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1	On the Use of the Similar Media Concept for Scaling Soil Air Permeability
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16 Abstract

Soil air permeability (k_a) is an important factor that controls subsurface gas transport and 17 exchange of gas across the soil-atmosphere interface. It is thus crucial to evaluate the spatial 18 distribution of k_a for both application and modeling purposes. However, relevant studies are still 19 very limited, partly due to the fact that the dependence of k_a on soil moisture levels cannot be 20 21 directly included in the methods such as geostatistical techniques for analyzing the spatial 22 distribution of k_a . To tackle this problem, the scaling scheme based on the similar media concept, which has been widely used in soil hydrology for characterizing spatial variability of soil 23 hydraulic properties, was employed for scaling k_a in this study. Four air permeability models, 24 25 including Millington and Quirk (1960)-MQ, Hunt (2005)-HT, Brooks and Corey (1964)-BC, and Kawamoto et al. (2006)-KA, were selected to test this method using two independent datasets. 26 27 For the first dataset that included k_a measured for river sediments, all of the four models were able to delineate the spatial distribution of k_a with a reference curve of k_a and a set of scaling 28 29 factors. Specifically, the MQ model gave the least satisfactory results due to the less flexibility of its form; whereas, there were no significant differences in the performances for the HT, BC, and 30 31 KA models. For the second dataset that contained k_a measured for agricultural soils, the overall 32 performance of the four models for scaling k_a deteriorated, largely due to the alterations in the microscopic structures of soil samples caused by repacking and compression of soil samples. 33 Nonetheless, as the first attempt, this study shows the viability of using the similar media 34 concept for scaling k_{a} . The merit of this method resides in the fact that the spatial variations of 35 36 moisture conditions and soil properties can be simultaneously included for analyzing the spatial 37 distribution of k_a . With a reference curve of k_a and the distribution of scaling factors, this method would be particularly suitable for modeling subsurface gas transport. 38

39

Keywords: Similar Media Concept, Scaling Factor, Soil Air Permeability, Air Permeability
Model

42 **1. Introduction**

43 As one of the key factors that control subsurface gas transport and exchange of gas across the soil-atmosphere interface, soil air permeability (k_a) describes the ability of a soil to transmit gas. 44 The rising interest in k_a is manifested by its broad range of applications in various fields, 45 46 including greenhouse gas emission (Ball et al., 1997; Conen et al., 1998), landfills (Jain et al., 2005; Wu et al., 2012), soil vapor extraction systems (Poulsen et la., 1998; Farhan et al., 2001), 47 crop growth (Lipiec and Hatano, 2003; Barrios et al., 2005). Moreover, due to easy operation and 48 cost effectiveness, k_a measured near field capacity has been used to predict soil saturated 49 hydraulic conductivity (Loll et al., 1999; Iversen et al., 2001). 50

51 Under natural conditions, k_a is affected by a number of soil factors (e.g., air-filled porosity and pore size distribution) and moisture conditions, all of which show various degrees of spatial 52 variations. As such, field measured k_a exhibits significant spatial variability (Iversen et al., 2003). 53 For application and modeling purposes, it is thus crucial to evaluate the spatial distribution of k_a 54 and its controlling factors. However, compared to relevant researches on soil hydraulic properties, 55 only few studies are available on the spatial distribution of $k_{\rm a}$, which nonetheless provided 56 valuable insights into the understanding of the spatial pattern of k_a (Poulsen et al., 2001; Iversen 57 et al., 2003, 2004). Geostatistical techniques were mainly used in previous studies to analyze the 58 spatial pattern of k_a . Based on the results of variograms, Iversen et al. (2003) found that the range 59 of k_a for sandy soils was larger than the one for a loamy soil, probably due to the difference in 60 61 the depositional processes of those two types of soils. Although geostatistical techniques have been proven to be powerful tools for investigating naturally occurred phenomena, there are 62 certain shortcomings for those techniques as pointed out by Henley (2001). Most notably, the 63 underlying processes associated with studied targets cannot be explicitly considered in 64 65 geostatistical techniques, which rather rely on statistical models for examining the spatial 66 correlations of the targets.

With respect to k_a , the main issue of using geostatistical techniques stems from the impact of soil moisture on k_a . It has been well known that k_a is highly dependent on soil moisture levels, but the spatial distribution of soil moisture cannot be directly included in the geostatistical analyses of k_a . To some degree, this is analogous to study the spatial distribution of soil unsaturated hydraulic conductivity without specifying moisture conditions. Therefore, 72 precautions were usually taken in previous field studies on the spatial distribution of k_a . As 73 conjectured by Iversen et al. (2003), when soil moisture contents reach field capacity, the air 74 flow takes place in the majority of soil pores; thus, the impact of moisture on k_a can be neglected 75 with moisture contents near and below field capacity. Although the assumption made by Iversen et al. (2003) is useful for studying the spatial distribution of k_a under dry conditions, it may fail at 76 regions with wet climates or shallow groundwater tables. For application purposes (e.g., landfills 77 and soil vapor extraction systems), the inclusion of moisture in analyzing the spatial distribution 78 of k_a is also inevitable. Therefore, it is desirable to seek alternative methods to assess the spatial 79 distribution of k_a under the influence of soil moisture. 80

