



HAL
open science

Study of a large scale powdered activated carbon pilot: Removals of a wide range of emerging and priority micropollutants from wastewater treatment plant effluents

R. Mailler, J. Gasperi, Y. Coquet, S. Deshayes, S. Zedek, C. Cren-Olivé, N. Cartiser, V. Eudes, Adèle Bressy, E. Caupos, et al.

► To cite this version:

R. Mailler, J. Gasperi, Y. Coquet, S. Deshayes, S. Zedek, et al.. Study of a large scale powdered activated carbon pilot: Removals of a wide range of emerging and priority micropollutants from wastewater treatment plant effluents. *Water Research*, 2015, 72, pp.315-330. 10.1016/j.watres.2014.10.047 . hal-01103115

HAL Id: hal-01103115

<https://enpc.hal.science/hal-01103115v1>

Submitted on 19 Jun 2018

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

1 Study of a large scale powdered activated carbon pilot: removals of a wide range of
2 emerging and priority micropollutants from wastewater treatment plant effluents

3
4 R. Mailler^{1*}, J. Gasperi^{1*}, Y. Coquet², S. Deshayes^{1,3}, S. Zedek¹, C. Cren-Olivé⁴, N. Cartiser⁴, V. Eudes³, A.
5 Bressy¹, E. Caupos¹, R. Moilleron¹, G. Chebbo⁵ and V. Rocher⁶

6
7 ¹ LEESU (UMR MA 102, Université Paris-Est, AgroParisTech), 61 avenue du Général de Gaulle, 94010
8 Créteil Cedex, France. (E-mail: *maillerr@leesu.enpc.fr*; *gasperi@u-pec.fr*)

9 ² SAUR, Direction de la Recherche et du Développement, 1 rue Antoine Lavoisier, 78064 Guyancourt,
10 France.

11 ³ LCPP (Laboratoire Central de la Préfecture de Police), 39 bis rue de Dantzig, 75015 Paris, France.

12 ⁴ ISA (UMR 5280, Institut des sciences analytiques, CNRS), 5 rue de la Doua, 69100 Villeurbanne, France.

13 ⁵ LEESU (UMR MA 102, Université Paris-Est, AgroParisTech), 6-8 avenue Blaise Pascal, Champs-sur-
14 Marne, 77455 Marne-la-Vallée Cedex 2, France.

15 ⁶ SIAAP, Direction du Développement et de la Prospective, 82 avenue Kléber, 92700 Colombes, France.

16 * Corresponding authors

17
18 **HIGHLIGHTS**

- 19 • Pharmaceuticals and hormones (PhPHs) are well removed (> 60%) by powdered activated carbon (PAC)
- 20 • 50-95% removals are also achieved for parabens, bisphenol A and pesticides by PAC
- 21 • Treatment degradation leads to substantially lower PhPHs removals by PAC
- 22 • The fresh PAC dose is strongly correlated to the process efficacy
- 23 • Molecular charge, hydrophobicity and molecular weight have a great influence on the micropollutant fate
- 24
- 25

26 ABSTRACT

27 The efficacy of a fluidized powdered activated carbon (PAC) pilot (CarboPlus®) was studied in both nominal (total
28 nitrification + post denitrification) and degraded (partial nitrification + no denitrification) configuration of the Seine
29 Centre WWTP (Colombes, France). In addition to conventional wastewater parameters 54 pharmaceuticals and
30 hormones (PhPHs) and 59 other emerging pollutants were monitored in influents and effluents of the pilot. Thus, the
31 impacts of the WWTP configuration, the process operation and the physico-chemical properties of the studied
32 compounds were assessed in this article.

33

34 Among the 26 PhPHs quantified in nominal WWTP configuration influents, 8 have high dissolved concentrations (>100
35 ng/L), 11 have an intermediary concentration (10-100 ng/L) and 7 are quantified below 10 ng/L. Sulfamethoxazole is
36 predominant (about 30% of the sum of the PhPHs). Overall, 6 PhPHs are poorly to moderately removed (<60%), such
37 as ibuprofen, paracetamol or estrone, while 9 are very well removed (>80%), i.e. beta blockers, carbamazepine or
38 trimethoprim, and 11 are well eliminated (60-80%), i.e. diclofenac, naproxen or sulfamethoxazole.

39 In degraded WWTP configuration, higher levels of organic matter and higher concentrations of most pollutants are
40 observed. Consequently, most PhPHs are substantially less removed in percentages but the removed flux is higher.
41 Thus, the PAC dose required to achieve a given removal percentage is higher in degraded WWTP configuration. For the
42 other micropollutants (34 quantified), artificial sweeteners and phthalates are found at particularly high concentrations
43 in degraded WWTP configuration influents, up to µg/L range. Only pesticides, bisphenol A and parabens are largely
44 eliminated (50-95%), while perfluorinated acids, PAHs, triclosan and sweeteners are not or weakly removed (<50%).
45 The remaining compounds exhibit a very variable fate from campaign to campaign.

46

47 The fresh PAC dose was identified as the most influencing operation parameter and is strongly correlated to
48 performances. Charge and hydrophobicity of compounds have been recognized as crucial for the micropollutant
49 adsorption on PAC, as well as the molecular weight. Finally, a PAC dose of 10 mg/L allows an average removal of 72-
50 80% of the sum of the PhPHs in nominal WWTP configuration. The comparison of the results with those from the
51 scarce other studies tends to indicate that an extrapolation of them to different PAC processes and to other WWTPs
52 could be possible and relevant, taking into account the differences of water quality from WWTP to WWTP.

53

54 KEYWORDS

55 Pharmaceuticals; emerging pollutants; adsorption; powdered activated carbon; wastewater treatment

56

57 INTRODUCTION

58

59 In addition to priority pollutants (EC 2013), many other emerging micropollutants such as pharmaceuticals and
60 hormones (PhPHs), personal care products (PCPs), pesticides, phthalates, artificial sweeteners, etc. are found in the
61 aquatic environment (Jones et al. 2001, Lange et al. 2012, Luo et al. 2014). As bioactive and toxic substances, their
62 environmental effects have been proven (Bolong et al. 2009, Daughton and Ternes 1999). For some of these compounds
63 such as PhPHs and PCPs, wastewater treatment plant (WWTP) discharges have been identified as an important source
64 in the aquatic environment (Halling-Sørensen et al. 1998, Heberer 2002).

65

66 Several studies have shown that conventional WWTPs (with primary and biological treatments) substantially remove
67 hydrophobic, volatile and biodegradable micropollutants even if they are not designed and/or optimized for that purpose
68 (Clara et al. 2007, Mailler et al. 2013, Ruel et al. 2012). However, hydrophilic or refractory organic compounds are still
69 present in the treated wastewater at ng/L to µg/L level (Loos et al. 2013). Besides, emerging micropollutants,
70 particularly pharmaceuticals such as diclofenac, may be included in future modifications of existing European (EC
71 2013) or national regulations. Among the solutions to reduce the emerging contaminant discharges into the
72 environment, the implementation of a tertiary treatment in conventional WWTPs is more and more considered by
73 WWTP managers and decision makers. These technologies, such as membrane filtration, advanced oxidation and
74 adsorption processes, are principally adapted from drinking water industry. They have been employed for a long time,
75 but the generalization of their applications to wastewater treatment has been seriously considered recently. However,
76 very few studies have shown the potential of such technologies to remove organic pollutants from treated wastewater
77 (Boehler et al. 2012, Margot et al. 2013) and their economic feasibility (Abegglen and Siegrist 2012).

78

79 In this context, a large scale powdered activated carbon (PAC) pilot has been set up at the Seine Centre WWTP, treating
80 wastewater from Paris conurbation and supervised by the Parisian public sanitation service (SIAAP). An extended
81 sampling strategy, both in number of campaigns (n=14) and compounds (N=113), has been built.

82

83 Thus, this paper aims at displaying and comparing the performances of such a process for micropollutant removal in
84 both nominal and degraded WWTP configurations, and to assess the influence of both operation parameters and
85 physico-chemical properties of the pollutants. First, the influent dissolved concentrations for both micropollutants and

86 conventional wastewater parameters are discussed and compared to the literature. Then, the removals of the molecules
87 are examined, as well as the effluent dissolved concentrations and the water quality. The difference between a normal
88 and a degraded configuration of the WWTP is particularly analyzed. Finally, the influence of operation parameters
89 (fresh PAC dose, total mass of carbon in the reactor, hydraulic velocity) and physico-chemical properties of the
90 compounds (charge, hydrophobicity, size, functional groups, etc.) are also examined. Such results are of high interest
91 and very relevant considering the poverty of the literature regarding the removal of micropollutants from wastewater by
92 PAC.

93

94

95 MATERIALS AND METHODS

96

97 1. Site and process description

98

99 In nominal configuration, the Seine Centre WWTP treats 240 000 m³ per day of domestic wastewater from Paris city,
100 corresponding to 900 000 population equivalents. In these conditions (total nitrification + post denitrification), the
101 treatment process is composed by three major steps: pre, primary and biological treatments. After the pretreatment
102 (screening, grit and oil removal units), a physicochemical lamellar settling unit (Densadeg[®]) removes a great part of the
103 particulate and colloidal pollution thanks to coagulant (ferric chloride) and flocculant (anionic polymer) addition.
104 Finally, a three stage biofiltration system achieves the biological treatment. In nominal configuration, the first stage
105 (Biofor[®] filters - biolite medium), allows the treatment of carbonaceous pollution in aerobic conditions, and the second
106 (Biostyr[®] filters - biostyrene medium) and third (Biofor[®] filters) stages remove the nitrogenous pollution respectively in
107 aerobic (total nitrification) and anoxic (denitrification) conditions (Rocher et al. 2012). This WWTP is able to switch to
108 a degraded configuration to treat 405 000 m³ of water per day but a partial nitrification and no denitrification are
109 performed. A detailed layout of the WWTP in both configurations is given in supporting material - Figure S1.

110

111 The tertiary treatment process studied (CarboPlus[®] - Figure 1) is fed by Seine Centre treated effluents and treats
112 between 700 and 1 200 m³/day. This pilot consists in a 5 m high reactor with a surface area of 4 m², where water flows
113 upstream through a fluidized bed of PAC. Depending on the mass of PAC inside the reactor and the hydraulic velocity,
114 the bed depth varies between 1 and 3 m. A dose of fresh PAC is continuously injected. At steady state, a comparable
115 amount of PAC is regularly removed from the reactor to obtain a solid retention time (SRT) of 5 to 7 days, and an

116 overall concentration of 5 to 10 g/L of PAC is maintained inside the bed. In addition, coagulant (ferric chloride) and
117 flocculant (anionic polymer) are injected continuously with the fresh PAC to stabilize the bed expansion and to prevent
118 the leakage of PAC with treated water. The hydraulic velocity can technically vary from 6 to 12 m/h and the hydraulic
119 retention time considering the reactor size is comprised between 25 and 50 min. The observed contact time between the
120 PAC bed and the water varies between 10 and 20 min considering the bed depth and the hydraulic velocity.

121 FIGURE 1

122 2. Study and sampling strategy

123
124
125 The study has been initially built to assess the process efficacy in different operation configuration (Table 1) during 14
126 campaigns (from July to December 2013). However, the Seine Centre WWTP operated in degraded configuration
127 during the 5 lasts campaigns (unplanned), because it received additional water from another WWTP that was in
128 maintenance. Taking this into account, the campaigns performed have been organized in two phases in the results and
129 discussion sections of the article. The first 9 campaigns, corresponding to phase 1, were performed during nominal
130 WWTP configuration. The last 5 campaigns were carried out under degraded WWTP configuration, corresponding to
131 phase 2.

132 In addition, different process configurations were tested during the first phase (Table 1) by varying the fresh PAC dose
133 in the ranges operationally feasible within the process. Thus, three moderate PAC doses (5-10-20 mg/L) and one high
134 PAC dose (70 mg/L), as well as four hydraulic velocities (6-7-8-10 m/h) were tested, allowing determining their
135 influences on performances. The 70 mg/L of PAC campaign was performed to determine whether some compounds
136 could still be recalcitrant to adsorption at high dose. Information about operation parameters during this phase are given
137 in Table 1. In contrary, a stable and intermediate configuration (10 mg/L of PAC and 7 m/h), considered as the
138 recommended configuration of the process, was maintained during the second phase.

139 Thus, this experimental design allows assessing the influences of both the process operation parameters and the
140 configuration of the WWTP.

141 TABLE 1

142
143 The same sampling procedure was applied for the 14 campaigns. Analyzes were performed on the dissolved phase. 24-h
144 composite samples of 20 L were collected using automatic refrigerated samplers (4°C) equipped with glass bottles,
145 properly cleaned, and Teflon[®] pipes to avoid any contamination or loss. When sampling was completed, glass bottles
146 were collected, very carefully homogenized and sub-samples were distributed for analyses. Each sample was filtered

147 using 0.7 μm GF/F filters (Whatman®). All samples were analyzed within 48-h after sampling, due to transport time to
148 the laboratories involved, except for sweeteners. As recommended by the laboratory performing the artificial
149 sweeteners analyses (internal tests), these samples were acidified (1% volumetric HCl) and stored in fridge (4°C) until
150 analyses.

