P. K. S.; Liu, J. Occurrence
  809 and Characteristics of Microplastic Pollution in Xiangxi Bay of Three Gorges Reservoir, China.
  810 *Environ. Sci. Technol.* 2017, *51* (7), 3794–3801. https://doi.org/10.1021/acs.est.7b00369.
- (60) Claessens, M.; Van Cauwenberghe, L.; Vandegehuchte, M. B.; Janssen, C. R. New Techniques for
  the Detection of Microplastics in Sediments and Field Collected Organisms. *Mar. Pollut. Bull.* 2013,
  70 (1–2), 227–233.
- 814 (61) Avio, C. G.; Gorbi, S.; Regoli, F. Experimental Development of a New Protocol for Extraction and
  815 Characterization of Microplastics in Fish Tissues: First Observations in Commercial Species from
- 816 Adriatic Sea. *Mar. Environ. Res.* **2015**, *111*, 18–26. https://doi.org/10.1016/j.marenvres.2015.06.014.
- (62) Catarino, A. I.; Thompson, R.; Sanderson, W.; Henry, T. B. Development and Optimization of a
  Standard Method for Extraction of Microplastics in Mussels by Enzyme Digestion of Soft Tissues. *Environ. Toxicol. Chem.* 2017, *36* (4), 947–951. https://doi.org/10.1002/etc.3608.
- (63) Cole, M.; Webb, H.; Lindeque, P. K.; Fileman, E. S.; Halsband, C.; Galloway, T. S. Isolation of
  Microplastics in Biota-Rich Seawater Samples and Marine Organisms. *Sci. Rep.* 2014, *4*, 4528.
  https://doi.org/10.1038/srep04528.
- (64) Railo, S.; Talvitie, J.; Setälä, O.; Koistinen, A.; Lehtiniemi, M. Application of an Enzyme Digestion
  Method Reveals Microlitter in Mytilus Trossulus at a Wastewater Discharge Area. *Mar. Pollut. Bull.*2018, *130*, 206–214. https://doi.org/10.1016/j.marpolbul.2018.03.022.
- (65) von Friesen, L. W.; Granberg, M. E.; Hassellöv, M.; Gabrielsen, G. W.; Magnusson, K. An Efficient
   and Gentle Enzymatic Digestion Protocol for the Extraction of Microplastics from Bivalve Tissue.
- 828 *Mar. Pollut. Bull.* **2019**, *142*, 129–134. https://doi.org/10.1016/j.marpolbul.2019.03.016.
- 829 (66) Tagg, A. S.; Harrison, J. P.; Ju-Nam, Y.; Sapp, M.; Bradley, E. L.; Sinclair, C. J.; Ojeda, J. J.
- 830 Fenton's Reagent for the Rapid and Efficient Isolation of Microplastics from Wastewater. *Chem.*
- 831 *Commun.* **2017**, *53* (2), 372–375. https://doi.org/10.1039/C6CC08798A.
- 832 (67) Horton, A. A.; Jürgens, M. D.; Lahive, E.; van Bodegom, P. M.; Vijver, M. G. The Influence of
- 833 Exposure and Physiology on Microplastic Ingestion by the Freshwater Fish Rutilus Rutilus (Roach)

834 in the River Thames, UK. *Environ. Pollut.* **2018**, *236*, 188–194.