81 Along the line of above thinking, the scaling scheme based on the similar media concept may provide a promising approach to investigating the spatial distribution of k_a with the consideration 82 of moisture conditions. First introduced by Miller and Miller (1956), the similar media concept is 83 based on the assumption that the internal geometry of similar media only differs by microscopic 84 85 length scales that can be characterized by scaling factors. The purpose of this scaling approach is to coalesce a range of functional relationships into a single curve through scaling factors that 86 87 depict the spatial distribution of those functional relationships. More specifically, scaling factors are used to relate soil properties at a given location to the mean properties at a reference location, 88 89 which are invariant of moisture conditions. This scaling method has been widely used in soil hydrology to characterize spatial variability of soil hydraulic properties and soil hydraulic 90 functions with associated model parameters (Warrick et al., 1977; Hopmans, 1987; Shouse and 91 Mohanty, 1998; Zavattaro et al., 1999; Hendrayanto et al., 2000; Tuli et al., 2001). With a 92 93 reference functional relationship and the distribution of scaling factors, this method is particularly suitable for modeling purposes (Peck et al., 1977; Kabat et al., 1997; Salvucci, 1998; 94 Oliveira et al., 2006). Given the similarities between water flow and gas transport in soils, one 95 can expect the viability of applying this scaling approach for investigating the spatial distribution 96 of k_a under the influence of soil moisture. 97

To our knowledge, this research was the first attempt to extend the use of the similar media concept for scaling k_a . The main objective of this study was to examine the feasibility of this approach using two datasets collected from USA and France. Four air permeability models were selected to delineate the functional relationship between k_a and saturation degree of air. The results of this study demonstrated the feasibility of using the similar media concept for scaling k_a , which also opened the door for utilizing this method for simulating gas transport in soils.

104

105 **2. Materials and Methods**

106 2.1 Similar Media Concept in Soil Hydrology

107 The use of the similar media concept for scaling soil water retention and hydraulic conductivity curves is well documented in the literature (see the review by Vereecken et al., 108 2007). So, only a brief overview is given here. Based on the similar media concept, it is assumed 109 that the microscopic structures (e.g., tortuosity, and relative particle size and pore size 110 distributions) of similar soils are identical and only differ by microscopic length scales that can 111 be characterized by scaling factors (Peck et al., 1977; Warrick et al., 1977). The scaling factor 112 (α) is thus defined by the ratio of the microscopic characteristic length of a soil (λ) to the 113 characteristic length of a reference soil (λ_m): 114

115
$$\alpha_i = \frac{\lambda_i}{\lambda_m} \tag{1}$$

where i=1, 2, ..., L is the location of the soil and the subscript *m* denotes the reference soil. By its definition, α is invariant of soil moisture conditions and only dependent upon the location of the soil. As such, the soil water retention curve at any location can be scaled to the reference water retention curve through α :

120
$$h_i(\theta) = \frac{h_m(\theta)}{\alpha_{w,i}}$$
(2)

where *h* is the soil matric potential, θ is the volumetric moisture content, and the subscript *w* denotes water. The scaling relationship of the soil hydraulic conductivity curve can be written as:

123 $K_{w,i}(\theta) = K_{w,m}(\theta) \,\alpha_{w,i}^2 \tag{3}$

where K_w is the hydraulic conductivity. Due to the fact that soil porosity may vary across locations, instead of θ , the saturation degree of moisture (S_w) is usually used (Warrick et al., 126 1977):

$$S_w = \frac{\theta}{\phi} \tag{4}$$

128 where ϕ is the soil porosity.

To describe the reference curves of $h_{\rm m}$ and $K_{\rm w,m}$, soil water retention and hydraulic conductivity models (e.g., Brooks-Corey model and van Genuchten model) have been used, although polynomial functions of $h_{\rm m}(S_{\rm w})$ and $K_{\rm w,m}(S_{\rm w})$ were also adopted (Warrick et al., 1977). By optimizing the differences between measured and calculated $h(S_{\rm w})$ and $K_{\rm w}(S_{\rm w})$ across all the locations, the model parameters for the reference curves of $h_{\rm m}$ and $K_{\rm w,m}$, and the scaling factor $\alpha_{\rm a,i}$ at each location can be obtained (Hopmans, 1987; Tuli et al., 2001).

135

136 2.2 Extension of the Similar Media Concept for Scaling Soil Air Permeability

Although the scaling theory based on the similar media concept has been widely used in soil 137 138 hydrology, there is a surprising lack of studies on its application for assessing the spatial distribution of k_a . Strictly speaking, soil water permeability (k_w) should be used in the scaling 139 procedure (Eq. (3)), as regardless of fluid properties, soil permeability, whether it is k_w or k_a , 140 represents the intrinsic properties of soils to transmit fluids and thus reflects the microscopic 141 142 structures of soils. However, given that the fluid of interest remains under the same conditions (e.g., temperature), the fluid properties do not change across locations and conductivities (e.g., 143 $K_{\rm w}$) can be used in Eq. (3). Therefore, one can write a similar scaling relationship for $k_{\rm a}$ based on 144 Eq. (3): 145

146
$$k_{a,i}(S_a) = k_{a,m}(S_a) \alpha_{a,i}^2$$

147 where $S_a=(1-S_w)$ is the saturation degree of air and the subscript *a* denotes air. For the same 148 reason for using S_w in scaling $h(S_w)$ and $K_w(S_w)$, S_a was used in this study instead of the 149 volumetric soil-air content. Note that α_a is also invariant of soil moisture conditions and only 150 dependent upon the location of the soil.