151

152 3. Activated carbon characterization

153

154 Activated carbon characteristics are very important in the adsorption mechanisms and can strongly influence the
155 micropollutant fate (Çeçen and Aktas 2012). Apart from BET (Brunauer, Emmett and Teller) specific surface and
156 particle size, the mesoporous structure is the best suited for micropollutants adsorption (Çeçen and Aktas 2012, Ebie et
157 al. 2001), as it reduces the impact of organic matter competition with micropollutants. Thus, a mesoporous PAC
158 (DaCarb PB-170) was tested in the pilot, after preliminary lab tests comparing different PACs. The characteristics of
159 this PAC are given in supporting material - Table 1. Globally, the selected PAC is featured by a high BET surface of
160 $957 \pm 28 \text{ g/m}^2$, which is close to BET specific surfaces found in the literature, often comprised between 700 and 1 500
161 m^2/g (Çeçen and Aktas 2012, de Ridder et al. 2013, Margot et al. 2013). In addition, its granulometry is large but its
162 median particle diameter is rather low ($<20 \mu\text{m}$), which theoretically enhance the adsorption kinetics. PAC observations
163 by scanning electron microscopy are provided in supporting material - Figure S2.

164

165 4. Pollutants and analytical procedures

166

167 In all samples, conventional wastewater parameters were analyzed by SIAAP French accredited laboratory (COFRAC,
168 supporting material - Table S2) to characterize the general quality of water. These parameters include: dissolved organic
169 carbon (DOC), chemical oxygen demand (COD), biological oxygen demand (BOD₅), UV absorption at 254 nm (UV-
170 254), total Kjeldahl nitrogen (TKN), NH_4^+ , NO_3^- , NO_2^- , PO_4^{3-} , total phosphorous (TP) and total suspended solids (TSS).
171 All analyses were performed according to standards (supporting material - Table S2).

172

173 113 micropollutants were monitored in the dissolved phase (Table 2). First, pharmaceuticals (N=44, N for the number
174 of compounds) and hormones (N=10) were analyzed for all campaigns, as they are the compounds targeted by the
175 process. In addition to these molecules, complementary analyses were performed for other micropollutants (N=59)
176 during phase 2. Thus, chlorinated solvents (N=3), perfluorinated acids (N=2), pesticides (N=23), bisphenol A,

177 alkylphenols (N=2) and phthalates (N=4) were analyzed in all campaigns of phase 2, while PAHs (N=13 - the 16 US-
178 EPA PAHs excluding naphthalene, acenaphthene and acenaphthylene) and PCPs (N=7) were monitored in 3 campaigns
179 and artificial sweeteners (N=4) in 4 campaigns. The full list of molecules is given in supporting material - Table S3, as
180 well as their respective limits of quantification (LQ).

181
182 Micropollutant analyses were performed by 4 laboratories: the Institute of Analytical Sciences (ISA - Villeurbanne,
183 France) from CNRS, in charge of the PhPHs, chlorinated solvents, perfluorinated acids, pesticides and bisphenol A, the
184 Water Environment and Urban Systems laboratory (LEESU - Créteil, France), in charge of PAHs and PCPs, the Central
185 laboratory of the Police Prefecture (LCPP - Paris, France), in charge of phthalates and alkylphenols, and the Water
186 Technology Center (TZW - Karlsruhe, Germany), in charge of artificial sweeteners.

187 TABLE 2

188 Information about the analytical procedures (Table 2) are available in the literature for PhPHs (Vulliet et al. 2011),
189 pesticides and chlorinated solvents (Barrek et al. 2009), PAHs (Bressy et al. 2012), alkylphenols and phthalates (Bergé
190 et al. 2014), sweeteners (Scheurer et al. 2009), and for PCPs (Gasperi et al. 2014).

191

192 5. Data processing

193

194 The statistical calculations were performed separately for the nominal (n=9) and the degraded (n=5) WWTP
195 configurations, for comparison purposes. When the number of values available was higher than six, box plots were
196 plotted with minimum, 1st quartile (Q1), 3rd quartile (Q3) and maximum values of the series. In contrary (≤ 6 values),
197 individual values were plotted. Finally, the results given in text, concentrations as well as removals, are average results
198 \pm standard deviations (>6 values) or minimum – maximum values (≤ 6 values).

199

200 As no direct assessment of the uncertainties was performed, and to have an idea of the robustness of the data, limit
201 values of 5 times the LQ were defined for each compound. The uncertainty was assumed moderate for concentrations
202 measured above these limit values ($<30\%$), while the uncertainty was considered high for concentrations below them
203 ($>30\%$). Such an approach has already been adopted by (Ruel et al. 2011), which stated that the uncertainty on the
204 micropollutant concentration in wastewater is generally comprised between 30 and 100% when the value measured is
205 lower than 2.5 to 10 times the LQ, depending on the compound, and lower than 30% when higher than this value.
206 Considering that, the removal of micropollutants was estimated only in several situations to limit the propagation of

207 uncertainties and provide more reliable results. When the compound was quantified above 5 times the LQ in influents
208 and above the LQ in effluents, the removal was conventionally calculated. In addition, when the compound was
209 quantified above the LQ in influents ($>LQ$ or $>5 \times LQ$) and below the LQ in effluents, the removal was estimated using
210 $LQ/2$ instead of the effluent concentration. The removals were not calculated when concentrations of both influents and
211 effluents were measured between the LQ and 5 times the LQ.

212

213 Finally, 4 classes of behaviors were defined regarding the obtained removals: very good ($>80\%$), good (60-80%),
214 moderate (30-60%) and poor ($<30\%$).

215

216

217 RESULTS AND DISCUSSION

218

219 1. Quality of influents in nominal and degraded WWTP configuration

220

221 *1.1. Conventional wastewater parameters in influents*

222

223 As displayed in Table 3, the pilot influents are characterized by relatively low values of the conventional wastewater
224 parameters in nominal WWTP configuration, particularly DOC (5.6 ± 0.9 mgC/L), UV-254 (0.110 ± 0.013 cm⁻¹), COD
225 and BOD₅ (26 ± 11 and 4.8 ± 3.5 mgO₂/L), TKN (1.5 ± 0.2 mgN/L), NH₄⁺ (< 0.3 mgN/L), TSS (3 ± 1 mg/L) and TP
226 (< 0.3 mgP/L). Indeed, this WWTP achieves very high removals of TSS (98%), COD (92%), TN (76%) and TP (95%)
227 in nominal configuration (Mailler et al. 2013). In addition, a substantial fraction of the DOC ($\approx 20\%$) is due to residual
228 methanol from the post denitrification step (SIAAP source), in contrary to DOC from the degraded WWTP
229 configuration (no denitrification).

230 In this study, the WWTP effluents are rather less concentrated (DOC, UV-254, etc.) than effluents from other studies
231 focusing on PAC adsorption from wastewater. Besides, effluents with average DOC values of 9.6 - 14.4 mgC/L and
232 UV-254 of 0.239 - 0.397 cm⁻¹ were recently studied by (Altmann et al. 2014), while DOC concentration of the water
233 studied by (Löwenberg et al. 2014) was 8.8 ± 1.2 mgC/L. Similarly, (Margot et al. 2013) performed PAC adsorption on
234 biological treatment effluents with DOC of 7.3 ± 1.9 mgC/L. Only the study of (Boehler et al. 2012) provides
235 micropollutant removal by PAC data with lower DOC concentration water, between 5.6 and 8.9 mgC/L.

236

237 The degraded configuration, with partial nitrification and no denitrification, induces a notable increase of
238 concentrations, particularly DOC (7.5 ± 0.5 mgC/L) and the UV-254 (0.139 ± 0.011 cm⁻¹). In addition, as no
239 denitrification is performed in this WWTP configuration, the composition of the DOC is different without residue of
240 methanol. The difference of DOC concentration is then even higher between the two WWTP configurations.

241 The concentration of NH₄⁺ in influents highlights the degradation of the nitrification step, with concentrations up to 4.4
242 ± 0.7 mgN/L. Similarly, the highest concentration of NO₃⁻ in influents from degraded configuration (23.7 ± 2.1 mgN/L)
243 reflects the lack of denitrification. Finally, the TSS concentration is doubled when WWTP configuration is degraded
244 (Table 3). Overall, the obtained quality of influents in this WWTP configuration is degraded and quite comparable to
245 the one from (Margot et al. 2013).

246 The higher quantity of organic matter and most of pollutants in degraded WWTP configuration should be unfavorable
247 to micropollutant adsorption. Indeed, it has been demonstrated that a higher DOC concentration induces a stronger
248 competition effect (Delgado et al. 2012, Margot et al. 2013).

249

250

251 *1.2. Micropollutant concentrations in influents from nominal WWTP configuration*

252

253 Among the 54 PhPHs monitored, a total of 26 compounds were quantified at least during the nominal WWTP
254 configuration. During this WWTP configuration, 14 were systematically quantified in influents, 12 in several samples
255 but 28 were never quantified (supporting material - Table S4). The dissolved concentrations of the 26 compounds, in
256 both nominal and degraded configuration of the WWTP, are presented in Table 2, with their occurrences.

257

258 8 compounds were measured in influents at average dissolved concentrations higher than 100 ng/L: sulfamethoxazole
259 (993 ± 817 ng/L), ofloxacin (412 ± 315 ng/L), carbamazepine (215 ± 85 ng/L), atenolol (185 ± 51 ng/L), diclofenac
260 (184 ± 91 ng/L), oxazepam (139 ± 128 ng/L) and erythromycin (124 ± 32 ng/L). 7 compounds were measured at
261 concentrations lower than 10 ng/L, especially estrone (9 ± 3 ng/L) and ibuprofen (9 ± 5 ng/L). The 11 remaining
262 compounds have average concentrations between 10 and 100 ng/L, i.e. naproxen (33 ± 28 ng/L), trimethoprim (64 ± 79
263 ng/L) or propranolol (97 ± 27 ng/L).

264 These concentrations are rather in accordance with data available in the literature for WWTP effluents (Deblonde et al.
265 2011, Loos et al. 2013, Luo et al. 2014, Miège et al. 2009), although they are overall in the lower part of the range (i.e.
266 analgesics). This could be due to i) the dilution of the Parisian wastewater (combined sewer) by parasite waters (Gasperi

267 et al. 2008) and ii) intense biological treatment including a total nitrification step, known to enhance the biodegradation
268 of micropollutants (Clara et al. 2005, McAdam et al. 2010). Sulfamethoxazole contributes about 30% of the total PhPHs
269 concentration, and its concentration lies in the upper part of the range found in the literature. No data were found for
270 testosterone, sulfadimerazine, sulfameter, and very scarce information are available for sulfadiazine, sulfathiazole,
271 fenofibrate and lorazepam. A short review as regard concentrations found in WWTP effluents for these compounds (23
272 references) is given in supporting material - Table S5.

273 In addition, most compounds are measured above 5 times their LQ (Table 3), leading to a moderate uncertainty on the
274 concentration. Sulfadimerazine, sulfadiazine, sulfameter, sulfathiazole, fenofibrate and testosterone are rather measured
275 between LQ and 5 times LQ, corresponding to high uncertainties, and the corresponding removals were consequently
276 not calculated.

277

278

279 *1.3. Micropollutant concentrations in influents from degraded WWTP configuration*

280

281 In the degraded configuration, 21 PhPHs were quantified. A similar pattern is observed but most of the PhPHs are
282 measured at higher average dissolved concentrations than in nominal WWTP configuration (Table 2), particularly
283 paracetamol (x200 based on average concentrations), ibuprofen (x100), ketoprofen (x10) and bezafibrate (x10). This is
284 mainly explained by both seasonal variations (phase 1 June-October, phase 2 November-December) and the lack of
285 nitrification since this step has been identified as crucial for micropollutant biodegradation in biological treatments,
286 especially for easily biodegradable molecules by biological treatments (Joss et al. 2005, Margot et al. 2013, Radjenović
287 et al. 2009). The degree of nitrification is then correlated to some PhPHs biodegradation (Margot et al. 2013), and a
288 total nitrification, as in nominal WWTP configuration, is known to be more efficient than a partial nitrification (Joss et
289 al. 2008), as in degraded WWTP configuration. This results in lower concentrations in discharges from nominal WWTP
290 configuration. Finally, the total concentration of the PhPHs doubles from $2\,729 \pm 1\,057$ ng/L in nominal WWTP
291 configuration to $4\,956 \pm 3\,628$ ng/L in degraded WWTP configuration, mainly due to the increase of paracetamol and
292 ibuprofen concentrations. In contrary, four compounds are measured at notably lower concentrations: sulfamethoxazole,
293 carbamazepine, ofloxacin and ciprofloxacin, probably because of seasonal variability.