- 835 https://doi.org/10.1016/J.ENVPOL.2018.01.044.
- (68) Collard, F.; Gilbert, B.; Eppe, G.; Parmentier, E.; Das, K. Detection of Anthropogenic Particles in
  Fish Stomachs: An Isolation Method Adapted to Identification by Raman Spectroscopy. *Arch.*
- 838 *Environ. Contam. Toxicol.* **2015**, *69* (3), 331–339. https://doi.org/10.1007/s00244-015-0221-0.
- 839 (69) Kühn, S.; van Werven, B.; van Oyen, A.; Meijboom, A.; Bravo Rebolledo, E. L.; van Franeker, J. A.
- 840 The Use of Potassium Hydroxide (KOH) Solution as a Suitable Approach to Isolate Plastics Ingested
- 841 by Marine Organisms. *Mar. Pollut. Bull.* **2017**, *115* (1–2), 86–90.
- 842 https://doi.org/10.1016/j.marpolbul.2016.11.034.
- 843 (70) Barrows, A. P. W.; Cathey, S. E.; Petersen, C. W. Marine Environment Microfiber Contamination:
- Global Patterns and the Diversity of Microparticle Origins. *Environ. Pollut.* **2018**, *237*, 275–284.
- 845 https://doi.org/10.1016/J.ENVPOL.2018.02.062.
- 846 (71) Darbra, R. M.; Gonzalez Dan, J. R.; Casal, J.; Agueda, A.; Capri, E.; Fait, G.; Schuhmacher, M.;
- 847 Nadal, M.; Rovira, J.; Grundmann, V.; et al. Additives in the Textile Industry. In *Global Risk-Based*848 *Management of Chemical Additives I. Production, usage, and environmental occurrence*; Bilitewski,
- 849 B., Darbra, R. M., Barceló, D., Eds.; The Handbook of Environmental Chemistry; Springer Berlin
- 850 Heidelberg: Berlin, Heidelberg, 2011; Vol. 18, p 290. https://doi.org/10.1016/0143-1471(82)90111-8.
- 851 (72) Robinson, T.; McMullan, G.; Marchant, R.; Nigam, P. Remediation of Dyes in Textile Effluent: A
- 852 Critical Review on Current Treatment Technologies with a Proposed Alternative. *Bioresour. Technol.*
- **2001**, 77 (3), 247–255. https://doi.org/10.1016/S0960-8524(00)00080-8.
- Pokharia, A.; Singh, S. Toxicological Effect of Textile Dyes and Their Metabolites: A Review. *Curr. Trends Biotechnol. Chem. Res.* 2015, 5 (1), 11–17.
- 856 (74) Remy, F.; Collard, F.; Gilbert, B.; Compère, P.; Eppe, G.; Lepoint, G. When Microplastic Is Not
- 857 Plastic: The Ingestion of Artificial Cellulose Fibers by Macrofauna Living in Seagrass
- 858 Macrophytodetritus. *Environ. Sci. Technol.* **2015**, *49* (18), 11158–11166.
- 859 https://doi.org/10.1021/acs.est.5b02005.
- (75) Andrady, A. L. Microplastics in the Marine Environment. *Mar. Pollut. Bull.* 2011, *62* (8), 1596–1605.
   https://doi.org/10.1016/j.marpolbul.2011.05.030.
- 862 (76) Renner, G.; Schmidt, T. C.; Schram, J. Analytical Methodologies for Monitoring

- 863 Micro(Nano)Plastics: Which Are Fit for Purpose? *Curr. Opin. Environ. Sci. Heal.* **2018**, *1*, 55–61.
- 864 https://doi.org/10.1016/j.coesh.2017.11.001.
- 865 (77) Collard, F.; Gilbert, B.; Eppe, G.; Roos, L.; Compère, P.; Das, K.; Parmentier, E. Morphology of the
  866 Filtration Apparatus of Three Planktivorous Fishes and Relation with Ingested Anthropogenic
- 867 Particles. *Mar. Pollut. Bull.* **2017**, *116* (1–2), 182–191.
- 868 https://doi.org/10.1016/j.marpolbul.2016.12.067.
- 869 (78) Löder, M. G. J.; Kuczera, M.; Mintenig, S.; Lorenz, C.; Gerdts, G. Focal Plane Array Detector-Based
- 870 Micro-Fourier-Transform Infrared Imaging for the Analysis of Microplastics in Environmental
- 871 Samples. *Environ. Chem.* **2015**, *12* (5), 563. https://doi.org/10.1071/EN14205.
- 872 (79) Song, Y. K.; Hong, S. H.; Jang, M.; Han, G. M.; Rani, M.; Lee, J.; Shim, W. J. A Comparison of
- 873 Microscopic and Spectroscopic Identification Methods for Analysis of Microplastics in
- 874 Environmental Samples. *Mar. Pollut. Bull.* **2015**, *93* (1–2), 202–209.
- 875 https://doi.org/10.1016/j.marpolbul.2015.01.015.
- 876 (80) GESAMP. Sources, Fate and Effects of Microplastics in the Marine Environment: Part Two of a
  877 Global Assessment; Kershaw, P. J., Rochman, C. M., Eds.; Rep. Stud. GESAMP No. 93, 2016.
- 878 (81) Roch, S.; Walter, T.; Ittner, L. D.; Friedrich, C.; Brinker, A. A Systematic Study of the Microplastic
- 879 Burden in Freshwater Fishes of South-Western Germany Are We Searching at the Right Scale? *Sci.*

880 *Total Environ.* **2019**, *689*, 1001–1011. https://doi.org/10.1016/j.scitotenv.2019.06.404.