To delineate the reference relationship of $k_{a,m}$, four air permeability models were selected (Ghanbarian-Alavijeh and Hunt, 2012). The first model was proposed by Millington and Quirk (1960) (denoted as MQ hereafter):

(5)

$$k_a(S_a) = k_o \times S_a^2 \tag{6}$$

155 where k_0 is the air permeability at the porosity ϕ .

The second model was developed by Hunt (2005) (HT), which is based on the continuum percolation theory:

158
$$k_a(S_a) = k_o \left(\frac{S_a - S_{a,t}}{1 - S_{a,t}}\right)^2$$
(7)

where $S_{a,t}$ is the percolation threshold used by Ghanbarian-Alavijeh and Hunt (2012). Note that the HT model can be reduced to Eq. (6) by assuming $S_{a,t}=0$.

161 The third model was based on the hydraulic conductivity function of Brooks and Corey 162 (1964) (BC):

163
$$k_a(S_a) = k_o S_a^l \left[1 + (1 - S_a)^{1 + \frac{2}{\gamma}} \right]$$
(8)

where γ is the pore size distribution index, and *l* is the tortuosity-pore connectivity factor and assumed to be 2 based on the percolation theory (Hunt and Ewing, 2009).

166 The last model was taken from Kawamoto et al. (2006) (KA):

167
$$k_a(S_a) = k_o \times (S_a)^{1+3/\eta}$$
 (9)

where $(1+3/\eta)$ represents the combined effects of tortuosity and connectivity of air-filled pores. In Kawamoto et al. (2006), the measured k_a at -100 cm H₂O was used instead of k_o . Essentially, the MQ model is a simplified version of Eq. (9) with $\eta=3$.

In parallel to previous studies on scaling $h(S_w)$ and $K_w(S_w)$, the scaling factors $\alpha_{a,i}$ and the model parameters (given in Tables 2 and 3) for describing the reference curves of $k_{a,m}$ were obtained by optimizing the differences between measured and calculated k_a . The following objective function was used:

175
$$O = \sum_{i=1}^{L} \left\{ \sum_{j=1}^{M(i)} \left[k_{a,i}^{obs}(S_{a,j}) - \alpha_{a,i}^2 \times k_{a,m}(S_{a,j}) \right]^2 \right\}$$
(10)

where M(i) is the total number of observed k_a at location i, $k_{a,i}^{obs}(S_{a,j})$ is the *j*th observed k_a corresponding to $S_{a,j}$ at location i, and $k_{a,m}(S_{a,j})$ is the calculated k_a at $S_{a,j}$ from Eq. (6) to Eq.(9). The generalized reduced gradient method was used to optimize Eq. (10) with the geometric mean of the scaling factors assumed to be unity (Hopmans, 1987; Zavattaro et al., 1999; Hendrayanto et al., 2000):

181
$$(\prod_{i=1}^{L} \alpha_i)^{1/L} = 1$$
(11)

To evaluate the performance of each air permeability model, the root mean square error(RMSE) was used:

184
$$RMSE = \left(\frac{1}{n}\sum_{k=1}^{n}(o_k - e_k)^2\right)^{\frac{1}{2}}$$
(12)

where *n* is the total number of observed k_a across all the locations, o_k is either observed $k_{a,i}(S_{a,j})$ or scaled $k_{a,i}(S_{a,j})$ (i.e., observed $k_{a,i}(S_{a,j}) \times \alpha_{a,i}^2$), and e_k is the corresponding $k_{a,m}(S_{a,j})$ calculated from the reference curve of $k_{a,m}$.

188

189 **2.3 Datasets**

Two independent datasets collected from USA and France were used in this study to test the 190 191 feasibility of applying the similar media concept for scaling k_a . The first dataset (denoted as Dataset I) included measured air permeability data using sediment samples taken from the 192 193 channel of the Platte River near Kearney, NE (Figure 1). The streambed sediments at the site are moderately-sorted, and mainly consist of sands and gravels with very low organic matter 194 195 contents (Table 1). A total of 25 samples with a length of 20 cm for each sample were collected 196 using transparent polycarbonate tubes with an inside diameter of 5.1 cm. The tubes were 197 vertically pressed to the depth of 20 cm into the sediments. The distance between neighboring 198 sampling locations was 1 m. After taking out the sample from the river channel, the bottom of the tube was wrapped by a figberglass screening to prevent sediments from falling off. A falling 199 head permeability test was carried out on site to determine the saturated hydraulic conductivity 200 201 of the sediments inside the tube (Chen, 2005). Water temperatures were also recorded during the test to calculate water dynamic viscosity, which was used to calculate k_{w} . 202