294

295 A total of 59 other emerging micropollutants were also monitored during phase 2 in order to improve and enlarge the
296 performance overview of the PAC treatment. Concentrations are presented in Table 3. Among them, monitored during

297 degraded WWTP configuration, 34 were measured above the LQ. Only 3 pesticides (atrazine, diuron and isoproturon)
298 out the 23 pesticides investigated were quantified due to both a low occurrence in treated wastewater and high LQ.
299 Most of the compounds were always found in influents, especially PFOS, PFOA, bisphenol A, nonylphenols (NP),
300 para-tert-octylphenol (t-OP), bis(2-ethylhexyl) phthalate (DEHP), artificial sweeteners, triclosan and parabens.

301

302 A wide variety of dissolved concentrations is observed, from a few ng/L to more than 1 000 ng/L. Phthalate levels are
303 rather high, ranging from 300 to 3 000 ng/L. Similarly, acesulfame and sucralose exhibit the highest concentrations by
304 far, respectively $8\,725 \pm 602$ and $7\,150 \pm 745$ ng/L, due to very poor removals by conventional WWTPs (Lange et al.
305 2012). Artificial sweeteners have been recently recognized as a new class of emerging environmental contaminants
306 (Lange et al. 2012), highly persistent, and their toxicity in the environment is still not well known. Bisphenol A, NP,
307 cyclamate and saccharin are found at lower concentrations (100-1 000 ng/L), as well as diuron, isoproturon, PFOS,
308 PFOA and parabens (10-100 ng/L). Then, all the PAHs present dissolved concentrations lower than 20 ng/L in the
309 effluents. Finally, as the concentrations presented for these micropollutants were acquired in degraded WWTP
310 configuration, it can be assumed that they should be lower or equal in nominal WWTP configuration. Levels found are
311 very similar to other studies for pesticides, perfluorinated acids, bisphenol A, triclosan and phthalates (Bergé et al.
312 2013, Deblonde et al. 2011, Loos et al. 2013, Luo et al. 2014, Zareitalabad et al. 2013). In contrary, substantially higher
313 levels of sweeteners (Berset and Ochsenbein 2012, Ordóñez et al. 2012) and PAHs (Fatone et al. 2011, Qiao et al. 2014,
314 Sánchez-Avila et al. 2011) were found in WWTP effluents in the literature. Similarly, levels of parabens found are
315 rather high compared to the literature (Chen et al. 2012, Yu et al. 2012b). In particular, concentrations are slightly
316 higher than those reported by (Geara-Matta 2012) for the same WWTP discharges, highlighting the impact of the
317 degraded configuration. More information about data available in the literature is given in supporting material - Table
318 S5.

319 Like PhPHs, most of these other emerging micropollutants feature concentrations higher than 5 times their LQ, except
320 three PAHs (anthracene, benzo[a]anthracene and dibenzo[ah]anthracene) and some campaigns for NP, benzylbutyl
321 phthalate (BBP), cyclamate and two parabens (benzyl and butyl paraben).

322

323

324 2. Efficacy of PAC in nominal and degraded WWTP configurations

325

326 2.1. Conventional wastewater parameter removal in nominal and degraded WWTP configurations

327

328 An overall improvement of the water quality is observed after the PAC treatment. In nominal WWTP configuration,
329 concentration removals of DOC ($35 \pm 24\%$), UV-254 ($32 \pm 14\%$), COD ($13 \pm 14\%$) and BOD₅ ($39 \pm 19\%$) are
330 observed (n=9, average PAC dose of 14 mg/L). These results are rather similar or lower than those of (Margot et al.
331 2013), despite the same average PAC dose. Actually, this difference could be due to the filtration unit after the PAC
332 contact tank in their study, which was recognized as the main cause of the general water quality improvements because
333 it was biologically active. Furthermore, they evaluated between 20 and 35% the DOC removal due to the PAC against
334 $35 \pm 24\%$ and $23 \pm 4\%$ in nominal and degraded configurations respectively. Similarly, (Boehler et al. 2012) observed
335 a DOC removal between 15 and 48%, consistent with previously cited results.

336

337 In degraded WWTP configuration, the influent quality (Table 3) is degraded (Figure 4). Indeed, DOC is rather better
338 removed both in percentage and concentration in nominal WWTP configuration, despite higher influent concentrations.
339 In contrary, UV-254, COD, BOD₅ and TKN are similarly removed in both WWTP configurations. In addition, the
340 presence of NH₄⁺ in degraded WWTP effluents (Table 3) allows highlighting that a nitrification process occurs in the
341 pilot, with a removal of about $20 \pm 7\%$. This confirms the findings of (Margot et al. 2013) even if authors quote a
342 higher NH₄⁺ removal ($85 \pm 20\%$) because of the biologically active filtration unit after the PAC contact tank. The
343 formation of NO₃⁻ and NO₂⁻ confirms the nitrification process.

344

345

346 *2.2. Micropollutant removal in nominal WWTP configuration*

347

348 The overall performances of the pilot in both nominal and degraded WWTP configuration are presented in Figure 2. For
349 each compound, the results from nominal configuration are given on the left hand and the results from degraded
350 configuration on the right hand. A short literature review about the micropollutant removals by PAC is also provided in
351 supporting material - Table S5.

352

353 In nominal WWTP configuration with a fresh PAC dose at 10 mg/L (n=3), paracetamol and ibuprofen are poorly
354 removed, and 4 compounds are moderately removed by the pilot, including estrone (32-61%, min-max). In contrary, 9
355 substances are very well removed, including propranolol (96-98%), atenolol (86-92%), trimethoprim (84-98%),
356 carbamazepine (86-97%), oxazepam (82-91%), bezafibrate (75-99%) and ciprofloxacin (76-91%). Finally, the 11

357 remaining compounds are well removed, i.e. ofloxacin (63-89%), naproxen (46-99%), sulfamethoxazole (53-72%) and
358 diclofenac (72-85%).

359

360 To the best of our knowledge, papers dealing with the removal of micropollutants from wastewater by PAC are very
361 scarce, especially at large scale. Among the 26 PhPHs quantified, only 16 have been already studied in a comparable
362 context - adsorption from treated wastewater by PAC. Despite rather large standard deviations in our study, removals
363 are similar to those of these studies (Altmann et al. 2014, Boehler et al. 2012, Löwenberg et al. 2014, Margot et al.
364 2013) for these 16 common PhPHs, as shown by supporting materials - Table S3 and Figure S3. Only ibuprofen and
365 estrone removals are notably lower in our study, probably because of their lower influent concentrations. In contrary,
366 ciprofloxacin is better removed than in the literature. As the average fresh PAC doses are comparable for these studies
367 (14-15 mg/L), this suggests that micropollutant adsorption on PAC is likely to be a very stable phenomenon and that
368 these results could be extrapolated to other PAC treatments. A generalization of these results to other WWTPs and PAC
369 processes could be then possible and relevant, even if efficacy in waters with slightly higher DOC concentrations
370 should be slightly lower. In addition, some papers have already observed comparable performances over PhPHs with
371 advanced oxidation processes such as ozonation (Altmann et al. 2014, Margot et al. 2013). For several compounds,
372 scarce data are available displaying their removal by grain activated carbon filter and/or PAC batch experiments with
373 surface water (paracetamol, roxithromycin, erythromycin and testosterone), while no reference was found for
374 sulfadimerazine, sulfameter, sulfadiazine, sulfathiazole, lorazepam and fenofibrate (supporting material - Table S5).
375 Overall, observed removals are similar to those from former studies (Altmann et al. 2014, Boehler et al. 2012,
376 Löwenberg et al. 2014, Margot et al. 2013, Ruel et al. 2012, Snyder et al. 2007, Westerhoff et al. 2005).

377

378 The removals vary due to variations of the fresh PAC doses (supporting material - Table S1) and of the influent
379 concentrations. Despite this, the variations of removal percentages are relatively moderate for most of the compounds,
380 around 5-15% between Q1 and Q3, except for paracetamol, sulfamethoxazole and bezafibrate, which have higher
381 variations of 20-30%. Moreover, the observed variations are lower for well and very well removed compounds (i.e. beta
382 blockers or trimethoprim), i.e. below 10%.

383

384

385 *2.3. Micropollutant removal in degraded WWTP configuration*

386

387 In degraded WWTP configuration, different impacts can be observed on PhPHs removal percentages, depending on the
388 compound (Figure 4). First, paracetamol and ibuprofen are better removed in percentage in degraded configuration,
389 most likely due to their higher influent concentrations. Several compounds are rather similarly removed in both
390 configurations, i.e. both beta blockers, trimethoprim, sulfadiazine and ofloxacin, despite variations of influent
391 concentrations. Then, lower removals in percentage are observed for a majority of compounds, during degraded WWTP
392 configuration (Figure 4), despite higher influent concentrations. Nevertheless, the removed pollutant load is higher in
393 degraded than in nominal WWTP configuration (ketoprofen, naproxen, diclofenac, trimethoprim, roxithromycin,
394 metronidazole, atenolol, propranolol, oxazepam, lorazepam, bezafibrate and estrone). In contrary, carbamazepine,
395 ciprofloxacin and sulfamethoxazole have lower removal in degraded configuration due to lower influent concentrations.
396 This decrease of the micropollutant removals can be explained by both the variations of influent concentrations and the
397 water quality change. As demonstrated by higher DOC and UV-254, the influents in degraded WWTP exhibit higher
398 organic matter levels, inducing a stronger competition with micropollutants for adsorption on the PAC by direct site
399 competition and pore blocking (de Ridder et al. 2011, Delgado et al. 2012). In particular, the negative relation between
400 DOC concentration and micropollutant removals has already been highlighted in the literature (Altmann et al. 2014,
401 Margot et al. 2013).

402 Finally, the sum of PhPHs is similarly removed in nominal and degraded WWTP configurations (72-80% vs 70-81%).
403 This is mainly explained by higher removals of the high concentrated pollutants such as paracetamol and ibuprofen and
404 slightly lower removals of the other compounds.

405
406 Regarding the other emerging micropollutants monitored (Figure 3), atrazine (51-64%, min-max), isoproturon (51-83%)
407 and to a greater extent diuron (82-96%) have a good affinity for PAC, confirming the suitability of the process for
408 pesticide removal. This confirms other studies observations (Margot et al. 2013, Ruel et al. 2012, Snyder et al. 2007),
409 even if atrazine is slightly less removed in our case (lower PAC dose). In addition, parabens (>70%), particularly ethyl
410 paraben (88-94%) and propyl paraben (91-96%), are also well or very well removed. As far as authors know, this is the
411 first study displaying their fates within PAC in wastewater, as well as for phthalates and PAHs. In contrary, acesulfame
412 (9-19%), sucralose (6-26%), triclosan (18-29%) and PFOS (6-52%) are poorly removed by adsorption. For PFOA, no
413 clear conclusion can be drawn, since a systematic negative removal was observed, with concentrations always higher in
414 effluents than in influents. No sampling blanks were performed for the study, but this potential contamination may
415 result from Teflon® pipes or elements of the automatic samplers.

416

417 The 9 remaining compounds have variable removals but are rather moderately removed, such as saccharin (33-54%),
418 DEHP (49-63%), bisphenol A (49-78%) or NP (3-97%). The removal of these substances from wastewater by
419 adsorption was poorly studied, but (Ruel et al. 2012) reported that within a granulated activated carbon treatment,
420 DEHP and NP are respectively moderately (30-70%) and poorly (<30%) removed. (Yu et al. 2008) also observed poor
421 removal for NP by activated carbon. Similarly, (Scheurer et al. 2010) observed that moderate to good removals can be
422 achieved by PAC adsorption for saccharin and sucralose, what is consistent with results obtained in our study. Finally,
423 bisphenol A was found to be slightly better removed from wastewater by PAC in (Margot et al. 2013). Regarding
424 PAHs, the variation of removal is high but these compounds are overall not eliminated by the process or poorly
425 removed (10-40% for pyrene, fluoranthene, indeno[123]pyrene and benzo[ghi]perylene).

426

427 Even if the number of samples for these substances is limited, the variations of pesticide, sweetener and paraben
428 removals from campaign to campaign are lower than 30%. Contrariwise, phthalates (except DEHP), PAHs,
429 alkylphenols and perfluorinated acids have large variations of removals.