- 881 (82) Vandermeersch, G.; Van Cauwenberghe, L.; Janssen, C. R.; Marques, A.; Granby, K.; Fait, G.;
- 882 Kotterman, M. J. J.; Diogène, J.; Bekaert, K.; Robbens, J.; et al. A Critical View on Microplastic
- 883 Quantification in Aquatic Organisms. *Environ. Res.* **2015**, *143*, 46–55.
- 884 https://doi.org/10.1016/j.envres.2015.07.016.
- 885 (83) Dümichen, E.; Barthel, A.-K.; Braun, U.; Bannick, C. G.; Brand, K.; Jekel, M.; Senz, R. Analysis of
- 886 Polyethylene Microplastics in Environmental Samples, Using a Thermal Decomposition Method.
- 887 *Water Res.* **2015**, *85*, 451–457. https://doi.org/10.1016/j.watres.2015.09.002.
- 888 (84) Käppler, A.; Fischer, D.; Oberbeckmann, S.; Schernewski, G.; Labrenz, M.; Eichhorn, K.-J.; Voit, B.
- Analysis of Environmental Microplastics by Vibrational Microspectroscopy: FTIR, Raman or Both?
- Anal. Bioanal. Chem. **2016**, 408 (29), 8377–8391. https://doi.org/10.1007/s00216-016-9956-3.
- 891 (85) Araujo, C. F.; Nolasco, M. M.; Ribeiro, A. M. P.; Ribeiro-Claro, P. J. A. Identification of

- 892 Microplastics Using Raman Spectroscopy: Latest Developments and Future Prospects. *Water Res.*
- **2018**, *142*, 426–440. https://doi.org/10.1016/J.WATRES.2018.05.060.
- 894 (86) Steer, M.; Cole, M.; Thompson, R. C.; Lindeque, P. K. Microplastic Ingestion in Fish Larvae in the
  895 Western English Channel. *Environ. Pollut.* 2017, *226*, 250–259.
- 896 https://doi.org/10.1016/j.envpol.2017.03.062.
- 897 (87) Bessa, F.; Barría, P.; Neto, J. M.; Frias, J. P. G. L.; Otero, V.; Sobral, P.; Marques, J. C. Occurrence
- of Microplastics in Commercial Fish from a Natural Estuarine Environment. Mar. Pollut. Bull. 2018,
- 899 *128*, 575–584. https://doi.org/10.1016/j.marpolbul.2018.01.044.
- 900 (88) Wieczorek, A. M.; Morrison, L.; Croot, P. L.; Allcock, A. L.; MacLoughlin, E.; Savard, O.;
- Brownlow, H.; Doyle, T. K. Frequency of Microplastics in Mesopelagic Fishes from the Northwest
  Atlantic. *Front. Mar. Sci.* 2018, *5*. https://doi.org/10.3389/fmars.2018.00039.
- 903 (89) Watts, A. J. R.; Lewis, C.; Goodhead, R. M.; Beckett, S. J.; Moger, J.; Tyler, C. R.; Galloway, T. S.
  904 Uptake and Retention of Microplastics by the Shore Crab Carcinus Maenas. *Environ. Sci. Technol.*905 2014, 48 (15), 8823–8830.
- 906 (90) Grigorakis, S.; Mason, S. A.; Drouillard, K. G. Determination of the Gut Retention of Plastic
- 907 Microbeads and Microfibers in Goldfish (Carassius Auratus). *Chemosphere* 2017, *169*, 233–238.
  908 https://doi.org/http://dx.doi.org/10.1016/j.chemosphere.2016.11.055.
- 909 (91) McNeish, R. E.; Kim, L. H.; Barrett, H. A.; Mason, S. A.; Kelly, J. J.; Hoellein, T. J. Microplastic in
  910 Riverine Fish Is Connected to Species Traits. *Sci. Rep.* 2018, *8* (1), 11639.
- 911 https://doi.org/10.1038/s41598-018-29980-9.
- 912 (92) Wesch, C.; Elert, A. M.; Wörner, M.; Braun, U.; Klein, R.; Paulus, M. Assuring Quality in
- 913 Microplastic Monitoring: About the Value of Clean-Air Devices as Essentials for Verified Data. *Sci.*
- 914 *Rep.* **2017**, *7* (1), 5424. https://doi.org/10.1038/s41598-017-05838-4.
- (93) Kosuth, M.; Mason, S. A.; Wattenberg, E. V. Anthropogenic Contamination of Tap Water, Beer, and
  Sea Salt. *PLoS One* 2018, *13* (4), e0194970. https://doi.org/10.1371/journal.pone.0194970.
- 917 (94) Yang, D.; Shi, H.; Li, L.; Li, J.; Jabeen, K.; Kolandhasamy, P. Microplastic Pollution in Table Salts
- 918 from China. *Environ. Pollut.* **2015**, *49*, 13622–13627. https://doi.org/10.1021/acs.est.5b03163.
- 919 (95) Iñiguez, M. E.; Conesa, J. A.; Fullana, A. Microplastics in Spanish Table Salt. Sci. Rep. 2017, 7 (1),
- 920 8620. https://doi.org/10.1038/s41598-017-09128-x.