After the measurements of saturated hydraulic conductivities, the sediment samples were brought back to the laboratory and air dried at room temperatures. On Day 2 and every three 205 days afterwards until Day 17, air permeability tests were performed using a portable air permeameter (Figure 2), which resulted in 6 pairs of data points of k_a and S_a for each tested 206 207 sample. To measure k_a , the polycarbonate tubes were connected to the air permeameter, which mainly consisted of an air tank, an air pressure regulator, a rotameter that measured the flow rate 208 of air through the sediments, and a manometer that recorded the pressure difference across the 209 length of the sediments. Darcy's Law was then used to calculate k_a (Kirkham, 1947). Due to the 210 slippage of gas molecules near the pore wall, the flow of gas in porous media may exhibit non-211 Darcian flow that leads to higher k_a (Klinkenberg, 1941). Therefore, additional tests were 212 performed to examine the applicability of Darcy's Law by measuring the pressure difference 213 across the sample length under varying flow rates. The linearity of the results proved the 214 applicability of Darcy's Law and confirmed the air permeameter was operational for measuring 215 216 $k_{\rm a}$. The merit of the air permeameter used in this study is that it allows to measure both saturated hydraulic conductivity and k_a of the same sample without the need to remove the sediments from 217 the permeameter for repacking. Thus, it minimizes the possible disturbance of the sample 218 structure. After each air permeability test, samples were immediately weighed to determine 219 220 moisture contents and corresponding volumetric soil-air contents. On Day 17, all of the samples were dried in the oven for 24 h at the temperature of 105 °C to determine bulk density and 221 222 porosity. The results of the grain size analyses showed that the texture of the sediments was sand according to the USDA classification (Table 1). 223

224 The second dataset (denoted as Dataset II) was taken from Tang et al. (2011) and contained nine remolded soil samples from an experimental forest site at Le Breuil located in central 225 226 France. The reason for choosing Dataset II was that several air permeability models, including the HT model, were successfully tested on those soil samples (Tang et al., 2011; Ghanbarian-227 Alavijeh and Hunt, 2012). However, unlike the air permeameter used in Dataset I, a specially 228 designed oedometer was adopted by Tang et al. (2011) to measure $k_a(S_a)$ by vertically 229 compressing soil samples, which led to varying moisture contents under different compressional 230 stresses. Except for one sample with 6 data points, there were 8 pairs of k_a and S_a for each 231 sample. The texture of the soils was sandy loam and their physical properties are reported in 232 Table 1. 233

234

235 **3. Results and Discussions**

236 **3.1 Dataset I**

It has been well known that besides soil properties, k_a is also affected by moisture conditions 237 and thus S_a . With the drying process, the mean value of S_a for Dataset I increased from 0.49 on 238 Day 2 to 0.87 on Day 17. Consequently, the mean value of k_a was almost quadrupled from 21.0 239 μ m² on Day 2 to 83.6 μ m² on Day 17, as more smaller pores opened up for transmitting air when 240 the sediments became drier. The results from Dataset I again illustrate the importance of 241 including moisture conditions in analyzing the spatial distribution of k_a . To further demonstrate 242 the impact of S_a on k_a , the results from four sediment samples in Dataset I (numbered as #1, #2, 243 244 #3, and #4) are plotted in Figure 3. As expected, k_a increased with increasing S_a or with 245 decreasing moisture levels; however, the $k_a(S_a)$ relationship also differed among different samples. More specifically, k_a showed similar ranges for #1 (13.7 to 41.5 μ m²) and #2 (13.2 to 246 44.5 μ m²) with S_a ranging from 0.40 to 0.81 and 0.56 to 0.86, respectively (Figures 3a and 3b). 247 In comparison, k_a for #3 (11.3 to 142.3 μ m²) and #4 (4.5 to 111.5 μ m²) varied over an order of 248 magnitude with S_a increasing from 0.46 to 0.90 and 0.42 to 0.89, respectively (Figures 3c and 249 3d). Clearly, a single curve of $k_a(S_a)$ is not sufficient to describe spatially distributed k_a , even at 250 251 plot scales (e.g., a reach scale in this study). This presents a great challenge for simulating subsurface gas transport with spatially varied k_a and associated model parameters. The same 252 253 challenge is also faced by soil hydrologists to simulate water movement in vadose zone, where 254 soil hydraulic properties may vary substantially across landscapes. One of the solutions to tackle this problem in soil hydrology relies on the use of scaling factors within the framework of the 255 similar media concept (Kabat et al., 1997; Salvucci, 1998; Oliveira et al., 2006). 256