430

431

432 *2.4. Water quality after the PAC treatment*

433

434 The tertiary PAC treatment doesn't radically change the micropollutant pattern between influent and effluent (Table 3),
435 but it substantially decreases their dissolved concentrations. As the generalization of tertiary treatments to all WWTPs
436 would depend strongly on regulations, having an idea of the levels that could be required for these compounds is
437 crucial. Instead of an overall removal (average removal of 80%), such as required in Switzerland, environmental quality
438 criteria (EQC) for freshwaters, comparable to environmental quality standards (EQS) for priority pollutants from (EC
439 2013), could be used as references. Such EQC have been proposed in Switzerland by the Ecotoxicity Centre of
440 EAWAG (supporting material - Table S6) (Götz et al. 2010, Kase et al. 2011). A comparison of the concentrations
441 measured in effluents from the PAC treatment with these EQC or EQS enables to state if this process efficacy is
442 sufficient to reach references. However, these EQC/EQS apply for freshwaters and not effluents from WWTP, so a
443 dilution factor should be considered when evaluating the quality of the treatment and the compliance with regulations.

444

445 For PhPHs, naproxen, trimethoprim and atenolol concentrations are already measured below these proposed values
446 before treatment. Ibuprofen, sulfamethoxazole, ciprofloxacin, propranolol, carbamazepine and bezafibrate are always

447 measured below levels proposed after the treatment, while a fresh PAC dose of 20 mg/L allows decreasing diclofenac,
448 erythromycin and estrone concentrations below the EQC proposed in Switzerland. For diclofenac, this is particularly
449 interesting as this compound is now listed on the first watch list of the (EC 2013).
450 Among the other emerging micropollutants measured, some are listed as priority pollutants (EC 2013). Maximum
451 admissible concentration (MAC-EQS) and annual average environmental quality standards (AA-EQS) have been set up
452 for them in surface waters (supporting material - Table S5). All the compounds with MAC-EQS defined were always
453 quantified below these values in the dissolved phase. For AA-EQS, different cases were encountered: pesticides, t-OP
454 and DEHP were always found below, NP was measured below only after treatment; PFOS and benzo[*a*]pyrene were
455 still found above after treatment. Bisphenol A and triclosan are not included in (EC 2013), but have EQC proposed by
456 the Swiss Ecotoxicity Centre of EAWAG. Regarding it, bisphenol A was always measured below this criterion, even in
457 influents and before dilution, while triclosan was still measured above its limit value after treatment. However, for these
458 compounds, the campaigns have been performed when the WWTP operated in degraded configuration, what should
459 lead to higher concentrations (Geara-Matta 2012). In addition, the WWTP discharges are diluted in the Seine River.

460

461

462 3. Influence of operation parameters and physico-chemical properties of the compounds

463

464 Different types of parameters can influence the adsorption processes (de Ridder et al. 2011, Delgado et al. 2012):
465 structures and properties of the adsorbent, physico-chemical properties of the targeted compounds, water quality and
466 composition and the operation parameters (dose of adsorbent, contact time, etc.). As explained previously, emerging
467 micropollutant removals are lower when the WWTP operates in degraded configuration, because of stronger
468 competition from effluent organic matter and higher concentrations. Results of this study also allow characterizing the
469 relationship between fresh PAC dose and performances, as well as investigating the relations between physico-chemical
470 properties of the compounds and their behaviors with PAC.

471

472 3.1. Operation parameters

473

474 *The fresh PAC dose*

475 To assess the influence of the fresh PAC dose, 4 doses (5-10-20-70 mg/L) were tested during nominal WWTP
476 configuration and the results of the 12 PhPHs quantified in every campaigns of the phase 1 (Figure 4) were considered.

477 Based on our results, the fresh PAC dose appears to be the leading operation parameter as regards its influence on
478 performances. In particular, a significant positive correlation is found between the removal of the 26 PhPHs and the
479 dose of PAC ($r_{\text{spearman}} = 0.962$; p-value < 0.001; $\alpha = 0.05$) considering the results in nominal WWTP configuration.
480 Similarly, significant correlations are found between their individual removals and the PAC dose, as well as for DOC
481 removal (supporting material - Table S6). In particular, diclofenac, sulfamethoxazole, metronidazole, sulfadiazine,
482 atenolol, propranolol and carbamazepine have all spearman coefficient of correlation higher than 0.8 with p-value lower
483 than 0.05.

484 As previously reported (Boehler et al. 2012, Snyder et al. 2007), the higher the PAC dose, the higher the removals.
485 Overall, the point of inflexion is reached around 10 mg/L for most of the compounds, explaining why the gain of
486 removal is relatively moderate between 10 and 20 mg/L and between 20 and 70 mg/L of PAC, despite the strong
487 correlation. In addition, while 9 compounds are very well removed at 10 mg/L of PAC, they are 14 at 20 mg/L and 3 at
488 5 mg/L (Figure 4). Similarly, the number of well or very well removed compounds increases from 12 at 5 mg/L of
489 PAC, to 20 at 10 mg/L and 26 at 20 mg/L. Finally, applying a high dose of fresh PAC (70 mg/L) enables to achieve a
490 very high elimination of micropollutants, with removals higher than 90% for all compounds. In particular, diclofenac
491 which is on the first watch list of the (EC 2013), is moderately removed at 5 mg/L of fresh PAC, well removed at 10
492 mg/L and very well removed at 20 mg/L. Similarly, sulfamethoxazole is very impacted by the PAC dose. Finally, the
493 removal of the total PhPHs concentration is around 76% at 10 mg/L of PAC, 45% at 5 mg/L and 83% at 20 mg/L.

494

495 *Total mass of the PAC in the bed*

496 In normal operation of the pilot, the total mass of the PAC bed is close to 40 kg. Despite substantial variations of the
497 total mass of PAC in the bed (supporting material - Table S1), no significant impact is observed on the performances,
498 highlighting that this parameter is minor. However, the presence of a high mass of PAC in the bed should slightly
499 enhance the performances compared to operating with only the fresh PAC dose, as it was observed with laboratory scale
500 experiments (not presented). This confirms that the observed removals on the pilot are mainly achieved by the fresh
501 PAC dose whereas the quantity of preloaded one in the bed allows a limited enhancement of the overall performances.
502 This observation has to be confirmed by laboratory scale tests in order to clearly identify the role of injected or
503 preloaded PAC.

504

505 *Other operation parameters*

506 The hydraulic velocity has theoretically an influence on contact time. Moreover, contact time has been identified as
507 significantly impacting the adsorption process in the literature (Snyder et al. 2007). However, the hydraulic velocity/the
508 water flow has not a significant impact on the observed contact time in the CarboPlus® process because the variation of
509 the hydraulic velocity is coupled with a modification of the bed depth. Indeed, a higher velocity means a higher bed
510 depth by natural expansion, which balances the decrease of contact time. In addition, the experimental design doesn't
511 allow deeply assessing this point, contrary to the PAC dose. Nevertheless, although a slight impact is observed on
512 efficacy for some compounds (supporting material - Figure S4) such as atenolol, diclofenac or oxazepam, it seems that
513 hydraulic velocity would not be a driving parameter of the process efficacy.

514

515 Finally, the solid retention time (SRT) of PAC and the presence of coagulant and flocculant were not studied in this
516 paper. (Margot et al. 2013) have nevertheless observed a slight increase of micropollutant removal by PAC adsorption
517 in presence of coagulant.

518

519

520 *3.2. Influence of the physico-chemical properties of the compounds on their removals*

521

522 The fate of many pollutants within PAC can be explained by their physico-chemical properties. First of all, the charge
523 of the compound is a crucial parameter in their removal. Indeed, all positively charged compounds (atenolol,
524 propranolol, trimethoprim, ciprofloxacin, norfloxacin) are well removed (>80%), regardless their other properties. This
525 has already been underlined in the literature (de Ridder et al. 2011, Margot et al. 2013). In fact, depending on the point
526 of zero charge (PZC) of the adsorbent, its surface can be neutral or slightly charged (positively or negatively) at influent
527 pH (7-8). In addition, sorption of effluent organic matter, generally negatively charged in wastewater, on activated
528 carbon surface can switch (if initially neutral or positive) or increase (if already negative) the charge, resulting overall in
529 a surface negatively charged (Margot et al. 2013, Yu et al. 2012a). Therefore, this indicates that in this case the PAC
530 surface has negative charges inducing strong electrostatic attraction of positive compounds. In spite of their positive
531 charges, erythromycin and roxithromycin are less removed (50-60%), probably as a results of their high molecular
532 masses (733.5 and 837.0 g/mol respectively; supporting material - Table S4), inducing a higher sensibility to
533 competition with organic matter and other compounds (Ji et al. 2010), and a size exclusion (Moreno-Castilla et al.
534 2003). For these compounds, a slight positive relation with hydrophobicity ($\log D_{ow}$) is moreover observed, i.e.
535 atenolol (80% - $\log D_{ow} = -1.99$) vs. propranolol (94% - $\log D_{ow} = 0.98$). The number of H-bond sites and the

536 compound structure do not seem to have any substantial influence for these compounds. Moreover, considering only
537 compounds with similar log D_{OW} and molecular mass, negatively charged compounds are less removed than positively
538 charged ones (i.e. trimethoprim, propranolol, naproxen, diclofenac, sulfamethoxazole).

539

540 For neutral and negatively charged compounds (Figure 2 and supporting material - Table S3), the hydrophobicity and
541 the structure of the compound seem to play a major role in removal. The influence of hydrophobicity, polarizability or
542 structure of the compound (functional groups allowing H-bond or π - π binding) has already been reported (de Ridder et
543 al. 2010, Delgado et al. 2012). Indeed, the adsorption of the neutral compounds seems to be influenced by their
544 hydrophobicity, as reflected by a significant positive correlation ($r_{\text{spearman}} = 0.587$; p-value < 0.05; $\alpha = 0.05$) found
545 between log D_{OW} and removals of PhPHs. (Westerhoff et al. 2005) observed a similar trend. In addition to the
546 hydrophobicity, specific interactions between PhPHs functional groups and PAC should be involved. Indeed, (de Ridder
547 et al. 2010) have reported that H-binding and π - π interactions become very important at low hydrophobicity, inducing
548 possible substantial differences in the fate of compounds with similar hydrophobicity.

549

550 No significant or specific relations were found for negative compounds between removal and properties.
551 Hydrophobicity was not identified as enhancing their adsorption, in contrary to what (Margot et al. 2013) observed.
552 Absence of direct relationship between adsorption and hydrophobicity was also observed for some antibiotics by (Ji et
553 al. 2010). These compounds are more likely influenced by their structures and their fates should be driven by a balance
554 between electrostatic repulsion and specific interactions with PAC surface. (Moreno-Castilla et al. 2003) concluded that
555 aromatic compounds are mainly physisorbed on activated carbon, confirming the prevalence of these weak interactions.
556 Some functional groups may enhance adsorption such as aromaticity and N-heterocycles (Delgado et al. 2012),
557 explaining the high removal of ofloxacin although this compound is negatively charged. For instance, ofloxacin and
558 diclofenac have three heterocycles, known to enhance adsorption on activated carbon (Delgado et al. 2012), while
559 sulfamethoxazole and sulfadiazine have only one heterocycle.

560

561

562 CONCLUSIONS

563

564 Among the solutions to reduce the emerging contaminant discharges into the environment, the implementation of a
565 tertiary treatment in actual WWTP is more and more considered. Besides, emerging micropollutants, particularly

566 pharmaceuticals, may be included in modifications of existing regulations. For instance, diclofenac, 17- α - and 17- β -
567 estradiols are now on the first watch list of the (EC 2013), and the Swiss Centre for applied ecotoxicology has proposed
568 environmental quality criteria for several emerging micropollutants.

569

570 In this context, the fate of 113 micropollutants within a large scale PAC adsorption pilot (CarboPlus[®]) has been studied.
571 54 pharmaceuticals and hormones (PhPHs) and 59 other micropollutants, such as pesticides, personal care products,
572 phthalates, PAHs or artificial sweeteners, were then monitored. The 14 campaigns performed allowed assessing the
573 overall performances of the process for these compounds, and the influence of the WWTP configuration, the operation
574 parameters and the physico-chemical properties of the micropollutants on the process efficacy.

575

576 26 PhPHs were quantified in influents from nominal WWTP configuration including 8 substances with concentrations
577 higher than 100 ng/L. Sulfamethoxazole is predominant with the highest average concentration, accounting for about
578 30% in average of the sum of the 26 PhPHs. In contrary, the remaining molecules were measured whether at low (<10
579 ng/L), like estrone or lorazepam, or at intermediate concentrations (10-100 ng/L), like naproxen or trimethoprim. 6 of
580 them are poorly or moderately removed by the process (<60%), i.e. paracetamol, ibuprofen or sulfamethoxazole, and 11
581 are rather well removed (60-80%), especially diclofenac, naproxen or oxazepam. The 9 remaining substances are very
582 well eliminated (>80%), i.e. beta blockers, carbamazepine or trimethoprim

583 The change of the WWTP configuration has a substantial impact on the influents quality (DOC, UV-254, BOD₅,
584 nitrogen species, TSS) and most of the PhPHs have higher concentrations in this configuration, highlighting both the
585 seasonal variations and the removal performed due to denitrification step in nominal configuration. Paracetamol and
586 ibuprofen concentrations are multiplied by 100. PhPHs are overall slightly less removed in percentage in this WWTP
587 configuration but a higher flux is eliminated. This is consistent considering the higher concentrations and the
588 degradation of the influents quality which probably enhances the competition effects.