- 921 (96) Campbell, S. H.; Williamson, P. R.; Hall, B. D. Microplastics in the Gastrointestinal Tracts of Fish
- 922 and the Water from an Urban Prairie Creek. *FACETS* 2017, 2 (1), 395–409.
- 923 https://doi.org/10.1139/facets-2017-0008.
- 924 (97) Possatto, F. E.; Barletta, M.; Costa, M. F.; do Sul, J. A. I.; Dantas, D. V. Plastic Debris Ingestion by
- 925 Marine Catfish: An Unexpected Fisheries Impact. *Mar. Pollut. Bull.* **2011**, *62* (5), 1098–1102.
- 926 https://doi.org/10.1016/j.marpolbul.2011.01.036.
- 927 (98) Sanchez, W.; Bender, C.; Porcher, J.-M. Wild Gudgeons (Gobio Gobio) from French Rivers Are
  928 Contaminated by Microplastics: Preliminary Study and First Evidence. *Environ. Res.* 2014, *128*, 98–
  929 100. https://doi.org/10.1016/j.envres.2013.11.004.
- 930 (99) Biginagwa, F. J.; Mayoma, B. S.; Shashoua, Y.; Syberg, K.; Khan, F. R. First Evidence of
- 931 Microplastics in the African Great Lakes: Recovery from Lake Victoria Nile Perch and Nile Tilapia.
- 932 *J. Great Lakes Res.* **2016**, *42* (1). https://doi.org/10.1016/j.jglr.2015.10.012.
- 933 (100) Faure, F.; Saini, C.; Potter, G.; Galgani, F.; de Alencastro, L. F.; Hagmann, P. An Evaluation of
  934 Surface Micro- and Mesoplastic Pollution in Pelagic Ecosystems of the Western Mediterranean Sea.
  935 *Environ. Sci. Pollut. Res. Int.* 2015. https://doi.org/10.1007/s11356-015-4453-3.
- 936 (101) Phillips, M. B.; Bonner, T. H. Occurrence and Amount of Microplastic Ingested by Fishes in
  937 Watersheds of the Gulf of Mexico. *Mar. Pollut. Bull.* 2015.
- 938 https://doi.org/10.1016/j.marpolbul.2015.08.041.
- 939 (102) Ferreira, G. V. B.; Barletta, M.; Lima, A. R. A.; Dantas, D. V.; Justino, A. K. S.; Costa, M. F. Plastic
- 940 Debris Contamination in the Life Cycle of Acoupa Weakfish (Cynoscion Acoupa ) in a Tropical
- 941 Estuary. ICES J. Mar. Sci. J. du Cons. 2016, 73 (10), 2695–2707.
- 942 https://doi.org/10.1093/icesjms/fsw108.
- 943 (103) Peters, C. A.; Bratton, S. P. Urbanization Is a Major Influence on Microplastic Ingestion by Sunfish
- in the Brazos River Basin, Central Texas, USA. *Environ. Pollut.* **2016**, *210*, 380–387.
- 945 https://doi.org/10.1016/j.envpol.2016.01.018.
- 946 (104) McGoran, A. R.; Clark, P. F.; Morritt, D. Presence of Microplastic in the Digestive Tracts of
- 947 European Flounder, Platichthys Flesus, and European Smelt, Osmerus Eperlanus, from the River
- 948 Thames. *Environ. Pollut.* **2017**, *220*, 744–751. https://doi.org/10.1016/j.envpol.2016.09.078.
- 949 (105) Silva-Cavalcanti, J. S.; Silva, J. D. B.; França, E. J. de; Araújo, M. C. B. de; Gusmão, F.

- 950 Microplastics Ingestion by a Common Tropical Freshwater Fishing Resource. *Environ. Pollut.* 2017,
- 951 *221*, 218–226. https://doi.org/10.1016/j.envpol.2016.11.068.
- 952 (106) Abbasi, S.; Soltani, N.; Keshavarzi, B.; Moore, F.; Turner, A.; Hassanaghaei, M. Microplastics in
- 953 Different Tissues of Fish and Prawn from the Musa Estuary, Persian Gulf. *Chemosphere* **2018**, *205*,

954 80–87. https://doi.org/10.1016/j.chemosphere.2018.04.076.