257 To have a thorough view of the effectiveness of using the similar media concept for scaling $k_{\rm a}$, the results of unscaled and scaled $k_{\rm a}$ from Dataset I are plotted in Figure 4 with the reference 258 curves of $k_{a,m}$ for all of the air permeability models. For the purpose of clarity, the results of 259 scaled k_a for each model are also plotted separately in Figure 4. The scaled k_a was calculated 260 using the obtained scaling factors (e.g., Eq. (5)). The obtained fitting parameters and RMSE 261 values for each model are given in Table 2. Note that the results of two sediment samples were 262 removed from the analyses, as they exhibited irregular patterns between k_a and S_a (e.g., k_a may 263 264 decrease with increasing S_a). Those irregular patterns were suspected to be the consequence of 265 the collapse of root holes in the samples during the drying process. As expected, a positive trend existed between unscaled k_a and S_a . When S_a was lower than 0.7, unscaled k_a largely remained 266 below 60 μ m², while unscaled k_a varied about 8 times from approximately 20 to 160 μ m² when 267 S_a reached above 0.8. Figure 5a further shows that a single curve of $k_a(S_a)$ is not sufficient to 268 delineate the distribution of the unscaled k_a as attested by the $k_{a,m}$ curves. By comparison, with 269 the aid of scaling factors, the scaled k_a coalesced along the $k_{a,m}$ curves with greatly reduced 270 271 scatter for all of the four air permeability models. Note that scaled k_a was systematically more scattered at the lower range of S_a for all of the models. It was due to the fact that more weights 272 were given to higher values of k_a in Eq. (10) during the optimization process. If k_a was replaced 273 by log k_a in Eq. (10), the scatter of scaled k_a at the lower range of S_a could be reduced; however, 274 the overall performance of the models deteriorated. Compared to the other three $k_{a,m}$ curves, the 275 one for the HT model slightly deviated; whereas, the resulting scaled k_a was comparable for all 276 of the four models (Figure 4b). Among all of the results, the RMSE value of the scaled k_a for the 277 MQ model was highest with $k_{a,m}$ overestimating scaled k_a at the lower end of S_a and 278 underestimating scaled k_a at the higher end of S_a (Figure 4c). This was because without 279 280 additional fitting parameters other than k_0 , the form of the MQ model was the least flexible. By contrast, the performances of the other three $k_{a,m}$ curves improved at both ends of S_a with the BC 281 282 model showing the most satisfactory results (Figures 4d-4f); however, there appeared no significant differences in the performances of those three air permeability models for scaling k_a 283 284 (Table 2), indicating the applicability of those models for delineating the distribution of k_a in the studied sediments. 285

The values of obtained k_0 varied among the four models (Table 2). The fitted k_0 in the KA 286 model was highest with the value of 106.0 μ m², while the one in the MQ model was only 92.5 287 μ m². As a result, the scaled k_a for the MQ model was lowest under the same level of S_a ; however, 288 289 the differences in the scaled k_a were minimum among different models. The value of the 290 obtained percolation threshold $S_{a,t}$ in the HT model was 0.168, which was comparable to the 0.1 threshold proposed by Ewing and Hunt (2006). In addition, the pore size distribution index γ (γ 291 292 =3.067 in this study) in the BC model fell within the range for sandy soils (Ghanbarian-Alavijeh and Hunt, 2012). In summary, Figure 4 shows that for the studied sediments, the spatially 293 distributed k_a could be described by a reference curve of $k_{a,m}$ and a set of scaling factors, which 294 295 attests the feasibility of using the similar media concept for scaling k_a under the influence of moisture. With a known reference curve of $k_{a,m}$ and the distribution of scaling factors, this method would be particularly suitable for modeling subsurface gas transport.

298 As previously explained, the spatially distributed k_a is mostly controlled by moisture 299 conditions and soil properties, both of which may vary at different spatiotemporal scales. 300 Essentially, the impacts of moisture levels and S_a on k_a can be determined through the reference curve of $k_{a,m}$; whereas, the spatial variations of soil properties are embedded in the distribution of 301 scaling factors, as scaling factors only depend on soil intrinsic properties (i.e., the microscopic 302 characteristic length). The probability plots of the scaling factors (both $\alpha_{a,i}$ and $\ln \alpha_{a,i}$) are given in 303 Figure 5 for the HT model. The resulting probability plots were similar for all of the four models, 304 305 so only the ones for the HT model are presented here. The respective mean and standard deviation was 1.03 and 0.23 for $\alpha_{a,i}$, and 0 and 0.23 for $\ln\alpha_{a,i}$. The values of the standard 306 307 deviations for $\alpha_{a,i}$ and $\ln \alpha_{a,i}$ were smaller than previously reported values derived from soil water retention and hydraulic conductivity data (Warrick et al., 1977; Hopmans, 1987; Hendrayanto et 308 309 al., 2000), probably due to the smaller number of sediment samples used in this study. Although those previous studies showed a log normal distribution of scaling factors, the Kolmogorov-310 311 Smirnov Test did not reject the null hypothesis that $\alpha_{a,i}$ and $\ln \alpha_{a,i}$ were normally distributed at the 312 significance level of 0.01. Thus, future studies are needed to include a larger number of samples, and to compare the distributions of scaling factors drawn separately from soil water retention and 313 hydraulic conductivity data and from air permeability data. 314