589 Concentrations of the other micropollutants during degraded WWTP configuration are variable but phthalates and
590 artificial sweeteners are present at very high concentrations, up to 1 000-10 000 ng/L, bisphenol A and NP between 100
591 and 1 000 ng/L, and pesticides, perfluorinated acids, PAHs and parabens below 100 ng/L. Except parabens and
592 pesticides, which are largely removed (50-95%), most of the other micropollutants are poorly to moderately removed.
593 However, similarly to PhPHs, it can be assumed that performances would be higher in nominal WWTP configuration
594 for these compounds.

595 The fresh PAC dose has been identified as the main operation parameter which significantly influences the

596 performances for micropollutant removals. The total mass of PAC in the bed was not identified as influencing the
597 performances in the studied range of mass. Overall, in nominal WWTP configuration, the CarboPlus® process reduces
598 in average the sum of the PhPHs of about 53% at a dose of PAC of 5 mg/L, 76% at 10 mg/L and more than 83% at 20
599 mg/L. Finally, a high dose of PAC allows to achieve very high removals (>90%) of all the PhPHs, but at higher costs.
600 The molecular charge seems to be the most important property influencing the fate of micropollutants. However, the
601 size of the molecule can be a limiting factor because high molecular weight compounds are more sensitive to organic
602 matter competition. A higher dose is then needed for heavy compounds compared to others, at comparable other
603 molecular properties. For neutral or negative compounds, hydrophobicity and structure of the molecule, particularly the
604 presence of specific functional groups, become very important in their fates.

605

606 This pilot is still operating in 2014, and a new type of activated carbon is tested. This micro-grain activated carbon
607 (μ GAC) has an intermediary size between PAC (<50 μ m) and grain (>1 mm), that facilitates the bed handling (no need
608 of chemical addition) together with good performances for micropollutants elimination. Moreover, μ GAC selected is
609 produced from regenerated activated carbon and is regenerated several times, ensuring a reduction of cost and a higher
610 sustainability of the process.

611

612 ACKNOWLEDGEMENT

613

614 This study has been performed within the framework of the OPUR research program. The authors would like to thank
615 the SIAAP (Céline Briand and Julien Pouillaude), the LEESU (Damien Lherm), the ISA (Audrey Bulete, Antoine
616 Vauchez, Mikaël Tournier and Loic Spinner) and the SAUR (Séverine Bareilles) teams for their technical support and
617 their active participation to the sampling campaigns. Similarly, authors thank the technical teams of the Technological
618 University of Compiègne (UTC) and the Water Technology Centre (TZW) for their participation in analyses.

619 REFERENCES

620

- 621 Abegglen, C. and Siegrist, H. (2012) Micropolluants dans les eaux résiduaires
622 urbaines. Etapes de traitement supplémentaire dans les stations d'épuration. (in
623 French). Environment, S.F.O.f.t. (ed), p. 87.
- 624 Altmann, J., Ruhl, A.S., Zietzschmann, F. and Jekel, M. (2014) Direct comparison
625 of ozonation and adsorption onto powdered activated carbon for micropollutant
626 removal in advanced wastewater treatment. *Water Research* 55(0), 185-193.
- 627 Barrek, S., Cren-Olivé, C., Wiest, L., Baudot, R., Arnaudguilhem, C. and
628 Grenier-Loustalot, M.-F. (2009) Multi-residue analysis and ultra-trace
629 quantification of 36 priority substances from the European Water Framework
630 Directive by GC-MS and LC-FLD-MS/MS in surface waters. *Talanta* 79(3), 712-722.
- 631 Bergé, A., Cladière, M., Gasperi, J., Coursimault, A., Tassin, B. and Moilleron,
632 R. (2013) Meta-analysis of environmental contamination by phthalates.
633 *Environmental Science and Pollution Research*, 1-20.
- 634 Bergé, A., Gasperi, J., Rocher, V., Gras, L., Coursimault, A. and Moilleron, R.
635 (2014) Phthalates and alkylphenols in industrial and domestic effluents: Case of
636 Paris conurbation (France). *Science of the Total Environment* 488-489(0), 26-35.
- 637 Berset, J.-D. and Ochsenbein, N. (2012) Stability considerations of aspartame in
638 the direct analysis of artificial sweeteners in water samples using high-
639 performance liquid chromatography-tandem mass spectrometry (HPLC-MS/MS).
640 *Chemosphere* 88(5), 563-569.
- 641 Boehler, M., Zwickenpflug, B., Hollender, J., Ternes, T., Joss, A. and Siegrist,
642 H. (2012) Removal of micropollutants in municipal wastewater treatment plants by
643 powder-activated carbon, International Water Association, London, Royaume-Uni.
- 644 Bolong, N., Ismail, A.F., Salim, M.R. and Matsuura, T. (2009) A review of the
645 effects of emerging contaminants in wastewater and options for their removal.
646 *Desalination* 239(1-3), 229-246.
- 647 Bressy, A., Gromaire, M.C., Lorgeoux, C., Saad, M., Leroy, F. and Chebbo, G.
648 (2012) Towards the determination of an optimal scale for stormwater quality
649 management: Micropollutants in a small residential catchment. *Water Research*
650 46(20), 6799-6810.
- 651 Çeçen, F. and Aktas, Ö. (2012) Activated carbon for water and wastewater
652 treatment integration of adsorption and biological treatment, Wiley-VCH,
653 Weinheim, Germany.
- 654 Chen, Z.-F., Ying, G.-G., Lai, H.-J., Chen, F., Su, H.-C., Liu, Y.-S., Peng, F.-
655 Q. and Zhao, J.-L. (2012) Determination of biocides in different environmental
656 matrices by use of ultra-high-performance liquid chromatography-tandem mass
657 spectrometry. *Analytical and Bioanalytical Chemistry* 404(10), 3175-3188.
- 658 Clara, M., Strenn, B., Gans, O., Martinez, E., Kreuzinger, N. and Kroiss, H.
659 (2005) Removal of selected pharmaceuticals, fragrances and endocrine disrupting
660 compounds in a membrane bioreactor and conventional wastewater treatment plants.
661 *Water Research* 39(19), 4797-4807.
- 662 Clara, M., Scharf, S., Scheffknecht, C. and Gans, O. (2007) Occurrence of
663 selected surfactants in untreated and treated sewage. *Water Research* 41, 4339-
664 4348.
- 665 Daughton, C.G. and Ternes, T.A. (1999) Pharmaceuticals and personal care
666 products in the environment: Agents of subtle change? *Environmental Health*
667 *Perspectives* 107, 907-938.
- 668 de Ridder, D.J., Villacorte, L., Verliefe, A.R.D., Verberk, J.Q.J.C., Heijman,
669 S.G.J., Amy, G.L. and van Dijk, J.C. (2010) Modeling equilibrium adsorption of
670 organic micropollutants onto activated carbon. *Water Research* 44(10), 3077-3086.
- 671 de Ridder, D.J., Verliefe, A.R., Heijman, S.G., Verberk, J.Q., Rietveld, L.C.,
672 van der Aa, L.T., Amy, G.L. and van Dijk, J.C. (2011) Influence of natural
673 organic matter on equilibrium adsorption of neutral and charged pharmaceuticals
674 onto activated carbon. *Water Sci Technol* 63(3), 416-423.
- 675 de Ridder, D.J., Verliefe, A.R.D., Schoutteten, K., van der Linden, B.,
676 Heijman, S.G.J., Beurroies, I., Denoyel, R., Amy, G.L. and van Dijk, J.C. (2013)
677 Relation between interfacial energy and adsorption of organic micropollutants
678 onto activated carbon. *Carbon* 53(0), 153-160.
- 679 Deblonde, T., Cossu-Leguille, C. and Hartemann, P. (2011) Emerging pollutants in
680 wastewater: A review of the literature. *International Journal of Hygiene and*
681 *Environmental Health* 214(6), 442-448.
- 682 Delgado, L.F., Charles, P., Glucina, K. and Morlay, C. (2012) The removal of
683 endocrine disrupting compounds, pharmaceutically activated compounds and
684 cyanobacterial toxins during drinking water preparation using activated carbon—
685 A review. *Science of the Total Environment* 435-436(0), 509-525.
- 686 Ebie, K., Li, F., Azuma, Y., Yuasa, A. and Hagishita, T. (2001) Pore
687 distribution effect of activated carbon in adsorbing organic micropollutants

688 from natural water. *Water Research* 35(1), 167-179.

689 EC (2013) Directive 2013/39/EU of the European Parliament amending Directives
690 2000/60/EC and 2008/105/EC as regards priority substances in the field of water
691 policy. JO-EU L226/1.

692 Fatone, F., Di Fabio, S., Bolzonella, D. and Cecchi, F. (2011) Fate of aromatic
693 hydrocarbons in Italian municipal wastewater systems: An overview of wastewater
694 treatment using conventional activated-sludge processes (CASp) and membrane
695 bioreactors (MBRs). *Water Research* 45(1), 93-104.

696 Gasperi, J., Kafi-Benyahia, M., Lorgeoux, C., Moilleron, R., Gromaire, M.C. and
697 Chebbo, G. (2008) Wastewater quality and pollutant loads in combined sewers
698 during dry weather periods. *Urban Water Journal* 5(4), 305-314.

699 Gasperi, J., Geara, D., Lorgeoux, C., Bressy, A., Zedek, S., Rocher, V., El
700 Samrani, A., Chebbo, G. and Moilleron, R. (2014) First assessment of triclosan,
701 triclocarban and paraben mass loads at a very large regional scale: Case of
702 Paris conurbation (France). *Science of the Total Environment* 493(0), 854-861.

703 Geara-Matta, D. (2012) Flux and sources of parabens, triclosan and triclocarban
704 in dense urban areas: comparison between Paris and Beyrouth (in French), *Ecole
705 des Ponts ParisTech*.

706 Götz, C., Stamm, C., Fenner, K., Singer, H., Schärer, M. and Hollender, J.
707 (2010) Targeting aquatic microcontaminants for monitoring: exposure
708 categorization and application to the Swiss situation. *Environmental Science and
709 Pollution Research* 17(2), 341-354.

710 Halling-Sørensen, B., Nors Nielsen, S., Lanzky, P.F., Ingerslev, F., Holten
711 Lützhøft, H.C. and Jørgensen, S.E. (1998) Occurrence, fate and effects of
712 pharmaceutical substances in the environment- A review. *Chemosphere* 36(2), 357-
713 393.

714 Heberer, T. (2002) Occurrence, fate, and removal of pharmaceutical residues in
715 the aquatic environment: a review of recent research data. *Toxicology Letters*
716 131(1-2), 5-17.

717 Ji, L., Liu, F., Xu, Z., Zheng, S. and Zhu, D. (2010) Adsorption of
718 Pharmaceutical Antibiotics on Template-Synthesized Ordered Micro- and Mesoporous
719 Carbons. *Environmental Science & Technology* 44(8), 3116-3122.

720 Jones, O.A.H., Voulvoulis, N. and Lester, J.N. (2001) Human Pharmaceuticals in
721 the Aquatic Environment a Review. *Environmental Technology* 22(12), 1383-1394.

722 Joss, A., Keller, E., Alder, A.C., Göbel, A., McArdell, C.S., Ternes, T. and
723 Siegrist, H. (2005) Removal of pharmaceuticals and fragrances in biological
724 wastewater treatment. *Water Research* 39(14), 3139-3152.

725 Joss, A., Siegrist, H. and Ternes, T.A. (2008) Are we about to upgrade
726 wastewater treatment for removing organic micropollutants? *Water Sci Technol*
727 57(2), 251-255.

728 Kase, R., Eggen, R.I.L., Junghans, M., Götz, C. and Hollender, J. (2011) Waste
729 Water - Evaluation and Management. *Einschlag, P.F.S.G. (ed), InTech*.

730 Lange, F., Scheurer, M. and Brauch, H.-J. (2012) Artificial sweeteners—a
731 recently recognized class of emerging environmental contaminants: a review.
732 *Analytical and Bioanalytical Chemistry* 403(9), 2503-2518.

733 Loos, R., Carvalho, R., António, D.C., Comero, S., Locoro, G., Tavazzi, S.,
734 Paracchini, B., Ghiani, M., Lettieri, T., Blaha, L., Jarosova, B., Voorspoels,
735 S., Servaes, K., Haglund, P., Fick, J., Lindberg, R.H., Schwesig, D. and Gawlik,
736 B.M. (2013) EU-wide monitoring survey on emerging polar organic contaminants in
737 wastewater treatment plant effluents. *Water Research* 47(17), 6475-6487.