- 955 (107) Xiong, X.; Zhang, K.; Chen, X.; Shi, H.; Luo, Z.; Wu, C. Sources and Distribution of Microplastics in
- 956 China's Largest Inland Lake Qinghai Lake. *Environ. Pollut.* **2018**, *235*, 899–906.
- 957 https://doi.org/10.1016/j.envpol.2017.12.081.
- 958 (108) Andrade, M. C.; Winemiller, K. O.; Barbosa, P. S.; Fortunati, A.; Chelazzi, D.; Cincinelli, A.;
- 959 Giarrizzo, T. First Account of Plastic Pollution Impacting Freshwater Fishes in the Amazon:
- 960 Ingestion of Plastic Debris by Piranhas and Other Serrasalmids with Diverse Feeding Habits.
- 961 *Environ. Pollut.* **2019**, *244*, 766–773. https://doi.org/10.1016/j.envpol.2018.10.088.
- 962 (109) Slootmaekers, B.; Catarci Carteny, C.; Belpaire, C.; Saverwyns, S.; Fremout, W.; Blust, R.; Bervoets,
  963 L. Microplastic Contamination in Gudgeons (Gobio Gobio) from Flemish Rivers (Belgium). *Environ.*

964 *Pollut.* **2019**, *244*, 675–684. https://doi.org/10.1016/j.envpol.2018.09.136.

- 965 (110) Yuan, W.; Liu, X.; Wang, W.; Di, M.; Wang, J. Microplastic Abundance, Distribution and
- 966 Composition in Water, Sediments, and Wild Fish from Poyang Lake, China. *Ecotoxicol. Environ. Saf.*
- **2019**, *170*, 180–187. https://doi.org/10.1016/j.ecoenv.2018.11.126.
- 968 (111) Su, L.; Deng, H.; Li, B.; Chen, Q.; Pettigrove, V.; Wu, C.; Shi, H. The Occurrence of Microplastic in
  969 Specific Organs in Commercially Caught Fishes from Coast and Estuary Area of East China. J.

970 *Hazard. Mater.* **2019**, *365*, 716–724. https://doi.org/10.1016/j.jhazmat.2018.11.024.

- 971 (112) Ferreira, G. V. B.; Barletta, M.; Lima, A. R. A. Use of Estuarine Resources by Top Predator Fishes.
- 972 How Do Ecological Patterns Affect Rates of Contamination by Microplastics? *Sci. Total Environ.*
- **2019**, *655*, 292–304. https://doi.org/10.1016/j.scitotenv.2018.11.229.
- 974 (113) Arias, A. H.; Ronda, A. C.; Oliva, A. L.; Marcovecchio, J. E. Evidence of Microplastic Ingestion by
  975 Fish from the Bahía Blanca Estuary in Argentina, South America. *Bull. Environ. Contam. Toxicol.*
- **2019**. https://doi.org/10.1007/s00128-019-02604-2.
- 977 (114) Su, L.; Nan, B.; Hassell, K. L.; Craig, N. J.; Pettigrove, V. Microplastics Biomonitoring in Australian
- 978 Urban Wetlands Using a Common Noxious Fish (Gambusia Holbrooki). *Chemosphere* **2019**, *228*,

- 979 65–74. https://doi.org/10.1016/j.chemosphere.2019.04.114.
- 980 (115) Hidalgo-Ruz, V.; Gutow, L.; Thompson, R. C.; Thiel, M. Microplastics in the Marine Environment:
  981 A Review of the Methods Used for Identification and Quantification. *Environ. Sci. Technol.* 2012, *46*982 (6), 3060–3075.
- (116) Zobkov, M. B.; Esiukova, E. E. Microplastics in a Marine Environment: Review of Methods for
  Sampling, Processing, and Analyzing Microplastics in Water, Bottom Sediments, and Coastal
  Deposits. *Oceanology* 2018, *58* (1), 137–143. https://doi.org/10.1134/S0001437017060169.
- 986 (117) Rodrigues, M. O.; Abrantes, N.; Gonçalves, F. J. M.; Nogueira, H.; Marques, J. C.; Gonçalves, A. M.
- 987 M. Spatial and Temporal Distribution of Microplastics in Water and Sediments of a Freshwater
- 988 System (Antuã River, Portugal). *Sci. Total Environ.* **2018**, *633*, 1549–1559.
- 989 https://doi.org/10.1016/j.scitotenv.2018.03.233.
- 990 (118) Lin, L.; Zuo, L.-Z.; Peng, J.-P.; Cai, L.-Q.; Fok, L.; Yan, Y.; Li, H.-X.; Xu, X.-R. Occurrence and
- Distribution of Microplastics in an Urban River: A Case Study in the Pearl River along Guangzhou
  City, China. *Sci. Total Environ.* 2018, *644*, 375–381. https://doi.org/10.1016/j.scitotenv.2018.06.327.
- 993 (119) Su, L.; Xue, Y.; Li, L.; Yang, D.; Kolandhasamy, P.; Li, D.; Shi, H. Microplastics in Taihu Lake,
  994 China. *Environ. Pollut.* 2016, *216*, 711–719. https://doi.org/10.1016/j.envpol.2016.06.036.
- 995 (120) Ballent, A.; Corcoran, P. L.; Madden, O.; Helm, P. A.; Longstaffe, F. J. Sources and Sinks of
- 996 Microplastics in Canadian Lake Ontario Nearshore, Tributary and Beach Sediments. *Mar. Pollut.*997 *Bull.* 2016, *110* (1), 383–395. https://doi.org/10.1016/j.marpolbul.2016.06.037.
- 998 (121) McCormick, A.; Hoellein, T. J.; Mason, S. A.; Schluep, J.; Kelly, J. J. Microplastic Is an Abundant
  999 and Distinct Microbial Habitat in an Urban River. *Environ. Sci. Technol.* 2014, 48 (20), 11863–
- 1000 11871. https://doi.org/10.1021/es503610r.
- 1001 (122) Claessens, M.; Meester, S. D.; Landuyt, L. V; Clerck, K. D.; Janssen, C. R. Occurrence and
  1002 Distribution of Microplastics in Marine Sediments along the Belgian Coast. *Mar. Pollut. Bull.* 2011,
  1003 62 (10), 2199–2204.
- (123) Vianello, A.; Boldrin, A.; Guerriero, P.; Moschino, V.; Rella, R.; Sturaro, A.; Da Ros, L. Microplastic
  Particles in Sediments of Lagoon of Venice, Italy: First Observations on Occurrence, Spatial Patterns
  and Identification. *Estuar. Coast. Shelf Sci.* 2013, *130*, 54–61.
- 1007 https://doi.org/10.1016/j.ecss.2013.03.022.