To test the applicability of the derived reference curves of $k_{a,m}$, the values of k_a on Day 17 315 were calculated using the obtained scaling factors and measured S_a based on Eq. (5). As 316 317 hypothesized by Iversen et al. (2003), when soil moisture contents reach field capacity, the air flow takes place in the majority of soil pores; thus, k_a measured at or near field capacity can be 318 used as a proxy of soil water permeability k_w to predict saturated hydraulic conductivity. On Day 319 320 17, the average value of soil moisture contents for all of the sediment samples was 0.04, which 321 was quite comparable to the observed field capacity of sandy soils (soil moisture data can be accessed from High Plains Regional Climate 322 the Center at <u>http://www.hprcc.unl.edu/awdn/soilm/</u>). Based on the above reasoning, the calculated k_a on Day 323 17 is plotted against k_w derived from measured saturated hydraulic conductivities in Figure 6. 324 Again, only the calculated k_a for the HT model is shown here as a demonstration. It can be seen 325

from Figure 6 that the calculated k_a was very close to k_w . On average, the calculated k_a only underestimated k_w by 11%. Figure 6 further corroborates the feasibility of using scaling factors within the similar media framework for describing spatially distributed k_a .

329

330 **3.2 Dataset II**

The previous section has demonstrated the feasibility of using the similar media concept for 331 scaling k_a in river sediments. To assess the viability of this method in agricultural soils, analyses 332 were carried out for Dataset II. A mean value of 0.02 for $S_{a,t}$ in the HT model was obtained by 333 Ghanbarian-Alavijeh and Hunt (2012) by fitting individual curves of k_a to soil samples in Dataset 334 II. Thus, to avoid the negative values of $(S_a - S_{a,t})$ in Eq. (7), $S_{a,t}$ was fixed to be 0.02 for Dataset 335 336 II. In addition, an unusually high value of γ in the BC model was obtained from the optimization, which was caused by fitting the $k_{a,m}$ curve to comparatively low values of k_a from one soil 337 338 sample (the data points with S_a between 0 and 0.2 shown in Figure 7a). Moreover, no physical 339 constraints for the fitting parameters were considered in the optimization process and $k_a(S_a)$ in 340 Eq. (8) became insensitive to γ for large γ values. Thus, instead of using this high value of γ , the class-averaged value of γ for sandy loam was used, which was derived from soil water retention 341 342 data (Rawls et al., 1982).

The unscaled and scaled k_a from Dataset II are presented in Figure 7, and the obtained fitting 343 parameters and RMSE values are given in Table 3. Overall, the unscaled k_a from Dataset II 344 exhibited a much wider distribution from 0.5 to 1980.8 μm^2 . Some of the remolded samples 345 (e.g., the one with the maximum k_a of 1980.8 μ m²) also showed different ranges of k_a (Figure 346 7a), which was probably caused by the repacking processes. As expected, the improvement of 347 the scaled k_a from the MQ model was least satisfactory (Figure 7c). By comparison, the 348 performances of the HT and BC models were considerably improved; however, the obtained $k_{a,m}$ 349 tended to overestimate most of the scaled k_a (Figures 7d and 7e), owing to the fact that the fitting 350 parameter (or shape factor) $S_{a,t}$ in the HT model and γ in the BC model were fixed during the 351 optimizations. Without the need to fix the fitting parameter η in the KA model, the $k_{a,m}$ curve 352 seemed to more reasonably delineate the distribution of the scaled k_a over the entire range of S_a 353 (Figures 7f), although it should be noted that a very high value of k_0 was obtained for the KA 354

355 model in order to match the data points from the sample with the maximum k_a of 1980.8 μ m². Using the same remolded samples from Dataset II, Ghanbarian-Alavijeh and Hunt (2012) 356 357 showed that the HT model was better than the other three models (e.g., MQ, BC, and KA) to describe the $k_a(S_a)$ relationship for individual samples; however, no significant improvement of 358 the HT model was found in this study for scaling k_a . One of the possible reasons for the different 359 360 conclusions about the model performances was due to the differences in the optimization procedures. In Ghanbarian-Alavijeh and Hunt (2012), the $k_a(S_a)$ curves were optimized for 361 individual samples and some of the fitting parameters (e.g., γ in the BC model) were obtained 362 through optimizing soil retention data rather than soil permeability data. 363

364 In comparison to the results from Dataset I, the overall performance of the four selected air permeability models for scaling k_a was less satisfactory for Dataset II, which can be attributed to 365 366 two reasons. First, for the soil samples in Dataset II, soil structures might have been altered during the repacking processes. Secondly, in order to achieve different moisture levels, soil 367 368 samples were compressed using an oedometer (Tang et al., 2011), which unavoidably changed 369 the microscopic structures of soils. The repacking processes and the compression of soil samples 370 might have led to variations in the microscopic structures among different soil samples and thus the less satisfactory performance of using the similar media concept for scaling k_a . Nonetheless, 371 Figure 7 shows that the use of scaling factors was able to narrow down the scatter of k_a around 372 the $k_{a,m}$ curves for the HT, BC, and KA models. In terms of the distribution of scaling factors 373 374 from Dataset II, the respective mean and standard deviation for the HT model was 1.01 and 0.14 for $\alpha_{a,i}$, and 0 and 0.13 for $\ln \alpha_{a,i}$. Again, the values of the standard deviations for $\alpha_{a,i}$ and $\ln \alpha_{a,i}$ 375 376 were smaller than the previously reported values derived from soil water retention and hydraulic 377 conductivity data (Warrick et al., 1977; Hopmans, 1987; Hendrayanto et al., 2000).