738 Löwenberg, J., Zenker, A., Baggenstos, M., Koch, G., Kazner, C. and Wintgens, T.
739 (2014) Comparison of two PAC/UF processes for the removal of micropollutants
740 from wastewater treatment plant effluent: Process performance and removal
741 efficiency. *Water Research* 56(0), 26-36.

742 Luo, Y., Guo, W., Ngo, H.H., Nghiem, L.D., Hai, F.I., Zhang, J., Liang, S. and
743 Wang, X.C. (2014) A review on the occurrence of micropollutants in the aquatic
744 environment and their fate and removal during wastewater treatment. *Science of
745 the Total Environment* 473-474(0), 619-641.

746 Mailler, R., Gasperi, J., Rocher, V., Gilbert-Pawlik, S., Geara-Matta, D.,
747 Moilleron, R. and Chebbo, G. (2013) Biofiltration vs conventional activated
748 sludge plants: what about priority and emerging pollutants removal?
749 *Environmental Science and Pollution Research*, 1-12.

750 Margot, J., Kienle, C., Magnet, A., Weil, M., Rossi, L., de Alencastro, L.F.,
751 Abegglen, C., Thonney, D., Chèvre, N., Schärer, M. and Barry, D.A. (2013)
752 Treatment of micropollutants in municipal wastewater: Ozone or powdered
753 activated carbon? *Science of the Total Environment* 461-462(0), 480-498.

754 McAdam, E.J., Bagnall, J.P., Koh, Y.K.K., Chiu, T.Y., Pollard, S., Scrimshaw,
755 M.D., Lester, J.N. and Cartmell, E. (2010) Removal of steroid estrogens in
756 carbonaceous and nitrifying activated sludge processes. *Chemosphere* 81(1), 1-6.

757 Miège, C., Choubert, J.M., Ribeiro, L., Eusèbe, M. and Coquery, M. (2009) Fate
758 of pharmaceuticals and personal care products in wastewater treatment plants -
759 Conception of a database and first results. *Environmental Pollution* 157(5),
760 1721-1726.

761 Moreno-Castilla, C., Bautista-Toledo, I., Ferro-Garcia, M.A. and Rivera-Utrilla,
762 J. (2003) Influence of support surface properties on activity of bacteria
763 immobilised on activated carbons for water denitrification. *Carbon* 41, 1743-
764 1749.

765 Ordóñez, E.Y., Quintana, J.B., Rodil, R. and Cela, R. (2012) Determination of
766 artificial sweeteners in water samples by solid-phase extraction and liquid
767 chromatography-tandem mass spectrometry. *Journal of Chromatography A* 1256(0),
768 197-205.

769 Qiao, M., Qi, W., Liu, H. and Qu, J. (2014) Occurrence, behavior and removal of
770 typical substituted and parent polycyclic aromatic hydrocarbons in a biological
771 wastewater treatment plant. *Water Research* 52(0), 11-19.

772 Radjenović, J., Petrović, M. and Barceló, D. (2009) Fate and distribution of
773 pharmaceuticals in wastewater and sewage sludge of the conventional activated
774 sludge (CAS) and advanced membrane bioreactor (MBR) treatment. *Water Research*
775 43(3), 831-841.

776 Rocher, V., Paffoni, C., Goncalves, A., Guerin, S., Azimi, S., Gasperi, J.,
777 Moilleron, R. and Pauss, A. (2012) Municipal wastewater treatment by
778 biofiltration: comparisons of various treatment layouts. Part 1: assessment of
779 carbon and nitrogen removal. *Water Science and Technology* 65(9), 1705-1712.

780 Ruel, S.M., Choubert, J.M., Esperanza, M., Miego, C., Navalon Madrugal, P.,
781 Budzinski, H., Le Menach, K., Lazarova, V. and Coquery, M. (2011) On-site
782 evaluation of the removal of 100 micro-pollutants through advanced wastewater
783 treatment processes for reuse applications. *Water Science and Technology* 63(11),
784 2486-2497.

785 Ruel, S.M., Choubert, J.M., Budzinski, H., Miego, C., Esperanza, M. and Coquery,
786 M. (2012) Occurrence and fate of relevant substances in wastewater treatment
787 plants regarding Water Framework Directive and future legislations. *Water*
788 *Science and Technology* 65(7), 1179-1189.

789 Sánchez-Avila, J., Fernandez-Sanjuan, M., Vicente, J. and Lacorte, S. (2011)
790 Development of a multi-residue method for the determination of organic
791 micropollutants in water, sediment and mussels using gas chromatography-tandem
792 mass spectrometry. *Journal of Chromatography A* 1218(38), 6799-6811.

793 Scheurer, M., Brauch, H.-J. and Lange, F. (2009) Analysis and occurrence of
794 seven artificial sweeteners in German waste water and surface water and in soil
795 aquifer treatment (SAT). *Analytical and Bioanalytical Chemistry* 394(6), 1585-
796 1594.

797 Scheurer, M., Storck, F.R., Brauch, H.-J. and Lange, F.T. (2010) Performance of
798 conventional multi-barrier drinking water treatment plants for the removal of
799 four artificial sweeteners. *Water Research* 44(12), 3573-3584.

800 Snyder, S.A., Adham, S., Redding, A.M., Cannon, F.S., DeCarolis, J.,
801 Oppenheimer, J., Wert, E.C. and Yoon, Y. (2007) Role of membranes and activated
802 carbon in the removal of endocrine disruptors and pharmaceuticals. *Desalination*
803 202(1-3), 156-181.

804 Vulliet, E., Cren-Olivé, C. and Grenier-Loustalot, M.-F. (2011) Occurrence of
805 pharmaceuticals and hormones in drinking water treated from surface waters.
806 *Environmental Chemistry Letters* 9(1), 103-114.

807 Westerhoff, P., Yoon, Y., Snyder, S. and Wert, E. (2005) Fate of Endocrine-
808 Disruptor, Pharmaceutical, and Personal Care Product Chemicals during Simulated
809 Drinking Water Treatment Processes. *Environmental Science & Technology* 39(17),
810 6649-6663.

811 Yu, J., Lv, L., Lan, P., Zhang, S., Pan, B. and Zhang, W. (2012a) Effect of
812 effluent organic matter on the adsorption of perfluorinated compounds onto
813 activated carbon. *Journal of Hazardous Materials* 225-226(0), 99-106.

814 Yu, K., Li, B. and Zhang, T. (2012b) Direct rapid analysis of multiple PPCPs in
815 municipal wastewater using ultrahigh performance liquid chromatography-tandem
816 mass spectrometry without SPE pre-concentration. *Analytica Chimica Acta* 738(0),
817 59-68.

818 Yu, Z., Peldszus, S. and Huck, P.M. (2008) Adsorption characteristics of
819 selected pharmaceuticals and an endocrine disrupting compound—Naproxen,
820 carbamazepine and nonylphenol—on activated carbon. *Water Research* 42(12), 2873-
821 2882.

822 Zareitalabad, P., Siemens, J., Hamer, M. and Amelung, W. (2013)
823 Perfluorooctanoic acid (PFOA) and perfluorooctanesulfonic acid (PFOS) in surface
824 waters, sediments, soils and wastewater - A review on concentrations and
825 distribution coefficients. *Chemosphere* 91(6), 725-732.

826
827
828

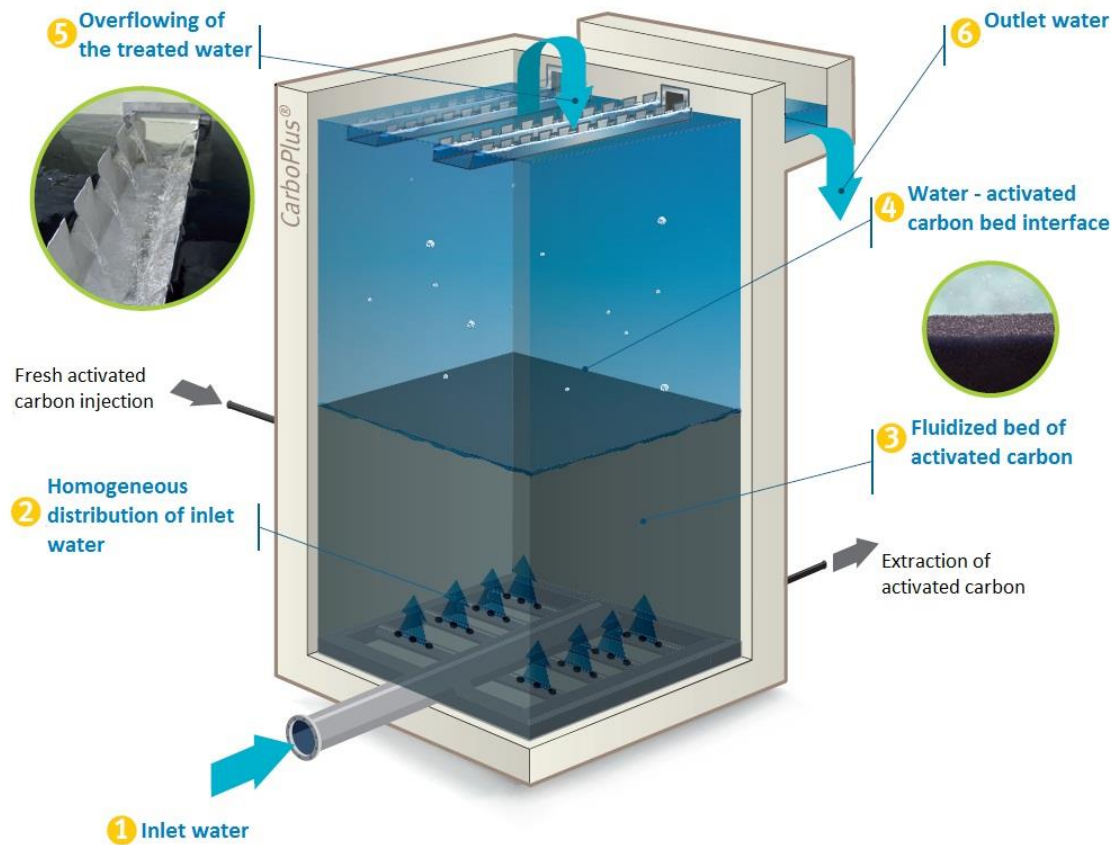


Figure 1 - Layout of the 20 m³ (5 m x 4 m²) CarboPlus® pilot (SAUR source)

829
830
831

832

Table 1 - Operation parameters for the 14 campaigns performed

	Date	WWTP configuration	PAC dose	Hydraulic velocity	Total PAC mass (kg)	
C1*	02/07/13	Nominal	70	8	-	
C2	09/07/13		20			
C3	16/07/13		5			
C4	23/07/13		20	10		
C5	30/07/13		10	6		
C6	06/08/13		5	8		
C7	20/08/13		10	7		
C8	27/08/13					62
C9	04/11/13					
C10	19/11/13	Degraded	10	7	46	
C11	27/11/13				52	
C12	03/12/13				53	
C13	10/12/13				45	
C14	17/12/13				58	

*C for campaign.

833

834

835

836

837

Table 2 - Pollutants studied and their analytical methods

Groups	N ^a	n ^b	Reference	Extraction ^c	Analysis ^d	LQ ^e
Antibiotics	31	14				0.2 - 110
Analgesic	5	14		SPE		0.04 - 1.0
Beta blockers	2	14		Autotrace®		0.02
Anxiolytics	4	14	(Vulliet et al. 2011)	StrataX® or Oasis HLB®	LC-MSMS	1.0 - 2.0
Hypolipemians	2	14		cartridges		0.7 - 2.4
Hormones	10	14				1.0
Chlorinated solvents	3	5	(Barrek et al. 2009)		GC-MS	0.2 - 23
Perfluorinated acids	2	5	(Vulliet et al. 2011)	SPE	LC-MSMS	1.0
Pesticides / herbicides / insecticides	23	5	(Barrek et al. 2009)	Autotrace®	GC-MS	6 - 177
				StrataX® cartridges	LC-MSMS	0.2
Bisphenol A	1	5	(Vulliet et al. 2011)		LC-MSMS	10
PAHs ^f	13	3	(Bressy et al. 2012)	SPE		0.2
Alkylphenols	2	5	(Bergé et al. 2014)	Manual	GC-MS	0.6 - 100
Phthalates	4	5	(Bergé et al. 2014)	Oasis HLB® cartridges		100
PCPs ^f	7	3	(Gasperi et al. 2014)		LC-MSMS	2.8 - 10
				SPE		
Sweeteners	4	4	(Scheurer et al. 2009)	Autotrace®	LC-MSMS	50 - 250
				Bakerbond SDB1 cartridges		
<i>Total micropollutants</i>	<i>113</i>					

^a N = number of substances.
^b n = number of campaigns performed.
^c SPE = solid phase extraction.
^d Analytical methods: LC = liquid chromatography, GC = gas chromatography, GC-MS = GC with mass spectrometry, LC-MSMS = LC with tandem mass spectrometry.
^e LQ = limit of quantification.
^f PAHs = polycyclic aromatic hydrocarbons; PCPs = personal care products.