- 1008 (124) Kim, S. W.; Chae, Y.; Kim, D.; An, Y.-J. Zebrafish Can Recognize Microplastics as Inedible
- 1009 Materials: Quantitative Evidence of Ingestion Behavior. *Sci. Total Environ.* **2019**, *649*, 156–162.
- 1010 https://doi.org/10.1016/j.scitotenv.2018.08.310.
- 1011 (125) Vroom, R. J. E.; Koelmans, A. A.; Besseling, E.; Halsband, C. Aging of Microplastics Promotes
- 1012 Their Ingestion by Marine Zooplankton. *Environ. Pollut.* 2017, 231, Part, 987–996.
- 1013 https://doi.org/https://doi.org/10.1016/j.envpol.2017.08.088.
- 1014 (126) Savoca, M. S.; Wohlfeil, M. E.; Ebeler, S. E.; Nevitt, G. A. Marine Plastic Debris Emits a Keystone
  1015 Infochemical for Olfactory Foraging Seabirds. *Sci. Adv.* 2016, *2* (11).
- 1016 (127) Savoca, M. S.; Tyson, C. W.; McGill, M.; Slager, C. J. Odours from Marine Plastic Debris Induce
- 1017 Food Search Behaviours in a Forage Fish. Proc. R. Soc. B Biol. Sci. 2017, 284 (1860), 20171000.
- 1018 https://doi.org/10.1098/rspb.2017.1000.
- (128) Cheung, P. K.; Cheung, L. T. O.; Fok, L. Seasonal Variation in the Abundance of Marine Plastic
  Debris in the Estuary of a Subtropical Macro-Scale Drainage Basin in South China. *Sci. Total Environ.* 2016, *562*, 658–665. https://doi.org/10.1016/j.scitotenv.2016.04.048.
- 1022 (129) Goldberg, E. D. Plasticizing the Seafloor: An Overview. *Environ. Technol.* 1997, *18* (2), 195–201.
   1023 https://doi.org/10.1080/09593331808616527.
- 1024 (130) Lusher, A. L.; McHugh, M.; Thompson, R. C. Occurrence of Microplastics in the Gastrointestinal
  1025 Tract of Pelagic and Demersal Fish from the English Channel. *Mar. Pollut. Bull.* 2013, 67 (1–2), 94–
  1026 99.
- 1027 (131) Chagnon, C.; Thiel, M.; Antunes, J.; Ferreira, J. L.; Sobral, P.; Ory, N. C. Plastic Ingestion and
- 1028 Trophic Transfer between Easter Island Flying Fish (Cheilopogon Rapanouiensis) and Yellowfin
- 1029 Tuna (Thunnus Albacares) from Rapa Nui (Easter Island). *Environ. Pollut.* 2018, 243, 127–133.
- 1030 https://doi.org/10.1016/j.envpol.2018.08.042.
- 1031 (132) Bravo Rebolledo, E. L.; Van Franeker, J. A.; Jansen, O. E.; Brasseur, S. M. J. M. Plastic Ingestion by
- 1032Harbour Seals (Phoca Vitulina) in The Netherlands. Mar. Pollut. Bull. 2013, 67 (1-2), 200-202.
- 1033 https://doi.org/10.1016/j.marpolbul.2012.11.035.
- 1034 (133) Donohue, M. J.; Masura, J.; Gelatt, T.; Ream, R.; Baker, J. D.; Faulhaber, K.; Lerner, D. T.
- 1035 Evaluating Exposure of Northern Fur Seals, Callorhinus Ursinus, to Microplastic Pollution through
- 1036 Fecal Analysis. *Mar. Pollut. Bull.* **2019**, *138*, 213–221.