378

379 4. Conclusions

The use of the scaling factors based on the similar media concept was tested for scaling air permeability k_a using two independent datasets in this study. With a reference curve of $k_{a,m}$ and a set of scaling factors, this method was shown to be able to delineate the spatial distribution of k_a for the first dataset, which included k_a measured under different moisture conditions by drying river sediment samples. For the second dataset that contained k_a measured for agricultural soils, 385 however, the use of the similar media concept for scaling k_a was less successful. It was most 386 likely due to the alterations in the microscopic structures of soil samples caused by repacking 387 and compression of soil samples. The merit of this method resides in the fact that the spatial variations of moisture conditions and soil properties can be simultaneously included for 388 analyzing the spatial distribution of k_a . At any given location, the impact of S_a and thus moisture 389 levels on k_a can be quantified by the reference curve of $k_{a,m}$; whereas, soil properties are 390 embedded in the scaling factors. Since this study was the first attempt to apply the similar media 391 concept for scaling k_a , more studies are needed to test this method on soils with different textures 392 and to examine the impacting factors that control the distribution of scaling factors. It would be 393 394 also useful to compare scaling factors derived independently from air permeability data, and from soil water retention and hydraulic conductivity data, which may further elucidate the 395 physical meanings of those scaling factors. Future studies are also needed to incorporate this 396 method in modeling subsurface gas transport. 397

398

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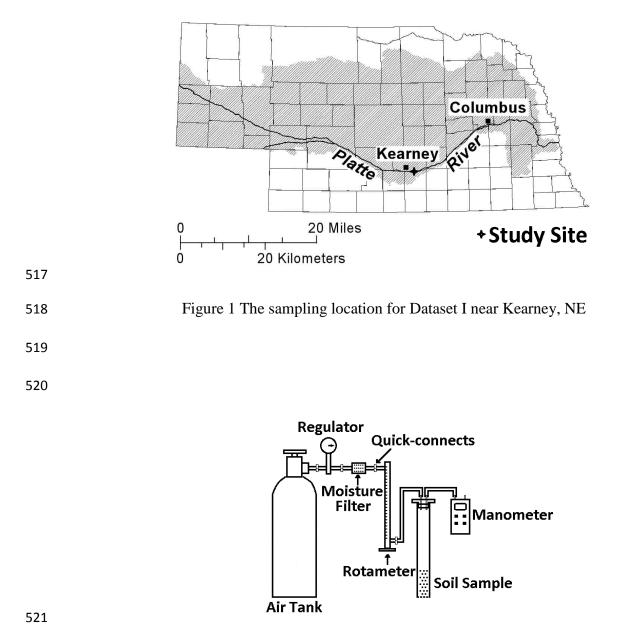
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522 Figure 2 Schematic of the air permeameter used for measuring air permeability in Dataset I

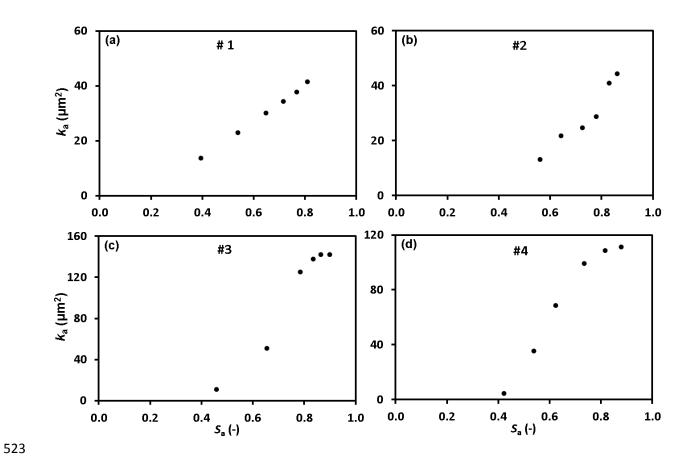
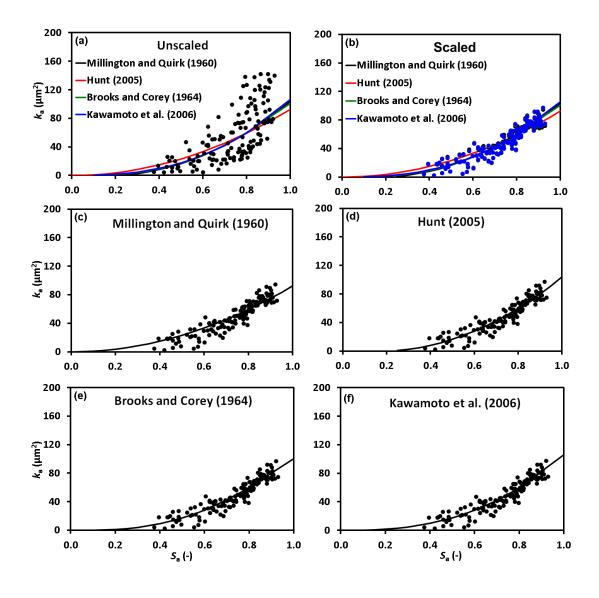