Table 3 - Micropollutant and conventional wastewater parameter concentrations in influents and effluents during nominal and degraded WWTP configuration

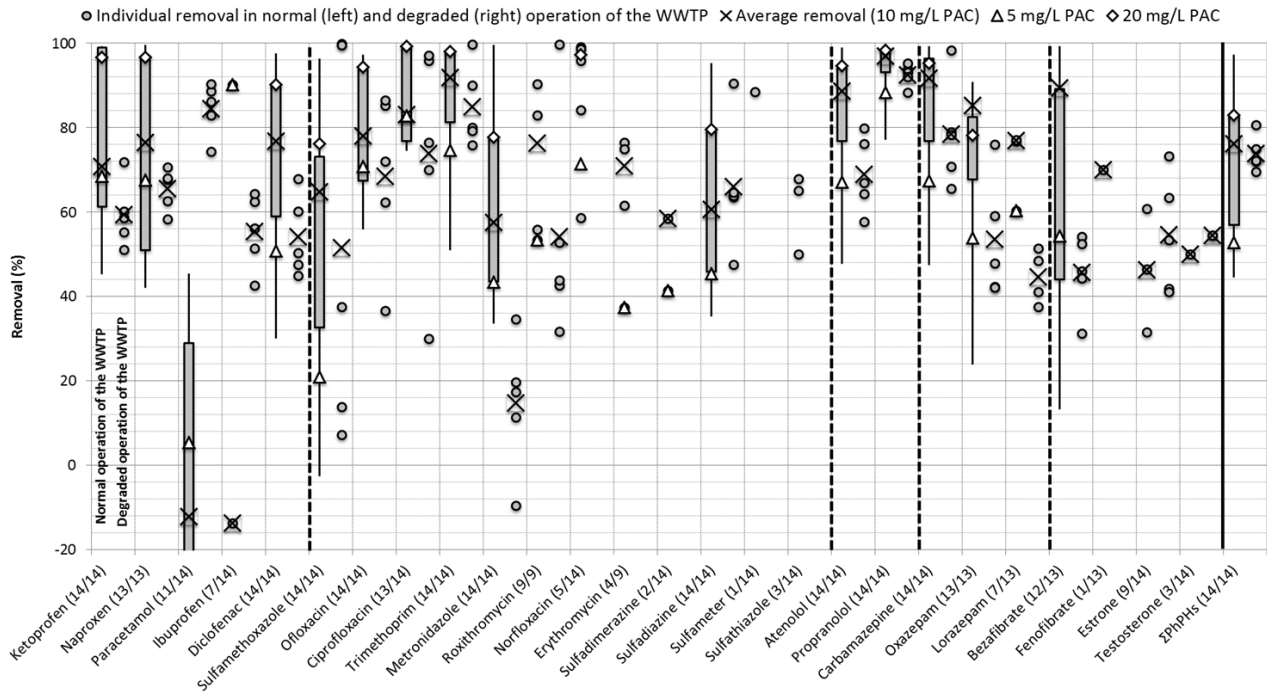
	LQ (ng/L)	Nominal WWTP configuration (n=9)				Degraded WWTP configuration (n=5)			
		Influent		Effluent		Influent		Effluent	
		N	Concentration (ng/L) Av. ± SD (min - max)	N	Concentration (ng/L) Av. ± SD (min - max)	N	Concentration (ng/L) Av. ± SD (min - max)	N	Concentration (ng/L) Av. ± SD (min - max)
<i>Conventional wastewater parameters</i>									
UV 254 nm (cm ⁻¹)	0.01	9/9	0.110 ± 0.013 (0.087 - 0.120)	9/9	0.075 ± 0.020 (0.047 - 0.106)	5/5	0.139 ± 0.011 (0.126 - 0.152)	5/5	0.109 ± 0.008 (0.099 - 0.120)
DOC (mgC/L)	0.5	9/9	5.6 ± 0.9 (4.4 - 7.0)	9/9	3.5 ± 1.2 (1.2 - 5.2)	5/5	7.5 ± 0.5 (7.1 - 8.3)	5/5	5.8 ± 0.4 (5.4 - 6.4)
COD (mgO ₂ /L)	4	9/9	26 ± 11 (17 - 47)	9/9	34 ± 22 (13 - 74)	5/5	26 ± 2 (23 - 28)	5/5	21 ± 2 (19 - 25)
BOD ₅ (mgO ₂ /L)	0.5	9/9	4.8 ± 3.3 (2.8 - 11.0)	9/9	3.5 ± 1.9 (1.3 - 7.0)	5/5	5.8 ± 0.9 (4.9 - 7.0)	5/5	3.0 ± 0.5 (2.4 - 3.7)
NH ₄ ⁺ (mgN/L)	0.3	0/9	<LQ	0/9	<LQ	5/5	4.4 ± 0.7 (3.2 - 5.0)	5/5	3.5 ± 0.5 (2.9 - 4.0)
NO ₂ ⁻ (mgN/L)	0.02	9/9	0.21 ± 0.17 (0.06 - 0.23)	8/9	0.15 ± 0.16 (<LQ - 0.30)	5/5	0.7 ± 0.02 (0.05 - 0.09)	5/5	0.29 ± 0.16 (0.03 - 0.46)
NO ₃ ⁻ (mgN/L)	0.4	9/9	9.1 ± 2.9 (5.6 - 11.7)	9/9	8.5 ± 2.8 (4.9 - 13.7)	5/5	23.7 ± 2.1 (20.6 - 25.4)	5/5	24.4 ± 2.5 (20.7 - 26.6)
TKN (mgN/L)	0.5	9/9	1.5 ± 0.2 (1.1 - 1.8)	9/9	1.2 ± 0.1 (0.9 - 1.3)	5/5	4.9 ± 0.5 (4.1 - 5.5)	5/5	3.9 ± 0.4 (3.4 - 4.4)
PO ₄ ³⁻ (mgP/L)	0.1	0/5	<LQ	0/5	<LQ	0/5	<LQ	0/5	<LQ
TP (mgP/L)	0.3	0/5	<LQ	0/5	<LQ	0/5	<LQ	0/5	<LQ
TSS (mg/L)	2.0	9/9	3 ± 1 (2 - 5)	9/9	9 ± 7 (2 - 21)	5/5	6 ± 2 (4 - 9)	5/5	5 ± 2 (2 - 6)
<i>Pharmaceuticals and hormones</i>									
Ketoprofen	0.3	9/9	34 ± 18 (13 - 61)	6/9	11 ± 10 (<LQ - 30)	5/5	367 ± 90 (273 - 501)	5/5	145 ± 28 (109 - 182)
Naproxen	0.7	8/8	33 ± 28 (8 - 83)	5/8	12 ± 13 (<LQ - 34)	5/5	154 ± 39 (122 - 220)	5/5	52 ± 9 (43 - 64)
Paracetamol	0.04	6/9	33 ± 22 (<LQ - 70)	6/9	32 ± 18 (<LQ - 56)	5/5	5 870 ± 2 597 (3 610 - 10 350)	5/5	1 030 ± 923 (349 - 2 650)
Ibuprofen	1.0	2/9	9 (<LQ - 13)	5/9	9 ± 5 (<LQ - 14)	5/5	951 ± 360 (590 - 1 439)	5/5	432 ± 200 (220 - 699)
Diclofenac	0.3	9/9	184 ± 91 (95 - 309)	9/9	52 ± 51 (3 - 166)	5/5	384 ± 76 (301 - 508)	5/5	171 ± 19 (147 - 196)
Sulfamethoxazole	1.0	9/9	993 ± 817 (175 - 3 010)	9/9	419 ± 318 (37 - 798)	5/5	233 ± 179 (70 - 470)	3/5	130 (<LQ - 235)
Ofloxacin	10	9/9	412 ± 314 (14 - 911)	8/9	70 ± 74 (<LQ - 218)	5/5	39 ± 7 (34 - 51)	3/5	18 (<LQ - 24)
Ciprofloxacin	1.0	8/9	175 ± 93 (64 - 312)	6/9	22 ± 17 (<LQ - 51)	5/5	13 ± 3 (11 - 17)	3/5	5 (<LQ - 8)
Trimethoprim	0.3	9/9	64 ± 79 (8 - 222)	8/9	4 ± 2 (<LQ - 8)	5/5	43 ± 28 (9 - 74)	4/5	6 ± 3 (<LQ - 9)
Metronidazole	0.2	9/9	19 ± 3 (15 - 24)	8/9	8 ± 3 (<LQ - 12)	5/5	30 ± 4 (23 - 33)	5/5	25 ± 4 (18 - 29)
Roxithromycin	1.0	4/4	99 ± 53 (57 - 173)	4/4	35 ± 33 (5 - 76)	5/5	175 ± 59 (126 - 271)	4/5	99 ± 42 (<LQ - 155)
Norfloxacin	1.0	5/9	80 ± 36 (<LQ - 118)	3/9	17 (<LQ - 29)	0/5	<LQ	0/5	<LQ
Erythromycin	1.0	4/4	124 ± 32 (97 - 170)	4/4	50 ± 38 (23 - 106)	0/5	<LQ	0/5	<LQ
Sulfadimerazine	1.0	2/9	4 (<LQ - 7)	1/9	(<LQ - 4)	0/5	<LQ	0/5	<LQ
Sulfadiazine	1.0	9/9	10 ± 6 (1 - 21)	7/9	4 ± 3 (<LQ - 8)	5/5	4 ± 1 (2 - 5)	4/5	1 ± 1 (<LQ - 2)
Sulfameter	1.0	1/9	(<LQ - 4)	0/9	<LQ	0/5	<LQ	0/5	<LQ
Sulfathiazole	1.0	3/9	1	0/9	<LQ	0/5	<LQ	0/5	<LQ

			(1 - 2)						
Atenolol	0.02	9/9	185 ± 51 (124 - 251)	9/9	28 ± 24 (2 - 67)	5/5	588 ± 156 (454 - 858)	5/5	176 ± 42 (126 - 236)
Propranolol	0.02	9/9	97 ± 27 (66 - 131)	9/9	5 ± 5 (1 - 15)	5/5	197 ± 46 (162 - 276)	5/5	14 ± 5 (10 - 23)
Carbamazepine	1.0	9/9	215 ± 85 (19 - 321)	8/9	41 ± 43 (<i><LQ</i> - 113)	5/5	30 ± 2 (27 - 32)	4/5	8 ± 2 (<i><LQ</i> - 11)
Oxazepam	1.2	8/8	139 ± 128 (20 - 354)	8/8	29 ± 24 (5 - 65)	5/5	409 ± 132 (251 - 615)	5/5	176 ± 31 (145 - 210)
Lorazepam	1.9	3/8	11 (<i><LQ</i> - 27)	1/8	(<i><LQ</i> - 6)	4/5	34 ± 2 (<i><LQ</i> - 36)	5/5	18 ± 3 (15 - 23)
Bezafibrate	0.7	8/8	36 ± 42 (1 - 102)	6/8	8 ± 9 (<i><LQ</i> - 26)	5/5	369 ± 330 (151 - 940)	5/5	187 ± 149 (84 - 447)
Fenofibrate	2.4	1/8	(<i><LQ</i> - 4)	0/8	<i><LQ</i>	0/8	<i><LQ</i>	0/8	<i><LQ</i>
Estrone	1.0	4/8	7 ± 2 (<i><LQ</i> - 10)	5/8	9 ± 10 (<i><LQ</i> - 26)	5/5	12 ± 3 (9 - 15)	5/5	5 ± 1 (4 - 7)
Testosterone	1.0	1/8	(<i><LQ</i> - 1)	0/8	<i><LQ</i>	2/5	(<i><LQ</i> - 1)	0/5	<i><LQ</i>
ΣPhPHs	-	-	2 729 ± 1 057 (1 358 - 5 158)	-	752 ± 517 (75 - 1 593)	-	9 892 ± 2 569 (6 844 - 13 910)	-	2 625 ± 1 007 (1 711 - 4 226)
<i>Other emerging micropollutants</i>									
Atrazine	0.2					5/5	4 ± 1 (3 - 5)	5/5	2 ± 1 (1 - 2)
Diuron	0.2					5/5	25 ± 5 (19 - 30)	5/5	4 ± 1 (2 - 5)
Isoproturon	0.2					5/5	34 ± 22 (11 - 62)	5/5	12 ± 6 (5 - 20)
PFOA	1.0					5/5	25 ± 9 (14 - 36)	5/5	37 ± 19 (16 - 57)
PFOS	1.0					5/5	44 ± 12 (24 - 56)	5/5	31 ± 13 (14 - 47)
Bisphenol A	10					5/5	259 ± 148 (142 - 513)	5/5	78 ± 24 (52 - 112)
NP	100					5/5	841 ± 681 (143 - 1 895)	3/5	286 (<i><LQ</i> - 390)
t-OP	0.6					5/5	17 ± 5 (10 - 22)	5/5	16 ± 15 (4 - 39)
DEHP	100					4/4	1 413 ± 862 (919 - 2 704)	4/4	861 ± 489 (377 - 1 300)
DEP	100					5/5	991 ± 443 (441 - 1 644)	5/5	437 ± 149 (298 - 646)
DnBP	100					5/5	932 ± 631 (355 - 1 986)	5/5	607 ± 383 (309 - 1 280)
Acesulfame	50					4/4	8 725 ± 602 (7 900 - 9 300)	4/4	7 525 ± 665 (6 700 - 8 200)
Cyclamate	50					2/4	240 (<i><LQ</i> - 430)	0/4	<i><LQ</i>
Saccharin	50					4/4	1 355 ± 572 (840 - 2 100)	4/4	703 ± 284 (450 - 1 100)
Sucralose	250					4/4	7 150 ± 545 (6 500 - 7 800)	4/4	5 875 ± 222 (5 600 - 6 100)
Triclosan	10					3/3	135 (121 - 158)	3/3	103 (86 - 119)
Methyl paraben	2.8					3/3	65 (31 - 102)	3/3	11 (10 - 13)
Ethyl paraben	3.5					3/3	40 (27 - 57)	2/3	5 (<i><LQ</i> - 7)
Propyl paraben	3.3					3/3	56 (37 - 81)	2/3	5 (<i><LQ</i> - 5)
Benzyl paraben	3.0					1/3	(<i><LQ</i> - 4)	1/3	(<i><LQ</i> - 3)
Butyl paraben	2.8					2/3	10 (<i><LQ</i> - 13)	0/3	<i><LQ</i>
Fluorene	0.2					3/3	1.7 (0.3 - 4.2)	3/3	3.0 (2.6 - 3.4)
Phenanthrene	0.2					3/3	16.7	3/3	24.2