- 1037 https://doi.org/10.1016/j.marpolbul.2018.11.036.
- 1038 (134) Clements, K. D.; Choat, J. H. Influence of Season, Ontogeny and Tide on the Diet of the Temperate
  1039 Marine Herbivorous Fish Odax Pullus (Odacidae). *Mar. Biol.* 1993, *117* (2), 213–220.
- 1040 https://doi.org/10.1007/BF00345665.
- 1041 (135) King, A. J. Ontogenetic Dietary Shifts of Fishes in an Australian Floodplain River. *Mar. Freshw. Res.*1042 2005, 56 (2), 215. https://doi.org/10.1071/MF04117.
- 1043 (136) Demain, D. K.; Gallego, A.; Jaworski, A.; Priede, I. G.; Jones, E. G. Diet and Feeding Niches of
- 1044 Juvenile Gadus Morhua, Melanogrammus Aeglefinus and Merlangius Merlangus during the
- 1045 Settlement Transition in the Northern North Sea. J. Fish Biol. 2011, 79 (1), 89–111.
- 1046 https://doi.org/10.1111/j.1095-8649.2011.02997.x.
- 1047 (137) Nunn, A. D.; Tewson, L. H.; Cowx, I. G. The Foraging Ecology of Larval and Juvenile Fishes. *Rev.* 1048 *Fish Biol. Fish.* 2012, *22* (2), 377–408. https://doi.org/10.1007/s11160-011-9240-8.
- (138) Anastasopoulou, A.; Mytilineou, C.; Smith, C. J.; Papadopoulou, K. N. Plastic Debris Ingested by
   Deep-Water Fish of the Ionian Sea (Eastern Mediterranean). *Deep. Res. Part I Oceanogr. Res. Pap.* 2013, 74, 11–13.
- 1052 (139) Lazzaro, X. A Review of Planktivorous Fishes: Their Evolution, Feeding Behaviours, Selectivities,
  1053 and Impacts. *Hydrobiologia* 1987, *146* (2), 97–167. https://doi.org/10.1007/BF00008764.
- (140) Ory, N. C.; Gallardo, C.; Lenz, M.; Thiel, M. Capture, Swallowing, and Egestion of Microplastics by
   a Planktivorous Juvenile Fish. *Environ. Pollut.* 2018, *240*, 566–573.
- 1056 https://doi.org/10.1016/J.ENVPOL.2018.04.093.
- 1057 (141) Carson, H. S. The Incidence of Plastic Ingestion by Fishes: From the Prey's Perspective. *Mar. Pollut.*1058 *Bull.* 2013.
- 1059 (142) Lima, A. R. A.; Costa, M. F.; Barletta, M. Distribution Patterns of Microplastics within the Plankton
  1060 of a Tropical Estuary. *Environ. Res.* 2014, *132*, 146–155.
- 1061 https://doi.org/10.1016/j.envres.2014.03.031.
- 1062 (143) van der Oost, R.; Beyer, J.; Vermeulen, N. P. . Fish Bioaccumulation and Biomarkers in
- 1063 Environmental Risk Assessment: A Review. *Environ. Toxicol. Pharmacol.* 2003, *13* (2), 57–149.
- 1064 https://doi.org/10.1016/S1382-6689(02)00126-6.
- 1065 (144) Browne, M. A.; Crump, P.; Niven, S. J.; Teuten, E.; Tonkin, A.; Galloway, T.; Thompson, R.

- 1066 Accumulation of Microplastic on Shorelines Worldwide: Sources and Sinks. *Environ. Sci. Technol.*
- **2011**, *45* (21), 9175–9179. https://doi.org/10.1021/es201811s.
- 1068 (145) Salvador Cesa, F.; Turra, A.; Baruque-Ramos, J. Synthetic Fibers as Microplastics in the Marine
- 1069 Environment: A Review from Textile Perspective with a Focus on Domestic Washings. Sci. Total

1070 *Environ.* 2017, *598*, 1116–1129. https://doi.org/10.1016/j.scitotenv.2017.04.172.