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Figure 4 Unscaled and scaled air permeability (k_a) with the reference curves of $k_{a,m}$ for the selected air permeability models from Dataset I: (a) unscaled k_a , (b) scaled k_a for all of the four air permeability models, (c) scaled k_a for the Millington and Quirk (1960) model, (d) scaled k_a for the Hunt (2005) model, (e) scaled k_a for the Brooks and Corey (1964) model, and (f) scaled k_a for the Kawamoto et al. (2006) model

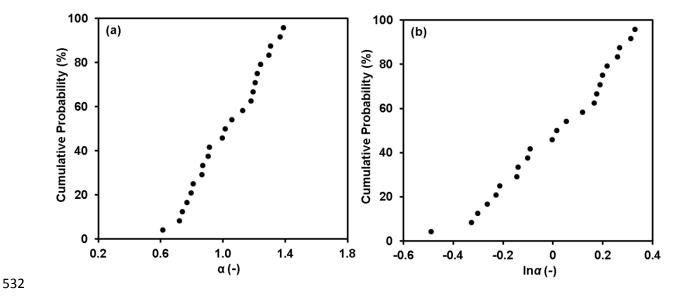
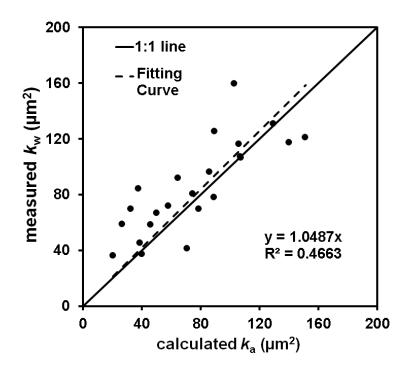


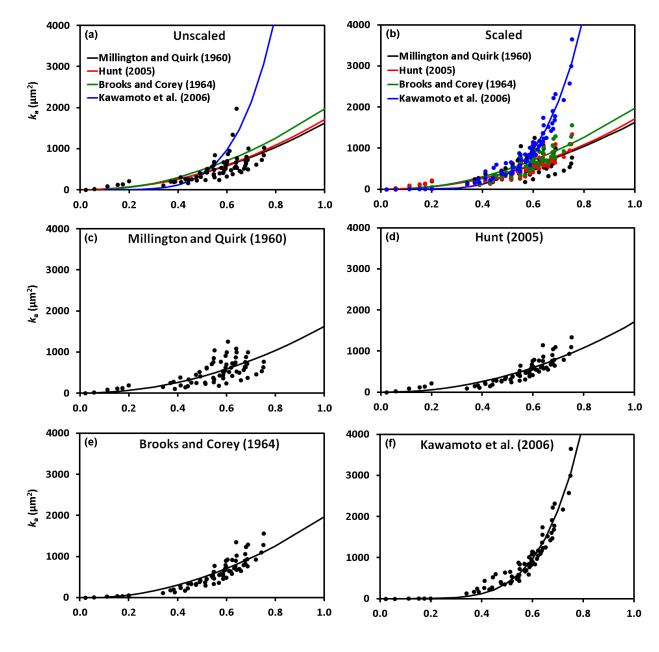
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Table 1 Summaries of the physical properties of sediment and soil samples from Dataset I and

554

	Dataset I	Dataset II
Texture [*]	Sand	Sandy Loam
Porosity	0.30	0.52
Gravel (>2mm)	10.9	-
Coarse Sand (2-0.5 mm)	53.0	-
Medium and Fine Sand (0.5-0.05 mm)	36.0	-
Silt and Clay (<0.05 mm)	0.1	-
Sand (2-0.05mm)	-	19.0
Silt (0.002-0.05mm)	_	23.0
Clay (<0.002mm)	_	58.0

Dataset II

555 * USDA classification

556

Table 2 Fitting parameters for the reference curves of $k_{a,m}$ with the corresponding RMSE values based on Dataset I

	Air Permeability Model	$k_{\rm o}(\mu{\rm m}^2)$	S _{a,t} (-)	γ(-)	η (-)	RMSE	
						Unscaled ka	Scaled ka
	Millington and Quirk (1960)	92.5	-	-	-	28.4	8.9
	Hunt (2005)	103.6	0.168	-	-	27.4	8.0
	Brooks and Corey (1964)	100.6	-	3.067	-	33.0	7.8
	Kawamoto et al. (2006)	106.0	-	-	1.893	27.4	7.9

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560

- Table 3 Fitting parameters for the reference curves of $k_{a,m}$ with the corresponding RMSE values
- 562

Air Permeability Model	$k_{\rm o}(\mu{\rm m}^2)$	S _{a,t} (-)	γ(-)	η (-)	RMSE	
All Fermeability Model					Unscaled ka	Scaled ka
Millington and Quirk (1960)	1625.3	-	-	-	228.7	217.3
Hunt (2005)	1709.4	0.02	-	-	229.5	135.0
Brooks and Corey (1964)	1966.2	-	0.378*	-	257.3	156.7
Kawamoto et al. (2006)	13390.6	-	-	0.724	719.7	188.7

based on Dataset II

^{*} Taken from Rawls et al. (1982)