Anthracene	0.2
Fluoranthene	0.2
Pyrene	0.2
Benzo[a]anthracene	0.2
Chrysene	0.2
Benzo[b]fluoranthene	0.2
Benzo[k]fluoranthene	0.2
Benzo[a]pyrene	0.2
Indeno[123]pyrene	0.2
Dibenzo[ah]anthracene	0.2
Benzo[ghi]perylene	0.2

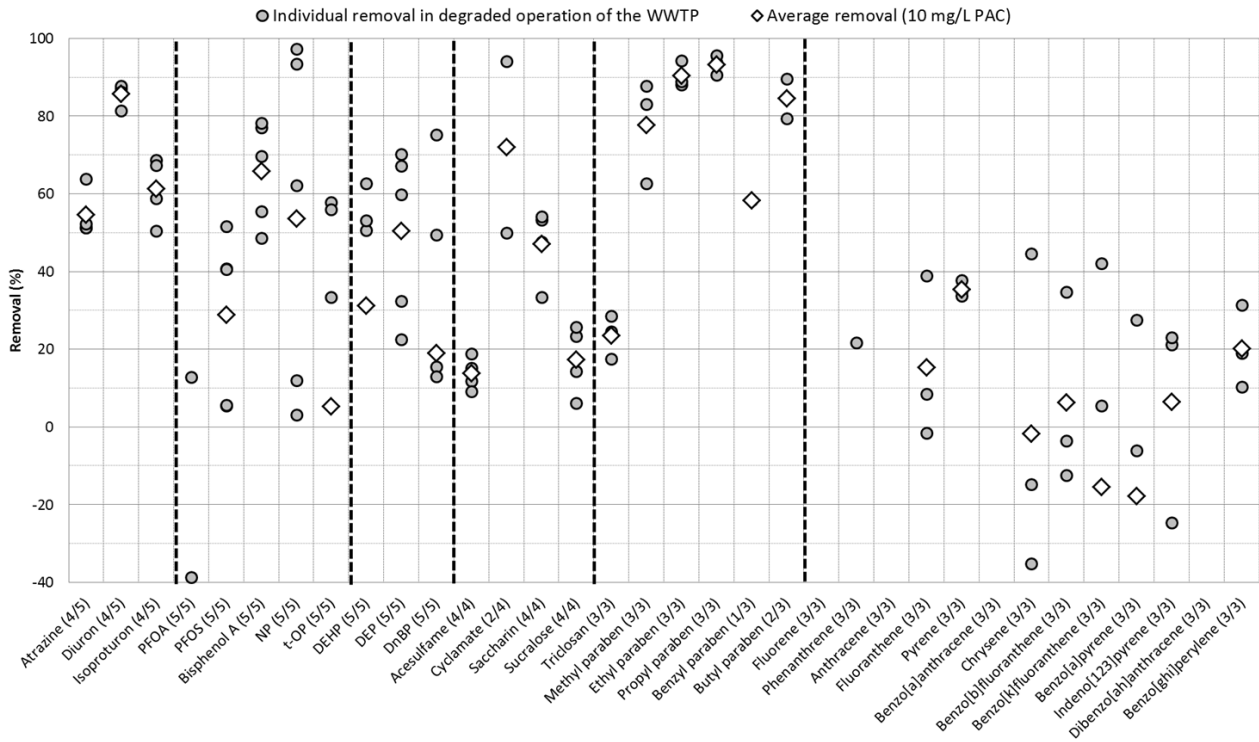
	(13.0 - 19.1)		(10.2 - 33.3)
3/3	0.6 (0.5 - 0.8)	3/3	16.7 (13.0 - 19.1)
3/3	3.2 (2.6 - 3.9)	3/3	2.6 (2.4 - 2.8)
3/3	2.4 (2.1 - 2.8)	3/3	1.6 (1.4 - 1.8)
0/3	<LQ	1/3	(<LQ - 0.9)
3/3	1.5 (1.2 - 1.6)	3/3	1.5 (0.7 - 1.9)
3/3	3.3 (2.2 - 4.0)	3/3	3.2 (1.5 - 4.2)
3/3	1.2 (0.9 - 1.4)	3/3	1.5 (0.5 - 2.5)
3/3	1.4 (0.9 - 1.7)	3/3	1.8 (0.7 - 3.0)
3/3	2.0 (1.0 - 2.6)	3/3	2.0 (0.8 - 3.3)
3/3	0.5 (0.3 - 0.6)	3/3	1.8 (0.3 - 4.3)
3/3	2.0 (1.1 - 2.6)	3/3	1.7 (0.7 - 2.3)

LQ = limit of quantification.
N = number of quantification / campaigns performed.
Av. ± SD = average ± standard deviation (calculated only when N > 3).



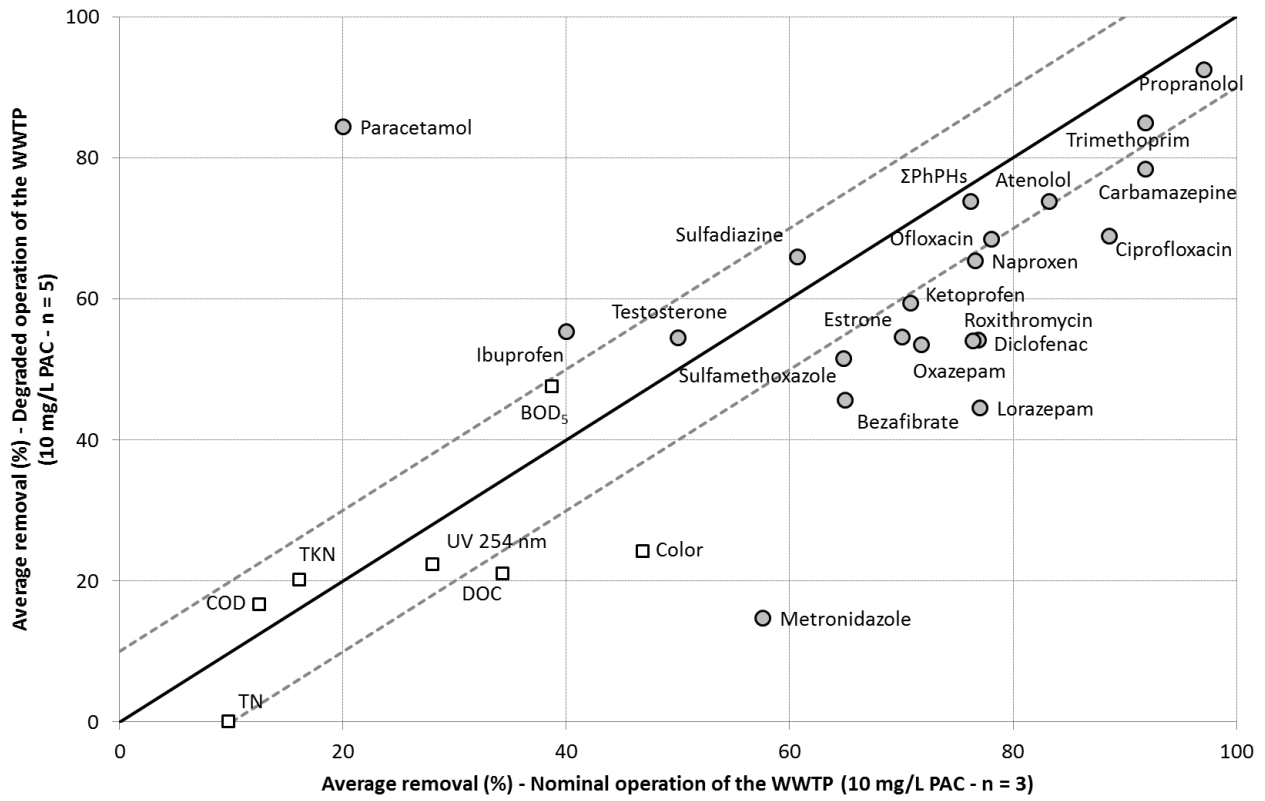
844
 845
 846
 847

Figure 2 - Removals of pharmaceuticals and hormones (N=26) by PAC treatment during nominal (left) and degraded (right) WWTP configurations (min - Q1 - Q2 - Q3 - max or individual values) - occurrence in brackets

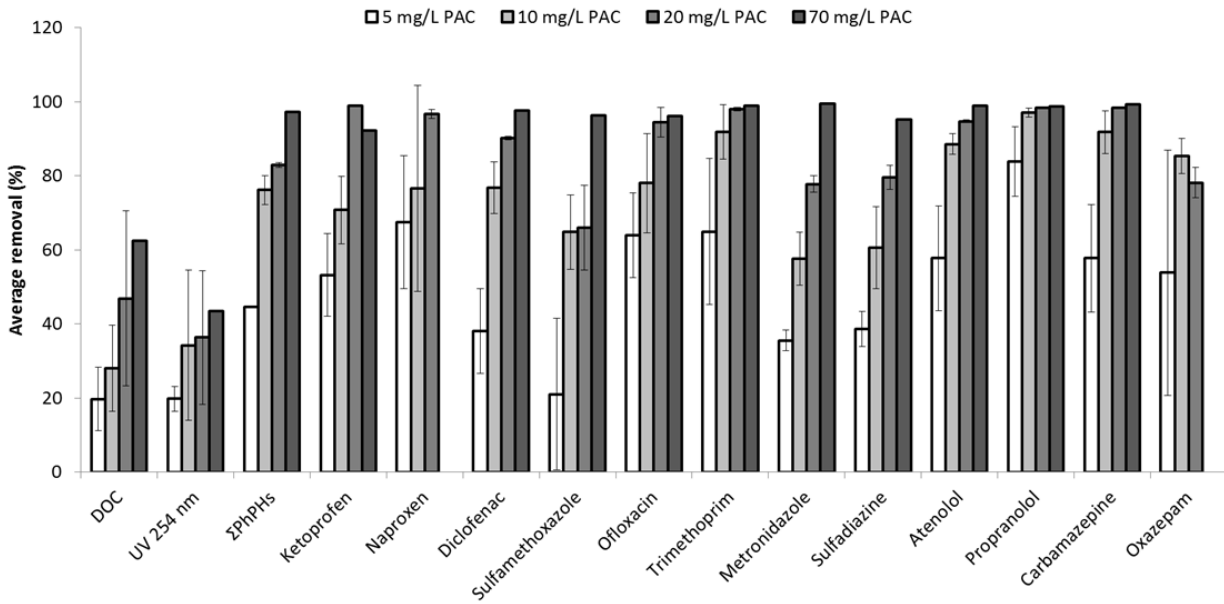


848
849
850

Figure 3 - Removals of the other emerging micropollutants by PAC, in degraded configuration of the WWTP.



851
852



853
854
855
856
857
858

Figure 4 - Parameters influencing the efficiency of PAC to remove emerging micropollutants from treated wastewater - influence of the water quality (top) and the fresh PAC dose (bottom)