- 1071 (146) Kase, R.; Korkaric, M.; Werner, I.; Ågerstrand, M. Criteria for Reporting and Evaluating Ecotoxicity
- 1072 Data (CRED): Comparison and Perception of the Klimisch and CRED Methods for Evaluating
- 1073 Reliability and Relevance of Ecotoxicity Studies. *Environ. Sci. Eur.* 2016, 28 (1), 7.
- 1074 https://doi.org/10.1186/s12302-016-0073-x.
- 1075 (147) Collard, F.; Gilbert, B.; Compère, P.; Eppe, G.; Das, K.; Jauniaux, T.; Parmentier, E. Microplastics in
- Livers of European Anchovies (Engraulis Encrasicolus, L.). *Environ. Pollut.* 2017, 229, 1000–1005.
  https://doi.org/10.1016/j.envpol.2017.07.089.
- 1078 (148) Ribeiro, F.; Garcia, A. R.; Pereira, B. P.; Fonseca, M.; Mestre, N. C.; Fonseca, T. G.; Ilharco, L. M.;
  1079 Bebianno, M. J. Microplastics Effects in Scrobicularia Plana. *Mar. Pollut. Bull.* 2017, *122* (1–2),
- 1080 379–391. https://doi.org/10.1016/j.marpolbul.2017.06.078.
- (149) Khan, M. B.; Prezant, R. S. Microplastic Abundances in a Mussel Bed and Ingestion by the Ribbed
   Marsh Mussel Geukensia Demissa. *Mar. Pollut. Bull.* 2018, *130*, 67–75.
- 1083 https://doi.org/10.1016/j.marpolbul.2018.03.012.
- 1084 (150) Rochman, C. M.; Browne, M. A.; Underwood, A. J.; van Franeker, J. A.; Thompson, R. C.; Amaral-
- Zettler, L. A. The Ecological Impacts of Marine Debris: Unraveling the Demonstrated Evidence from
   What Is Perceived. *Ecology* 2016, *97* (2), 302–312. https://doi.org/10.1890/14-2070.1.
- 1087 (151) Mazurais, D.; Ernande, B.; Quazuguel, P.; Severe, A.; Huelvan, C.; Madec, L.; Mouchel, O.;
- 1088 Soudant, P.; Robbens, J.; Huvet, A.; et al. Evaluation of the Impact of Polyethylene Microbeads
- 1089 Ingestion in European Sea Bass (Dicentrarchus Labrax) Larvae. Mar. Environ. Res. 2015, 112, 78–
- 1090 85. https://doi.org/10.1016/j.marenvres.2015.09.009.
- 1091 (152) de Sá, L. C.; Oliveira, M.; Ribeiro, F.; Rocha, T. L.; Futter, M. N. Studies of the Effects of
- 1092 Microplastics on Aquatic Organisms: What Do We Know and Where Should We Focus Our Efforts
- in the Future? *Sci. Total Environ.* **2018**, *645*, 1029–1039.
- 1094 https://doi.org/10.1016/j.scitotenv.2018.07.207.

- 1095 (153) Crain, D. A.; Eriksen, M.; Iguchi, T.; Jobling, S.; Laufer, H.; LeBlanc, G. A.; Guillette, L. J. An
- 1096 Ecological Assessment of Bisphenol-A: Evidence from Comparative Biology. *Reprod. Toxicol.* 2007,
- 1097 24 (2), 225–239. https://doi.org/10.1016/j.reprotox.2007.05.008.
- 1098 (154) Oehlmann, J.; Schulte-Oehlmann, U.; Kloas, W.; Jagnytsch, O.; Lutz, I.; Kusk, K. O.; Wollenberger,
- 1099 L.; Santos, E. M.; Paull, G. C.; Van Look, K. J. W.; et al. A Critical Analysis of the Biological
- 1100 Impacts of Plasticizers on Wildlife. *Philos. Trans. R. Soc. Lond. B. Biol. Sci.* 2009, 364 (1526), 2047–
- 1101 2062. https://doi.org/10.1098/rstb.2008.0242.
- 1102 (155) Johnson, L. L.; Anulacion, B. F.; Arkoosh, M. R.; Burrows, D. G.; da Silva, D. A. M.; Dietrich, J. P.;
- 1103 Myers, M. S.; Spromberg, J.; Ylitalo, G. M. Effects of Legacy Persistent Organic Pollutants (POPs)
- 1104 in Fish—Current and Future Challenges. In Organic Chemical Toxicology of fishes; Tierney, K. B.,
- 1105 Farrell, A. P., Brauner, C. J., Eds.; Elsevier, 2013; pp 53–140. https://doi.org/10.1016/B978-0-12-
- **1106 398254-4.00002-9**.

1107